

1 **Contribution to Circular Economy options of mixed agricultural wastes management: Coupling**  
2 **anaerobic digestion with gasification for enhanced energy and material recovery**

3  
4 **Antoniou, N.<sup>1</sup>, Monlau, F.<sup>2</sup>, Sambusiti, C.<sup>3</sup>, Ficara, E.<sup>4</sup>, Barakat, A.<sup>3,5</sup>, Zabaniotou, A.<sup>1,\*</sup>**

5 <sup>1</sup> Biomass Group, Dept. of Chemical Eng., Aristotle University of Thessaloniki, Greece.

6 <sup>2</sup> APESA, Plateau technique, Lescar, France

7 <sup>3</sup>UMR, IATE, CIRAD, Montpellier SupAgro, INRA, Université de Montpellier, France.

8 <sup>4</sup> Politecnico di Milano, DICA, Milano, Italy.

9 <sup>5</sup>AgroBiosciences Department, Mohammed VI Polytechnic University Ben Guerir, Morocco.

10  
11 **Abstract.**

12 Anaerobic digestion (AD) is an established process for the treatment of organic wastes and the  
13 production of renewable energy. However, high amounts of digestate produced by AD plants require  
14 enhancement for further use. This study investigates a conceptual model for the digestate enhancement  
15 by using a downstream gasification. It is based on a ‘systemic approach’ considering the interactions  
16 of every contributing process into the dual system. The digestate was provided by an Italian AD plant,  
17 that treats mixed agricultural wastes of pig manure (43%), cow manure (20%), maize and  
18 triticale silages (25%), and cereal bran (12%). Digestate air gasification experiments were conducted,  
19 in a downdraft fixed-bed reactor, at temperature range from 750 °C to 850 °C, with  $\lambda$  varying from  
20 0.14 to 0.34. Results have shown that gasification of digestate at 850 °C with  $\lambda = 0.24$ , increased  
21 producer gas yield (65.5 wt %), and its LHV (2.88 MJ Nm<sup>-3</sup>). The gas is classified as medium heating  
22 value fuel, suitable to generate electricity of 971 kWhel day<sup>-1</sup> to enhance the AD plant's economic  
23 viability. A carbonaceous material rich in macronutrients (P, K, Ca, Mg) was produced, with  
24  $R_{50} = 0.48$ , suitable for carbon sequestration. The study offers a resource closed loop approach of  
25 converting AD digestate into energy and soil fertilizer. Useful suggestions for policy makers and  
26 business can be drawn.

27  
28 **Keywords:** Anaerobic digestion, digestate, gasification, industrial symbiosis, circular economy, char,  
29 energy.

30 *Corresponding authors: Anastasia Zabaniotou, professor, [azampani@gmail.com](mailto:azampani@gmail.com)*

