

# **Toxicological assessment method for evaluating**

## **the occupational risk of workers involved in dynamic olfactometry**

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The paper aims to propose a new method to evaluate the occupational exposure risk for examiners involved in dynamic olfactometry. Indeed, during olfactometric analysis, human examiners are possibly exposed, at increasing concentration, to hazardous pollutants potentially present in odorous samples. A standardized method to evaluate the examiners' occupational safety is not yet available and the literature model existing present some critical aspect if applied to real odorous samples, mainly related to selection of the reference concentration to be applied and to the presence in the samples of compounds for which toxicity threshold is available. Due to these critical aspects, a deepening of assessment procedure to evaluate the occupation exposure risk for panellists involved in dynamic olfactometry is necessary. Therefore, this paper aims to propose a standardized approach for risk assessment in dynamic olfactometry. The proposed approach allows the quantification synthetic and conservative risk indices, which in turn can be used to calculate the minimum dilution factor useful for protecting the health of exposed subjects. In this model, the use of the hazard index (HI) for the mixture of compounds that make up the odour sample was proposed to assess the non-carcinogenic risk; the calculation of the inhalation risk (IR) was applied to estimate the carcinogenic risk. The methodology proposed in this paper uses different databases to retrieve proper occupational exposure limits, according to a hierarchical basis. These implementations allow obtaining the complete characterization of real mixtures and an accurate toxicological evaluation. From this elaboration, a more precise and conservative minimum dilution value can be evaluated and applied to protect the panellists' health.

**Keywords:** occupational health; odorous emissions, olfactometric analysis, olfactometric assessors, occupational risk assessment

## **1 Introduction**

Concerns related to odorous emissions from different industrial plants have increased in recent years, even related to their potential health impact [1]–[4]. Indeed, the statement that unpleasant odour may lead a health hazard persists, leading to complaints from exposed citizens [5]–[7]. For these reasons, the monitoring of odorous emissions emitted by industrial plants has become increasingly important and increasingly required to companies by local control agencies and several countries currently have specific regulations on this topic [8]–[12]. Among the various analytical techniques available, the most diffuse and the only currently regulated at the European level to quantify odour is dynamic olfactometry [13]. Dynamic olfactometry provides the sample odour concentration, expressed in European odour units per cubic meter ( $ou_E/m^3$ ). The odour concentration represents the number of dilutions with neutral air that are necessary to reach the odour threshold value. This technique is standardized by EN 13725:2003 and directly involves humans examiners, called panellists [14]. Dynamic olfactometry uses the nose as a sensor, due to the human nose's ability to perceive and distinguish the presence of an odour. During the analysis, odorous samples are diluted with neutral air by specific equipment, called olfactometer, and presented to human examiners at increasing concentration. Therefore, even at more diluted level, panellists are directly exposed to odorous and hazardous compounds potentially contained in the emission' sample. For

47 this reason, by conducting olfactometric analysis, an undefined occupational exposure risk exists for  
48 panellists involved. The exposure assessment for human health risk is therefore fundamental to  
49 define appropriate measures to protect the health of the panellists and thus to safely conduct  
50 olfactometric analyses. The specific occupational exposure of panellists and the regulations on  
51 sample handling impose strict limits on the possible protective measures to be taken to ensure the  
52 safety of the involved personnel: (i) Chemical transformations of odour samples or removal of  
53 hazardous pollutants are not allowed to avoid alteration of the sample; (ii) personal protective  
54 equipment (i.e., respiratory protective equipment) to protect the panellists against inhalation of  
55 hazardous substances cannot be adopted, for the same principles underlying the method as  
56 described above. In addition, exposure monitoring of panellist is not practicable in this exposure  
57 scenario. As the panellists are exposed to increasing concentrations of odorous emissions, a  
58 valuable protection measure that can be taken is the definition of a minimum dilution value, not to  
59 be overcome during analyses. The necessity to estimate the exposure risk for panellists involved in  
60 olfactometric analysis is a requirement already present in the EN 13725:2003 and additional remarks  
61 on this topic have been made in the current revision of the standard (prEN 13725:2019). Currently,  
62 the standard, solely prescribes to control and minimise the potential toxicological exposure risk and  
63 to inform exposed workers of the potential risk related to the olfactometric analysis. For these  
64 reasons, the standard revision should introduce guidance on risk assessment for panel members,  
65 prescribing the use of current relevant exposure limits. However, at the current state, a standardized  
66 method for estimating employment risk is not yet described. The absence of a standardized method  
67 to evaluate the exposure risk for panellist during olfactometric analysis can lead to several critical  
68 problems in the toxicological assessment. In particular, the absence of a reference concentrations  
69 may lead to non-comparable and even incorrect assessment of the panellists' exposure risk. Indeed,  
70 for workers exposure, different occupational exposure limits coexist: they can vary according to the  
71 exposure time and the exposure route considered. In addition, the occupational exposure limits can  
72 varied in accordance with the risk assessment approach applied [15]–[17]. In the literature, only two  
73 articles have proposed a methodology to assess occupational risk for workers involved in dynamic  
74 olfactometry. In these two studies, a methodology to estimate the minimum dilution value to be  
75 applied has been proposed. The first article [18], based on chemical data available in the literature,  
76 establishes the minimum dilution value to be used during the analysis of odour samples collected  
77 from different categories of industrial plants. The second study conducts a toxicological evaluation  
78 for odour emissions from an Italian municipal solid waste (MSW) incinerator, based on the authorised  
79 concentrations at stack emission for chemicals of potential concern (COPs) [19]. Both articles  
80 calculate the minimum dilution value to be applied by estimating the non-carcinogenic and  
81 carcinogenic risk for panellists' activity. However, in the absence of specific and detailed method in  
82 the standards about dynamic olfactometry, the assessment can be conducted using different  
83 reference databases. In addition, the two papers do not describe how to consider, in the toxicological  
84 evaluation of real samples, the compounds observed for which there is no exposure limit in the  
85 proposed databases. This problem can be crucial in presence of a high number of undefined  
86 compounds in an odorous sample. For all these reasons, there is a clear need to discuss an  
87 assessment method that considers these critical issues and that can be easily implemented and  
88 used to assess minimum dilution levels for real samples, in which compounds may be present for  
89 which an exposure threshold is not readily available. Therefore, this article aims to present a robust  
90 method for assessing occupational risk for dynamic olfactometry panellists. The proposed method  
91 has been constructed to be easily usable by olfactometric operators and considering the criticalities  
92 observed in the literature in this field.