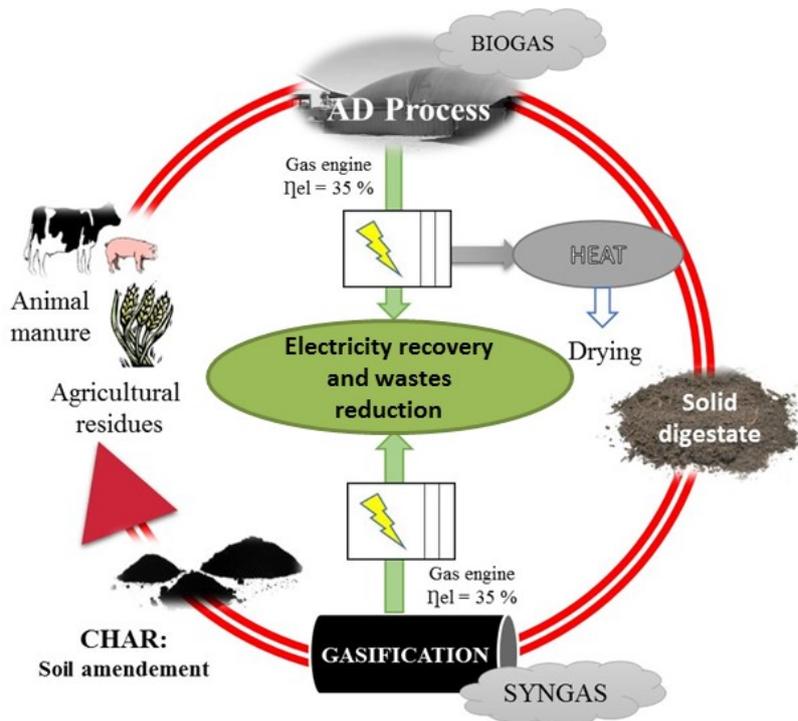
31

32

33

34

35 **Graphical abstract**



36

37 **Highlights**

- 38 • A Circular Economy concept of AD and gasification system was investigated.
- 39 • Industrial symbiosis of AD with gasification increased electricity production by 12%.
- 40 • Solid digestate drying requirements fulfilled by AD derived surplus heat.
- 41 • Optimal digestate gasification conditions were: 850 °C and  $\lambda=0.24$
- 42 • Carbonaceous gasification derived char was rich in nutrients.
- 43 • Char's  $R_{50} = 0.48$  makes it suitable for carbon sequestration.

44

45

46 **Abbreviations**

47 AD Anaerobic Digestion

48	CHP	Cogeneration Heat and Power
49	DM	Dry Matter
50	HRT	Hydrolytic Retention Time
51	OLR	Organic Loading Rate
52	$\lambda$ (ER)	Equivalence ratio
53	VFAs	Volatile Fatty Acids
54	VS	Volatile Solids

## 55 **1 Introduction**

56

57 The European Commission, Council and Parliament provisionally agreed the Circular  
 58 Economy Package of measures in wastes. This agreement moves the EU towards a higher level of  
 59 sustainability in waste management. Anaerobic digestion (AD) is a key process for developing a  
 60 Circular Economy. By closing the loops on the previously linear processes, AD can tackle waste,  
 61 energy, sustainable food production and nutrient recycling challenges in a sustainable and circular  
 62 manner. New ways could strengthen the AD industry as part of developing a Circular Economy.

63 In many European countries, anaerobic digestion (AD) plays an important role in the agro-industrial  
 64 sector, displacing emission-intensive waste management strategies such as landfilling ([Moretti et al.,](#)  
 65 [2018](#)). AD of agricultural residues is an established process, that not only contributes to the reduction  
 66 of greenhouse gas emissions from the agricultural sector, but also produces biogas, a biofuel that can  
 67 be used either to produce heat and electricity, or, after upgrading, to be injected into the natural gas  
 68 grid ([Monlau et al., 2013a](#), [Sambusiti et al., 2013](#)).

69 AD process seems to be a promising route for the treatment of organic wastes ([Santi et al., 2015](#)), but  
 70 widespread uptake of small scale AD is limited due to economic costs and the safe disposal of  
 71 digestate ([Fuldauer et al., 2018](#)). In addition, AD solves only partially the problem of material and  
 72 energy recovery, since a significant part of the organic matter (*i.e.* polysaccharides, lignin) remains in  
 73 the so-called “digestate” ([Monlau et al., 2015a](#), [Santi et al., 2015](#)), not reaching to the complete  
 74 recovery of the organic value existing in the wasted matter. The efficiency of the organic matter (OM)  
 75 conversion through mesophilic (*i.e.* 35 °C) or thermophilic (*i.e.* 55 °C) AD is generally in the range of  
 76 13–65%, and depends on the type of substrate fed to the digester, as well as on anaerobic  
 77 reactor parameters, such as the organic loading rate (OLR) and the hydraulic retention time (HRT)  
 78 ([Monlau et al., 2015a](#)). Legislation framework, does not help very much, because although it promotes  
 79 AD broader implementation across the European continent, clear and solid directions on further  
 80 valorisation of the obtained digestate, are not provided ([Saveyn and Edder, 2014](#)).