## 93 **2 Method development**

### 94 **2.1 Occupational exposure of panel involved in dynamic olfactometry**

95 To conduct a toxicological assessment for olfactometric examiners and determinate the minimum  
96 dilution values useful to guarantee their safety, preliminary observations about their exposure should  
97 be conducted. Firstly, it is essential to assess the type of exposure. Panellists have to be considered  
98 as workers and therefore occupational exposure limits have to be used within the toxicological

99 assessment. However, the work exposure of these employees involved in olfactometric analysis is  
100 unusual. Indeed, panellists usually work in daily sessions of 1 or 2 hours and the examiner is called  
101 for a maximum of one analysis session per day in order to avoid nose fatigue. In addition, each odour  
102 presentation lasts for a maximum of 15 seconds, according to standard requirements [14]. Therefore,  
103 an olfactometric examiner will never be exposed for the same length of time as a common worker,  
104 who is generally exposed for a conventional 8-hour workday and a 40-hour workweek, for the entire  
105 working lifetime [20][21].

106 Nevertheless, it is necessary to investigate the salubrity of the panel because of the specific  
107 concentration to which they may be exposed during analysis. According to the sampling  
108 requirements of EN 13725:2003, odour emissions to which examiners are exposed during the  
109 analysis have to be collected directly at the odour source [14]. Therefore, the odour emissions  
110 analysed are not diluted by atmospheric dispersion, but only by the olfactometer during analysis  
111 session. As described previously, odour samples are diluted by this instrument with neutral air and  
112 presented to panellists at increasing concentration, until the examiners can distinguish the presence  
113 of an odour different from neutral air. Consequently, examiners may potentially be exposed to not  
114 negligible concentrations of hazardous pollutants. This exposure concentration is potentially higher  
115 than the exposure concentration of employees working directly inside the plant, precisely because  
116 the emission is not diluted in ambient air. For all these reasons, it remains necessary to assess the  
117 toxicological risk for examiners involved in dynamic olfactometry, using a precautionary approach  
118 but considering all the aspects associated with such a peculiar exposure scenario.

## 119 **2.2 Method description**

120 For what has been described above, in the considered scenario the source of exposure for panellists  
121 is the sample itself (which is diluted with neutral air) and the exposure pathway is limited to inhalation  
122 of chemicals in air through the olfactometer. This should be considered as a combined exposure  
123 scenario (i.e., a scenario in which exposure to multiple chemicals occurs by a single route, from one  
124 source of release). The odour samples could also be generally classified as a discharge mixture (i.e.,  
125 a substance combination that are emitted by a single industrial site) which is typically complex and  
126 variable in composition [22]. Then, a method was defined with the aim of characterising the health  
127 risk posed by the odour samples with a general and simplified, but robust approach. As a function of  
128 the exposure modes, a non-carcinogenic risk assessment was calculated, based on short-term  
129 exposure.

130 Further, for carcinogenic health effects, an excess lifetime cancer risk was calculated considering a  
131 “non-threshold” toxicity and assuming a linear dose-response relationship, as described in 2.2.2. For  
132 this study, it was assumed that no chemical transformations or abatements are induced during the  
133 dilution steps.

### 134 **2.2.1 Risk Assessment (Non-carcinogenic health effects)**

135 The potential non-carcinogenic chronic risk was evaluated using the hazard index (HI) (Equation 1).  
136 The HI is equal to the sum of each chemical component’s Hazard Quotient ( $HQ = \text{Exposure} \div \text{Safe}$   
137  $\text{Dose} = \text{Chemical concentration in odour sample} \div \text{Exposure reference value}$ ), as reported in  
138 Equation 2 [22].

$$HI = \sum_i^N HQ_i \quad \text{Equation 1}$$

139

$$HQ = \frac{\text{Exposure}}{\text{Occupational Limit Value}} \quad \text{Equation 2}$$

140

141 HI is one of the more straightforward ways to assess combined exposures to multiple chemicals and  
142 this approach is widely adopted for occupational chemical risk assessment. For example, ECHA  
143 suggests the use of risk characterization ratios (RCRs), calculated as the ratio between the exposure  
144 to a chemical and the corresponding derived no-effect level (DNEL) for that chemical [23]. A recent

145 EU standard [21] also suggests a similar approach (Appendix C.2 - Exposure Index) as “Tier 1”  
146 method for the risk assessment of simultaneous workplace exposure to several chemical agents.  
147 Another example of application of a similar method is also provided by ACGIH (American Conference  
148 of Governmental Industrial Hygienists), which applies a similar method for applying its own Threshold  
149 Limit Values for Mixtures [20]. Overall, HI represents a simple and flexible approach that can be  
150 quickly applied to different samples in the same scenario. The HI approach provides a transparent  
151 index of acceptable risk (i.e.,  $HI \leq 1$ : risk acceptable;  $HI > 1$ : potential concern). Further, it does not  
152 require an in-depth analysis of the toxicokinetic and toxicodynamic characteristics (or an evaluation  
153 of the specific toxicological endpoints) of each chemical in the mixture. For this reason, applying HI,  
154 it is possible to consider at the same time chemicals belonging to different groups and easily  
155 quantifying the percent contribution of each substance, which in turn permit the identification of target  
156 chemicals for risk management [22].

157 This approach also has some drawbacks: HI could be potentially over-conservative, which at the  
158 same time can be seen as an advantage, especially in the case in which the mixture under  
159 investigation is complex and poorly characterized from a toxicological point of view. Further, HI is a  
160 rather simplified approach, that does not consider chemical interactions and toxicokinetic or  
161 toxicodynamic differences. More complex approaches are available, but they require different  
162 assumptions and deeper toxicological characterizations and are difficult to apply for the purpose of  
163 this study [22], [24]–[28]. Finally, to apply the HI method, reference values for all chemicals need to  
164 be available these values are often not available.

165 This last issue could be one of the most critical aspects regarding the applicability of the HI method  
166 because very often reference values for occupational human exposure are not available to derive  
167 the single HQ. On the contrary, for some chemicals, several different reference concentration values  
168 are available. More in general, different occupational exposure limit values for chemicals exist at  
169 European and International level and were developed within different legislative frameworks. This is  
170 particularly critical for the case study under consideration, the panel's occupational exposure risk  
171 assessment, for which at the regulatory level (EN 13725:2003 and prEN 13725:2019) no  
172 toxicological parameter is specified to be adopted in the evaluation. Concerning this selection, the  
173 type of reference values used within a single HI calculation should be consistent, because each  
174 value is derived using slightly different assumptions, or since different critical effects may have driven  
175 the derivation of the reference values for the components. For example, OELs are developed for  
176 occupational safety and health purposes while DNELs were initially intended to play a role within risk  
177 management processes and prevention scheme in a regulatory framework. This could result in  
178 difficulty for the selection of the most appropriate limit value by experts and for the application of  
179 such limit values in the risk characterization in real exposure scenarios [12].

180 To overcome these problems and to allow the application of the HI method in this context, in this  
181 study a method has been proposed for the searching and for selecting the most appropriate  
182 reference value for the different chemicals that can be identified in the mixture of chemical agents  
183 present in odour samples. Furthermore, considering that chemical-specific reference values may not  
184 be found for all components of the mixture, two different approaches for using appropriate generic  
185 reference values have been proposed. Thus, for the purpose of the risk assessment of odorous  
186 samples, the assessor must apply a precise and careful evaluation procedure, divided into the  
187 following points [12]:

- 188 • Search among the proper occupational exposure limit values (i.e., short-term, inhalation  
189 exposure route) available for each component.
- 190 • Ensure that the selected value is equivalent or more conservative than the binding value, if  
191 this is available.
- 192 • Evaluate the derivation procedure of the limit value and the starting points in order to select  
193 the most robust limit values for the specific purpose of the assessment, always in a path that  
194 is documented as much as possible to ensure consistency, transparency and professional  
195 correctness.

196  
197 The role and competence of the assessor remains crucial for a proper choice. However, considering  
198 that the purpose of this work is to suggest a method to be implemented and used by any olfactometric  
199 laboratory to assess panel safety, it was decided to construct and propose a toxicological  
200 assessment approach widely usable and more flexible. In addition, the decision-making process

201 described in this paper has been structured to solve the main criticalities observed in the state of the  
202 art, in particular, the unavailability of a specific reference threshold to be adopted in the evaluation  
203 and the presence of compounds for which no toxicity threshold is provided.

204 To conduct the toxicological assessment, it is essential to investigate the chemical nature of real  
205 samples to be analysed by dynamic olfactometry. The most applied analytical technique to obtain  
206 qualitative and quantitative information about compounds present in gaseous environmental  
207 matrices, such as odour samples, is the Gas Chromatography coupled with Mass Spectrometry (GC-  
208 MS). Indeed, GC-MS is the widely used analytical technique to achieve the resolution of complex  
209 mixtures and the identification and quantification of unknown molecules [13], [29], [30].