81 Digestate is rich in nitrogen, phosphorous and stabilized carbon. Its conventional valorisation route  
 82 lies in its utilization as soil amendment and/or as fertilizer, under the condition that it fulfils the  
 83 restrictions set by European Nitrate Directive (91/676/EEC), especially in the cases of livestock  
 84 intensive areas ([Nitrate Directive, 1997](#)). Due to the accumulation of biogas plants in certain regions,

85 in which intensive livestock farming is encountered, an oversupply of digestate is expected  
86 ([Kuligowski and Luostarinen, 2011](#), [Lacroix et al., 2014](#)). If these agricultural areas cannot fully  
87 process large quantities of digestate, the surplus material must be transported to regions with nutrients  
88 deficit, thus increasing the operational cost of a biogas plant. Digestate composition depends on the  
89 input materials (feedstock) characteristics and the AD process conditions and needs enhancement for  
90 sustainable further uses. Digestate enhancement technological options need to be capable of dealing  
91 with a large range of inputs and feedstocks and low-cost, to achieve significant market penetration.  
92 The valorisation of anaerobic digestate via thermal processes (*i.e* combustion, pyrolysis, gasification)  
93 is gaining interest ([Monlau et al., 2015b](#), [Sheets et al., 2015](#)). Thermo-chemical processes are certainly  
94 interesting and complementary to anaerobic digestion. Anaerobic digestion (AD) is best suited to  
95 organic, putrescible waste streams, while thermochemical is currently best suited to highly  
96 homogenous dry materials (dried fibrous digestate). Research studies have reported the successful use  
97 of solid digestate as a fuel for combustion, considering emissions and the overall combustion  
98 behaviour ([Kratzeisen et al., 2010](#), [Pedrazzi et al., 2015](#)). An anaerobic digestion (AD) process  
99 coupled with digestate composting, to produce a soil amendment of good quality was proposed  
100 by [Cuadros Blázquez et al. \(2018\)](#). Recently, the involvement of pyrolysis process on the further  
101 valorisation of solid digestate for biofuels production (*i.e* bio-oil, pyrolysis gas), and biochar, has  
102 attracted a lot of attention ([Li et al., 2014](#), [Monlau et al., 2015b](#), [Monlau et al., 2016](#), [Troy et al.,](#)  
103 [2013](#)). The produced bio-oil and gas from pyrolysis, can fuel CHP systems, whereas the biochar, due  
104 to its physico-chemical properties, can be either used complementary on mineral fertilizers, or in soil  
105 preservation methodologies ([Monlau et al., 2015b](#), [Monlau et al., 2015c](#)). In livestock intensive and  
106 isolated agricultural areas, the conjunction of AD plants with thermochemical conversion pathways,  
107 can assure a partial reuse of excess heat, which sometimes is lost, for the drying of the solid digestate,  
108 after a mechanical liquid/solid separation of the digestate ([Monlau et al., 2015a](#), [Monlau et al., 2015b](#)).  
109 From the literature review, it was revealed that only few publications reported the utilization of  
110 gasification for the digestate down-stream processing ([Kuligowski and Luostarinen, 2011](#), [Lacroix](#)  
111 [et al., 2014](#)), and fewer studies modelled the potentials of incorporation of gasification process and  
112 anaerobic digestion in a dual system ([Allesina et al., 2015](#), [Li et al., 2015](#), [Yao et al., 2017](#)).  
113 Gasification is a thermal conversion process, which includes the stages of drying, pyrolysis, char  
114 gasification and combustion, that converts organic or fossil fuel based carbonaceous materials into a  
115 gaseous product, of low to medium heating value ([Basu, 2010](#)). This is achieved by material  
116 conversion at high temperatures (>700 °C), under a controlled amount of oxygen and/or steam.  
117 The gaseous mixture, consisted of hydrogen, carbon monoxide, methane, carbon dioxide and fractions  
118 of light hydrocarbons, can be used either as an energy carrier for clean energy or co-fired with other  
119 fuels in current power systems ([Manara and Zabaniotou, 2014](#), [Zabaniotou et al., 2014](#)). The producer  
120 gas can be used to produce combined heat and power, through a CHP unit ([Zabaniotou, 2014](#)).

121 Aiming to improve the sustainability of the existing AD plants, several related biorefinery schemes  
122 have been investigated, during the last decades (Fabbri and Torri, 2016, Monlau et al., 2015a, Sheets  
123 et al., 2015). In most biorefinery cases, the utilization of the residual polysaccharides and lignin  
124 content of the solid-digestate, resulted to an extra-energy recovery and valuable materials production  
125 (Monlau et al., 2015a, Sheets et al., 2015).