210 Thus, to evaluate the HI value for a specific odour sample, once its chemical composition has been  
211 defined, in terms of nature and concentration of pollutants present, the following scheme has been  
212 constructed with basic elements for selecting the most relevant limit values to be used within the risk  
213 assessment process. For each component of the mixture identified by GC-MS analysis, the assessor  
214 should:

- 215  
216 1. Check among the occupational limit values (OELs) for short-term exposure (15-minute) that  
217 have the force of law at the international level. In this case the following OELs have been  
218 identified as those of interest:
  - 219 - *Indicative occupational exposure limit values (IOELV)*: health-based, non-binding values,  
220 derived from the most recent scientific data available and taking into account the  
221 availability of reliable measurement techniques [12]. The last update of list of the available  
222 IOELV was performed with the publication of Directive 2019/1831 (former European  
223 Directives reporting IOELVs are Directives 91/322/CEE, 2000/39/EC, 2006/15/Ec,  
224 2009/161/EU, 2017/164/EU);
  - 225 - *Binding occupational exposure limit values (BOELV)*: these are to be understood as  
226 binding threshold limit values, specially developed for the protection of the health of  
227 workers exposed to substances with genotoxic, carcinogenic or respiratory sensitization  
228 effects for which it is not possible to define a No Observed Adverse Effect Level (NOAEL)  
229 or a Lowest Observed Adverse Effect Level (LOAEL) [12]. BOELV are reported in  
230 European Directives 98/24/CE, 2003/18/CE, 2004/37/CE, (UE) 2017/2398, (UE)  
231 2019/130, (UE) 2019/983. The methodology to be adopted to define a IOELVs and  
232 BOELVs is reported in a specific guideline [31].
- 233  
234 2. Check among Derived No Effect Levels (DNELs) for workers' short-term exposure and to  
235 prevent systematic effects. DNELs are levels of exposure above which humans should not  
236 be exposed. DNELs must be declared by the registrants in the framework of REACH  
237 Regulation (Regulation (EC) 1907/2006)). The methodology to be adopted to develop a  
238 DNEL is reported in a specific guideline [32]. Several DNEL types may be established for the  
239 same chemical. For the purpose of this study, Acute – inhalation, systemic effects DNELs  
240 for workers should be considered. DNELs for registered substances are publicly available on  
241 ECHA website [33].
- 242  
243 3. Check for other short-term occupational limit values without mandatory value, defined at  
244 national level, choosing the most conservative value among those available at international  
245 level.

246  
247 The updated international OELs are available in GESTIS database - International limit values for  
248 chemical agents [34]. This database contains a collection of occupational limit values from 32 lists  
249 from 27 countries: various European states, Australia, Canada (Ontario and Québec), Israel, Japan,  
250 New Zealand, Singapore, South Korea, The People's Republic of China, Turkey, and the United  
251 States. IOELVs and BOELVs are also collected in this database. Limit values of more than 2,000  
252 substances are listed. ACGIH Threshold limit Values are not included in this database, but given  
253 their use is widely diffused, this source has also been consulted [20]. It should be noted that limit  
254 values defined by the various expert bodies and authorities differ in the criteria for their derivation,  
255 the level of protection which they offer, and their legal relevance.

256 After consulting the various sources mentioned above, in search of the respective exposure limit  
257 values, the assessor will be able to define, based on those available, which limit value to use.  
258 The proposed hierarchical order to follow is: (1) short-term OEL with legal value at international level;  
259 (2) DNEL (for workers' acute - inhalation, to prevent systemic effects); (3) Short term OEL without  
260 force of law, defined at national level. The assessor should ensure that the selected threshold value  
261 should be equivalent or more conservative than the binding value, anyway.  
262 The application of this procedure makes it possible to provide a toxicity value useful for calculating  
263 the HQ of a wide range of compounds present in a real odour sample.  
264 However, even consulting different sources according to the above scheme, it is possible that for  
265 some of the components of a mixture of an odour sample it may not be possible to establish a specific  
266 threshold value and that a non-negligible percentage of the mixture may not be characterized in  
267 terms of risk. In order to resolve this critical issue and complete the HI calculation, an approach is  
268 proposed.  
269 Indeed, for each compound for which a toxicity threshold is not available in the above databases, the  
270 assessor could use an occupational limit value defined for a family or group of chemical agents, if  
271 available (e.g., Hydrocarbons, aliphatic, C6-C8; Hydrocarbons, aliphatic, C9-C14; Hydrocarbons,  
272 aromatic, C9-C14; etc.) instead of the specified limit value for a single chemical. Discussing  
273 hydrocarbon molecules, another methods for defining a limit value for a family or group of chemical  
274 agents can be considered, such as the reciprocal calculation procedure (RCP) for deriving exposure  
275 limits, as described in [20].  
276 In any case, with the application of this approximation, it is considered possible to be able to  
277 characterize the HI of the entire mixture of chemicals that make up the odour sample, thus helping  
278 to complete its risk characterization. An acceptable non-carcinogenic risk level is obtained for HI  
279 lower than 1. It is however possible that even applying these two approaches, for a minor component  
280 of the odour sample it could not be possible to identify an exposure limit value and therefore that this  
281 component may not be characterized in terms of risk (i.e., HQ could not be calculated for some  
282 components). This aspect possibly introduces an element of uncertainty in the risk assessment,  
283 which must be considered.

## 284 2.2.2 Risk Assessment (Carcinogenic risk)

285 A method was also defined and applied to characterise the health risk posed by the odour sample.  
286 Carcinogenic excess lifetime risk was calculated based on Equation 3, as reported in existing  
287 guideline [35].  
288

$$\text{Inhalation Risk} = \text{CDI} \times \text{IUR} \quad \text{Equation 3}$$

289

290 where IUR is the Inhalation Unit Risk and CDI was the chronic daily intake, expressed in  $\mu\text{g}/\text{m}^3$ . IUR  
291 values were retrieved by the Risk Assessment Information System [36], while CDI has to be  
292 calculated in accordance with the working exposure time of the examiners involved in the  
293 olfactometric analysis. Indeed, CDI can be calculated as follow (Equation 4):  
294

$$\text{CDI} = \frac{C_{\text{air}} \times EF_{\text{iw}} \times ED_{\text{iw}} \times ET_{\text{iw}}}{AT_{\text{iw}} \times LT} \quad \text{Equation 4}$$

295

296 where  $C_{\text{air}}$  is concentration observed of pollutant [ $\mu\text{g}/\text{m}^3$ ],  $EF_{\text{iw}}$  is the exposure frequency [day/year],  
297  $ED_{\text{iw}}$  is the exposure duration [year],  $ET$  is exposure time [hours/day],  $AT_{\text{ir}}$  is averaging time  
298 [days/years]  $LT_{\text{iw}}$  is lifetime [years].

299 Therefore, to calculate CDI, it will be necessary to define these parameters for each olfactometric  
300 laboratory, with reference to the exposure of the panels involved in the analyses. Indeed, exposure  
301 parameters are closely correlated to the duration and frequency of panellists' exposure to pollutants  
302 in odour samples. Therefore, to assess a realistic specific risk, it will be necessary for each

303 olfactometric laboratory responsible to record in detail the representative work activity (i.e., the  
 304 “worst-case scenario”) of the examiners involved.

305 Observing the scientific literature about panellists’ exposure, the parameters used to calculate the  
 306 CDI are reported in In the model proposed in this paper, the inhalation risk was calculated for each  
 307 component classified as carcinogenic or potential carcinogenic, according to the definition supplied  
 308 in Part 3 of Annex VI to Regulation (EC) No 1272/2008:

- 309 • Carcinogen if the compound is classified as carcinogens of category 1A or 1B.
- 310 • Suspected to have cancerogenic potential for human if the compound is classified as  
 311 carcinogens of category 2.

312 Inhalation Risk factor for each carcinogen and suspected carcinogen pollutants was calculated and  
 313 were summed to obtain the total Inhalation Risk of the sample. An acceptable carcinogenic risk level  
 314 is defined for an Inhalation Risk lower than  $< 10^{-6}$  for a single compound and lower than  $10^{-5}$  in the  
 315 case of mixtures.  
 316

317 Table 1 as an example. In a previous study [19], two hypothetical exposure scenario are described,  
 318 in which panellists were considered to be exposed in the performance of their tasks, but with a  
 319 different workload. The references for the parametrization of these two scenarios were derived from  
 320 two different olfactometric laboratories: a "commercial" laboratory (owned by a private company) and  
 321 an "institutional" laboratory (owned by an environmental inspection authority). As shown in In the  
 322 model proposed in this paper, the inhalation risk was calculated for each component classified as  
 323 carcinogenic or potential carcinogenic, according to the definition supplied in Part 3 of Annex VI to  
 324 Regulation (EC) No 1272/2008:

- 325 • Carcinogen if the compound is classified as carcinogens of category 1A or 1B.
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 327 carcinogens of category 2.

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 329 were summed to obtain the total Inhalation Risk of the sample. An acceptable carcinogenic risk level  
 330 is defined for an Inhalation Risk lower than  $< 10^{-6}$  for a single compound and lower than  $10^{-5}$  in the  
 331 case of mixtures.  
 332

333 Table 1, the exposure parameters vary considerably depending on the exposure scenario, modifying  
 334 the resulting CDI.

335 In the model proposed in this paper, the inhalation risk was calculated for each component classified  
 336 as carcinogenic or potential carcinogenic, according to the definition supplied in Part 3 of Annex VI  
 337 to Regulation (EC) No 1272/2008:

- 338 • Carcinogen if the compound is classified as carcinogens of category 1A or 1B.
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 340 carcinogens of category 2.