126 The literature review made evident that although, AD is a key process for developing a Circular  
127 Economy and an important pillar of the European Circular Economy and a part of the European bio-  
128 economy improving European resource-efficiency (EBA, 2015), further technological leaps are  
129 needed. In this respect, a conceptual dual system of AD and digestate upgrading via downstream  
130 gasification process is proposed in this study. The study aims to fill that gap in knowledge and  
131 experimental data by investigating the gasification of the dried solid digestate, by-product of  
132 agricultural mixed wastes-based biogas plant. The study focuses on the gasification  
133 parameters optimisation, an evaluation of the agronomic properties of the char produce and an overall  
134 energy balance enhancement of the dual AD/gasification, providing recommendations for its  
135 deployment and further commercialisation.

136 The study contributes to the options of mixed agricultural wastes management, since it promotes the  
137 implementation of an environmentally-friendly solution, capable of acting either alternatively or  
138 complementary with the traditional agriculture processes, in closed loops. The technical innovations of  
139 this study are based on the materials and energetic fluxes exchanges, such as part of the heat produced  
140 through CHP system, which is used for solid digestate drying. The deployment of this system could  
141 also contribute to the wider implementation of resource efficiency principles, through: (i)  
142 maximization of energy recovery (*i.e* biogas, syngas) from agricultural residues, (ii) efficient reduction  
143 of wastes, (iii) production of a carbonaceous material for various application, such as for soil  
144 preservation and long term carbon sequestration (WRAP, 2012, Zabaniotou et al., 2015).

145

## 146 **2 Materials and Methods**

147

### 148 *2.1 Digestate*

149

150 The solid digestate was provided by a biogas plant located in Italy, with conventional  
151 configuration, including a first digester followed by a post-digester. The digester with the main  
152 characteristics of volume ~5,840 m<sup>3</sup>, OLR of 120 t FM d<sup>-1</sup>, HRT of 53 days and temperature at 45°C,  
153 was daily fed with a mixture of 43% animal sewage, 20% cow manure, 25% maize and triticale silages  
154 and 12% cereal bran. A daily production of 6.7 t DM day<sup>-1</sup> of solid-digestate was obtained. Prior to  
155 any analysis and gasification trials, the sample was oven dried overnight at 105°C. Table 4  
156 summarized the main characteristics of the AD plant.

157

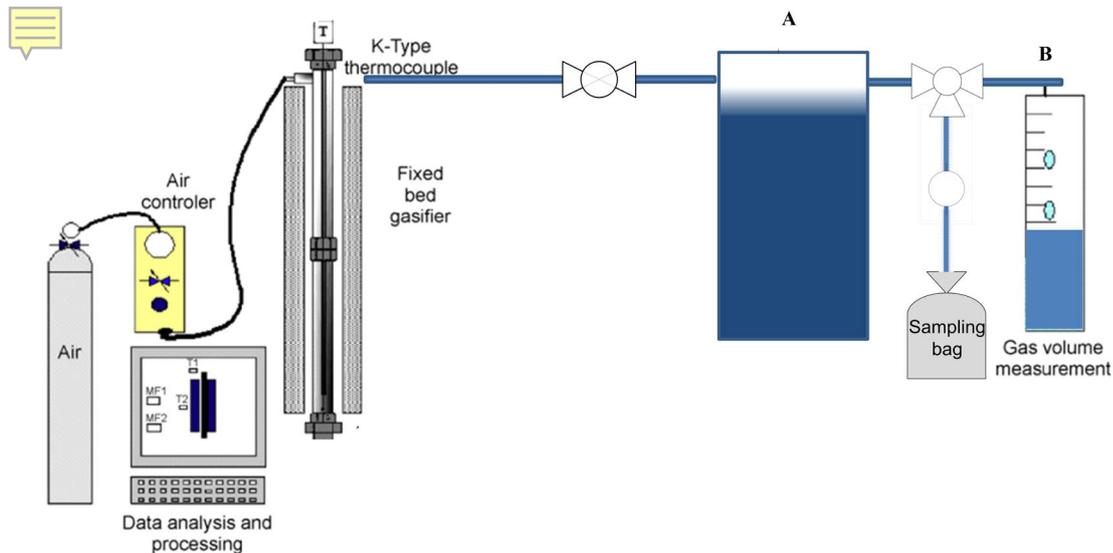
158 2.2 Gasification protocol

159

160 In this study, gasification was carried out at medium to high temperatures, using air as  
161 gasifying agent, in a laboratory scale downdraft fixed-bed gasifier, at ambient pressure (Skoulou et al.,  
162 2008). The gasification system consisted of:

- 163 (i) stainless-steel reactor (height=500mm, diameter=12.5mm),  
164 (ii) an individually controlled vertical electric furnace, with temperature measurement and  
165 modification in three different spots across the tube,  
166 (iii) medium (N<sub>2</sub> or Air) providing system, using a vertical pipe in a downward flow, not preheated  
167 before entry (ambient conditions) and  
168 (iv) gas collection system (Fig. 1).