341 Inhalation Risk factor for each carcinogen and suspected carcinogen pollutants was calculated and  
 342 were summed to obtain the total Inhalation Risk of the sample. An acceptable carcinogenic risk level  
 343 is defined for an Inhalation Risk lower than  $< 10^{-6}$  for a single compound and lower than  $10^{-5}$  in the  
 344 case of mixtures.  
 345

346 Table 1. Parameter for inhalation Risk calculation (derived from [19])

| Parameter [unit]  | “Commercial Lab” Scenario  | “Institutional Lab” scenario |
|---|--|------------------------------|
| Inhalation Unit Risk -IUR<br>[ $\mu\text{g}/\text{m}^3$ ] <sup>-1</sup>   | Retrieved by the Risk Assessment Information System ( <a href="https://rais.ornl.gov/">https://rais.ornl.gov/</a> )<br>for carcinogen chemical in the odour sample (Carcinogen cat. 1A, 1b, 2) |                              |
| Exposure concentration -<br>C <sub>air</sub> [ $\mu\text{g}/\text{m}^3$ ] | Chemical's concentration in the odour sample<br>(Carcinogen cat. 1A, 1B, 2)  |                              |
| Exposure Frequency -<br>EF <sub>iw</sub> [day/year]                       | 90   | 10                           |
| Exposure Duration - ED <sub>iw</sub><br>[year]                            | 10   | 7                            |

|  |       |      |
|--|-------|------|
| Exposure Time - ET <sub>iw</sub><br>[hours/day]      | 0,073 | 0,17 |
| Averaging time - AT <sub>iw</sub><br>[days/years]    | 365   | 365  |
| Life-time of the exposed<br>individuals - LT [years] | 70    | 70   |

347

### 348 2.2.3 Minimum dilution of the odour sample

349 As previously described, during olfactometric analysis odorous samples are presented to examiners  
350 at increasing concentration. Due to this, for every odour sample analysed by dynamic olfactometry,  
351 a minimum dilution level can be defined to avoid any health impact for the involved panellists. As  
352 stated above, no preventive or protective measures other than sample dilution can be applied, due  
353 to the requirements and specifications of the analysis. The minimum dilution level is defined as the  
354 dilution step not to be exceeded in order not to expose examiners to dangerous concentrations of  
355 pollutants within an odour sample. The determination of the minimum dilution value is defined on the  
356 results of the characterization of the non-carcinogenic risk (HI calculation) and of the carcinogenic  
357 risk (Inhalation Risk calculation). Indeed, by comparing the results obtained using the method  
358 described above with the acceptable safety criteria for carcinogenic and non-carcinogenic risk, it is  
359 possible to define the minimum dilution level for an odour sample. In Table 2 the acceptability criteria  
360 for non-carcinogenic and carcinogenic risk were reported: if HI or IR are higher than these  
361 parameters, a minimum dilution value must be set to protect panellist's health.

362

363 Table 2. Risk acceptability criteria adopted in this method

| Risk type             | Acceptance parameters                      |
|-----------------------|--|
| Non-carcinogenic (NC) | HI < 1                                     |
| Carcinogenic (C)      | IR < 10 <sup>-6</sup> for single substance |
|                       | IR < 10 <sup>-5</sup> for gaseous mixture  |

364

365 First of all, it is necessary to evaluate if an odour sample, according to the elaboration performed,  
366 exceeds the acceptability criteria (Table 2) for either or both criteria. As described above, in dynamic  
367 olfactometry, the odour concentration of an odorous gas sample is determined by presenting the  
368 sample to a panel of selected and screened human examiners, varying the concentration by diluting  
369 with neutral air. The dilution factor needed to reach the panel detection threshold determines the  
370 quantity value for the odour concentration of the undiluted odour sample. Concerning the minimum  
371 dilution, as mentioned, the goal is to protect the panellists' health. For this purpose, if both HI and IR  
372 respect the acceptability values shown in Table 2, it is not necessary to define a minimum level of  
373 dilution as regards the potential impact on health. Contrariwise, if calculated HI and/or IR values  
374 exceed the acceptability parameters for the sample as-it-is (i.e., for the undiluted sample), it is  
375 necessary to determine the minimum dilution value, which allows the above thresholds to be  
376 respected. For this purpose, knowing a priori the different dilution steps applied by the olfactometer,  
377 it is easy to evaluate the minimum level of dilution to be applied to avoid incurring an unacceptable  
378 carcinogenic or non-carcinogenic risk. It should be noted that the dynamic olfactometry method still  
379 uses a minimum dilution for technical and procedural reasons, in particular this dilution is typically  
380 fixed at a 4. This implies that the definition of a minimum dilution value can be considered  
381 conservative, even with respect to any elements of uncertainty that can be introduced with some  
382 simplifications and assumptions that become necessary. For the conduction of olfactometric  
383 analysis, it is necessary to calculate HI and IR on the as-it-is sample, and then apply this parameter  
384 to define, on the dilution scale applicable by the olfactometer, the minimum dilution step. The entire  
385 decision-making process is reported in **Errore. L'origine riferimento non è stata trovata.**

## 2.2.4 Limitations and Strengths of the method

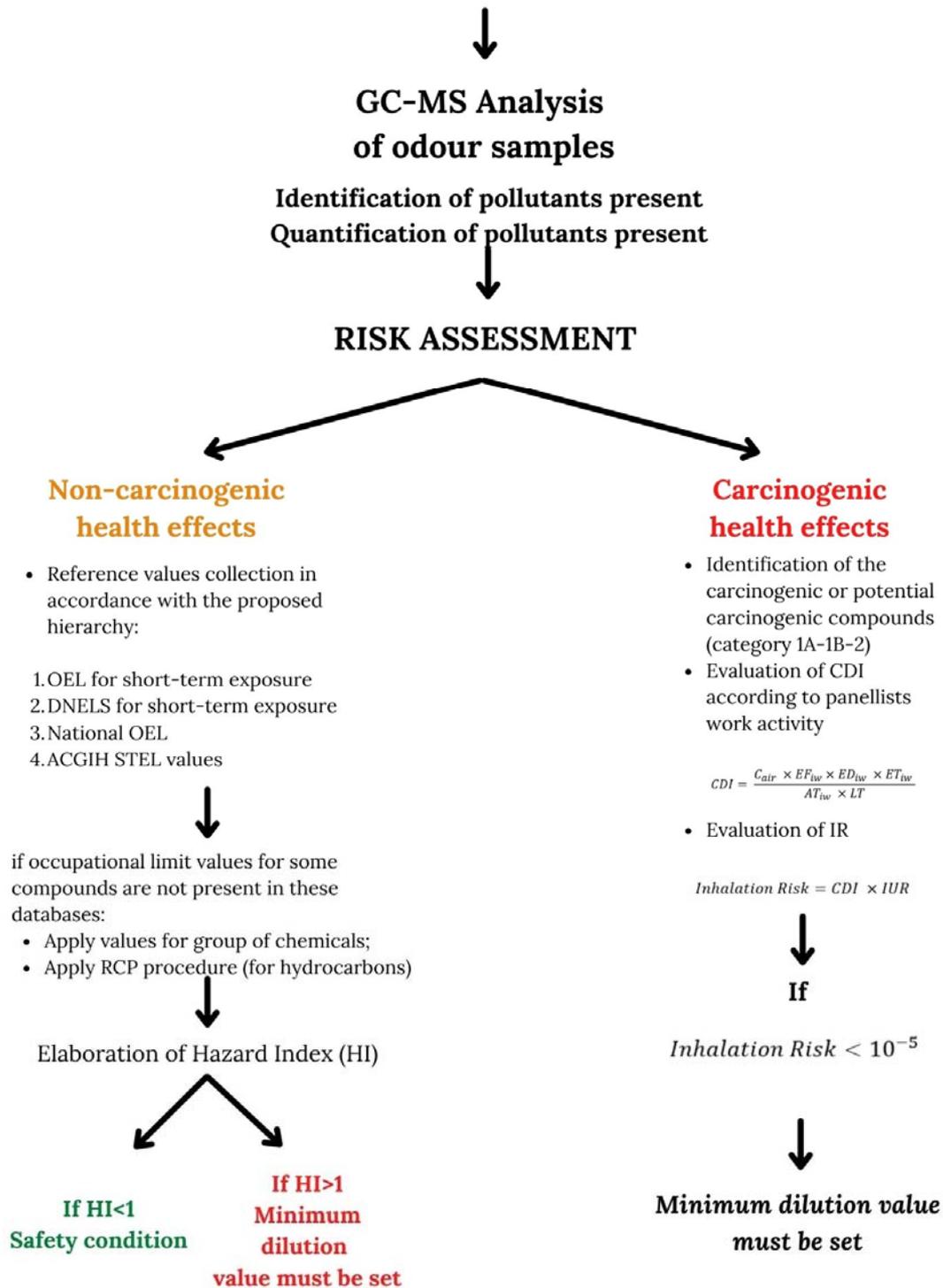
387 Some assumptions and limitations must be considered in the interpretation of the results of this  
388 procedure as the study design limits the general applicability of the findings. Major limitations of the  
389 present method (which should be considered in the interpretation of the results and the general  
390 applicability of the findings) are summarized hereafter. (i) As mentioned in paragraph 2.2.1, HI is a  
391 simplified approach, that does not consider chemical interactions and toxicokinetic or toxicodynamic  
392 differences. Our aim was to define a simple and flexible approach, anyway. This needs to be rather  
393 simple - but conservative enough - to be quickly applied to different samples in the same scenario,  
394 to allow the performance of dynamic olfactometry tests while protecting the health of the panellists.  
395 The application of the HI method should be seen as a "first-tier method" for risk assessment in a wider  
396 tiered approach. This Tier-1 method, to be effective, must be simple (i.e., data poor) and  
397 conservative (protective). However, it is an implicit limit of the method which can lead to a certain  
398 degree of uncertainty. Therefore, it is advisable to suggest the application of "higher tier" risk  
399 assessment methods if further, specific in-depth toxicological assessments prove to be necessary,  
400 or if decidedly non-negligible risks are highlighted by means of the tier 1 assessment phase. As  
401 reported above, more complex approaches are available, requiring different assumptions and  
402 deeper toxicological characterizations [22], [24]–[28]. In any case, it is good to remember that skills  
403 and expertise in the field of toxicology and health impact assessment must not be taken for granted,  
404 therefore in case of doubt in the application of the above method, it is good to refer to experts in the  
405 field. (ii) Further, difficulties may arise for the selection of the most appropriate occupational limit  
406 value for HQ and calculation. The proposed method respects some general principles of caution and  
407 invites to refer to reliable sources for the choice of the limit values. It should be remembered that  
408 various exposure limit values are available, which are derived according to different principles and  
409 methods. In this regard, differences could be observed by comparing exposure with different  
410 exposure levels and this could result, for the same sample, in defining different probabilities of risk  
411 [15], [17], [37], [38]. The proposed hierarchical criteria should make it possible to take a choice in  
412 compliance with the binding exposure limit values and in general following a principle of  
413 conservatism. The selection of limit values from different sources could have introduced a bias in  
414 the risk assessment process, thus the assessor should have sufficient knowledge on this topic and  
415 should ensure that the selected threshold value should be appropriate.

416 Despite these limitations, this study has several strengths, as it provides important insights into the  
417 risk assessment procedure for panellists exposed to odours sample in dynamic olfactometry, an  
418 unusual occupational exposure scenario characterized by the impossibility of applying the classic  
419 hierarchy of risk management and mitigation measures (i.e., elimination / replacement of the risk  
420 agent, confinement of the risk agent, use of collective protection devices and use of personal  
421 protection equipment).

422 In summary, (i) to the best of our knowledge, this is one of the first studies aimed to propose a  
423 standardized approach for risk assessment in dynamic olfactometry. Further, (ii) the proposed  
424 approach allows the quantification of synthetic and conservative risk indices, which in turn can be used  
425 to calculate the minimum dilution factor useful for protecting the health of exposed subjects.

426 The method proposed in this study may need to be updated in the future, according to the evolution of  
427 knowledge on the topic and results obtained in real exposure scenarios. However, at the current state,  
428 this study could represent an advancement of the knowledge about the risk related to occupational  
429 exposure in dynamic olfactometry.

# Occupational exposure risk for olfactometric panellists



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433

### 434 3 Conclusions

435 Dynamic olfactometry is one of the most widely applied techniques to quantify odour concentration.  
 436 The analysis is conducted directly involving human examiners. During the analysis of odorous  
 437 samples, these panellists are directly exposed, at increasing concentration, to odour and the

Fig. 1 Process of decision-making proposed

438 pollutants potentially toxic it contains. Therefore, the evaluation of the occupational exposure risk of  
439 these workers is fundamental to conduct dynamic olfactometry under safe conditions. Despite the  
440 importance of this problem, nowadays a standardized method is not described in the standards  
441 available about dynamic olfactometry (EN 13725:2003 and prEN 13725:2019). In addition, in the  
442 scientific literature, only two papers have deepened this argument, proposing two models for  
443 conducting this evaluation. However, the proposed models do not consider some critical aspects  
444 that may be observed during the analysis of real odour samples. In particular, the models suggest  
445 using different toxicological databases and don't describe how to consider compounds for which  
446 there is no toxicity threshold. Therefore, the aim of this work has been to propose a general and  
447 simplified, but robust method for the assessment of occupational exposure risk for panellists involved  
448 in dynamic olfactometry. The method proposed in this article is based on the assessment of  
449 carcinogenic and non-carcinogenic risks using values established by international regulations and  
450 commonly applied reference standards in the field of toxicology. Applying common and easily  
451 available assessment methods based on properly selected occupational limit values, this  
452 assessment method can be applied by different olfactometric laboratories to assess the occupational  
453 risk of their examiners, achieving uniformity of results. In addition, several reference databases were  
454 proposed for the assessment of non-carcinogenic effects, with a hierarchy between them, based on  
455 the adequacy of available limits, to obtain toxicological information for the largest number of  
456 compounds present in a real odour sample. By minimising the number of compounds for which no  
457 hazard information is available, it is possible not to underestimate the potential exposure risk. To  
458 assess the minimum level of dilution to be adopted, the proposed model uses the calculation of HI  
459 and IR. Although simplified, the HI calculation appears, in according to the particular exposure of  
460 examiners to pollutants during testing activities, conservative and can be easily applied to the case  
461 of occupational exposure of examiners involved in dynamic olfactometry. Therefore, the model  
462 proposed in this paper appears to be a practical solution, combining a robust occupational risk  
463 assessment procedure and an easily implementable approach. This is particularly important  
464 considering the need to have, regarding this issue, a procedure that can be used by  
465 various operators.

466

## 467 **Declaration of interest**

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469 The authors declare no conflict of interest.

470

## 471 **Author Contribution**

472

473 Elisa Polvara: Investigation, Visualisation, Writing – Original draft preparation; Andrea Spinazzè:  
474 Methodology, Visualisation, Writing – Original draft preparation; Marzio Invernizzi: Writing - Review  
475 & Editing; Andrea Cattaneo: Writing - Review & Editing; Selena Sironi: Writing - Review & Editing,  
476 Supervision, Project administration; Domenico Cavallo: Writing - Review & Editing, Supervision.

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