169



170

171 **Figure 1.** Downdraft fixed bed gasifier.

172

173 Gasification experiments were conducted at a temperature range of 750–850°C. Experiments  
174 were replicated twice, to achieve repeatability of results, and reproducibility of the process. Prior to  
175 each experiment, the reactor was dismantled, and the digestate, in powder form, was fed manually, by  
176 batches, from the top of the reactor. The reactor was sealed and nitrogen was purged continuously for  
177 30 min, to remove the included air. The electric furnace was heated up (heating rate=30°C min<sup>-1</sup>) to the  
178 desired temperature. By the time the desired temperature (750, 800 or 850 °C) was reached, the reactor  
179 was placed vertically on it. Temperature, measured by a K-type thermocouple (NiCr-Ni), increased  
180 and the gas medium was switched from nitrogen to air, when its instant temperature reached the  
181 desired one. During gasification, the produced gas through water displacement was collected in bottle  
182 A. When gasification ended, the produced volume was measured in bottle B, and the gas, through a

183 gas pump, was transferred into a gas bag for GC analysis. Gas samples were taken after a few minutes  
184 for chromatographic analysis and at two additional time intervals.

185 The produced gas composed of CO, CO<sub>2</sub>, H<sub>2</sub>, CH<sub>4</sub>, and other gaseous hydrocarbons (H<sub>x</sub>C<sub>y</sub>). The  
186 determination of the composition of the producer gas was performed offline, in a gas chromatographer  
187 (Model GC 6890N, Agilent Technologies), fitted with two columns, HP-PlotQ and HP-Molsive type  
188 (Manara and Zabaniotou, 2013). After the completion of the experiment, the reactor was removed  
189 from the furnace, left to cool down, dismantled and the produced ash was collected and weighted,  
190 whereas tar was collected on a solvent and its yield was estimated. Operational parameters such as  
191 temperature and  $\lambda$  were studied.  $\lambda$  is defined as the ratio of the actual amount of oxidising agent  
192 provided for oxidation to the theoretical amount required for complete combustion (Zabaniotou et al.,  
193 2014). Based on the elemental analysis of the raw material, the desired  $\lambda$  value for air gasification and  
194 the air-flow rate, the required air volume and consequently the air mass was determined.

195

### 196 2.3 Analytical procedure

197

198 Total solid (TS), volatile solid (VS) and ash (AS) were analyzed according to the APHA  
199 methods for digestate and char (APHA, 2005). Additionally, elemental analysis (CHNS) was also  
200 performed using an “Elementar Vario Macro Cube” analyser. The higher heating value (HHV) of the  
201 feedstock was calculated using the results from the elemental analysis of the sample and the following  
202 equation (Channiwala and Parikh, 2002):

$$203 \text{HHV (MJ Kg}^{-1}\text{)} = 0.3491*\text{C}+1.1783*\text{H}+0.1005*\text{S}-0.1034*\text{O}-0.0151*\text{N}-0.0211*\text{Ash} \quad (1)$$

204 Digestate and char pH determination was carried out by adding de-ionized water in a mass  
205 ratio 1:20. The solution was then hand shaken and allowed to stand for 5 min before measuring pH  
206 using a Basic Crison 20<sup>®</sup> pH meter. *Structural-carbohydrates* from cellulose and hemicelluloses  
207 together with “klason lignin” of solid digestate were measured, using a strong acid hydrolysis method,  
208 previously described by Monlau et al. (Monlau et al., 2015b). All monosaccharides (*i.e.* glucose,  
209 xylose, arabinose) were analysed by HPLC (Agilent<sup>®</sup> 1260) coupled to refractometric detection. The  
210 analysis was carried out with a Hi-PLex H column at 50°C. The eluent corresponded to 5 mM H<sub>2</sub>SO<sub>4</sub>  
211 under a flow rate of 0.3 mL min<sup>-1</sup>. A refractive index detector was used to quantify the carbohydrates.  
212 The system was calibrated with glucose, xylose and arabinose standards (Sigma–Aldrich<sup>®</sup>).  
213 Thereafter, cellulose and hemicelluloses contents were estimated as follows:

$$214 \text{Cellulose (\%TS)} = \text{Glucose (\%TS)} / 1.11 \quad (2)$$

$$215 \text{Hemicelluloses (\%TS)} = [\text{Xylose (\%TS)} + \text{Arabinose (\%TS)}] / 1.13 \quad (3)$$

216 Where:

217 1.11 is the conversion factor for glucose-based polymers (glucose) to monomers and 1.13 is the  
218 conversion factor for xylose-based polymers (arabinose and xylose) to monomers.

219 Elemental analysis (micro- and macro-elements) was performed by digesting 38.5 mg of sample  
 220 in 2 mL of HNO<sub>3</sub> (68%) and 5mL of distilled water, in a closed vessel microwave digester (*µondes*  
 221 *US49*), for three temperature cycles of 140, 170 and 190°C, respectively. The metals in the solution  
 222 were analysed by inductively coupling plasma spectrometry-optical emission spectroscopy (ICP-OES  
 223 Agilent 720).

224 To evaluate the carbon sequestration potential of both digestate and char, the R<sub>50</sub> coefficient was  
 225 calculated. R<sub>50</sub> corresponds to an index for quantifying char recalcitrance and screening char with  
 226 respect to their carbon sequestration potential (Harvey et al., 2012). The R<sub>50</sub> uses the energy required  
 227 for thermal oxidation of the char (normalized to that needed for the oxidation of graphite) as a measure  
 228 of recalcitrance. The R<sub>50</sub> recalcitrance index was calculated according to the following equation:

$$229 R_{50Ch} = T_{50Ch} / T_{50 GRAPH} \quad (4)$$

230 Where:

231 T<sub>50Ch</sub> and T<sub>50 GRAPH</sub> are the temperature values corresponding to 50% oxidation/volatilization of  
 232 chars (Ch) and graphite, respectively.

233 Values for R<sub>50Ch</sub> are obtained directly from TG thermograms appropriately corrected for water  
 234 and ash content. T<sub>50GRAPH</sub> was assumed to be 886 °C (Harvey et al., 2012). The R<sub>50</sub> was interpreted  
 235 considering the following recalcitrance/carbon sequestration classes: Class A (R<sub>50</sub> ≥ 0.70), Class B  
 236 (0.50 ≤ R<sub>50</sub> < 0.70), or Class C (R<sub>50</sub> < 0.50).

237

#### 238 2.4 Energy analysis

239

240 The energetic requirements for drying the solid digestate were estimated based on data provided  
 241 by the owner of the AD plant. The energetic requirements for drying the solid digestate E<sub>DD</sub> (kWh<sub>th</sub>  
 242 day<sup>-1</sup>) on a daily basis, was calculated using the modified equation used by Barakat et al. (Barakat et  
 243 al., 2014).

$$244 E_{DD} = E_{Heat} + E_{Evaporation} \quad (5)$$

245 where:

246 E<sub>Heat</sub> (kWh<sub>th</sub> day<sup>-1</sup>) is the energetic requirement to increase the temperature of water and digestate from  
 247 25°C to 105°C and

248 E<sub>Evaporation</sub> (kWh<sub>th</sub> day<sup>-1</sup>) corresponds to the energetic requirement for water evaporation at 105°C.

249 E<sub>Heat</sub> and E<sub>Evaporation</sub> were calculated according to the equations 6 and 7:

$$250 E_{Heat} = m \times Cp \times [T_{Final} - T_{Initial}] / 3600 \quad (6)$$

251 m (kg day<sup>-1</sup>) is the daily mass of water and solid digestate;

252 Cp is the water specific heat (4.18 kJ kg<sup>-1</sup>°C<sup>-1</sup>);

253 T<sub>Initial</sub> (°C) is the initial temperature of the substrate suspension, assumed as 25°C;

254 T<sub>Final</sub> (°C) is the final temperature at 105°C.

$$255 E_{Evaporation} = [m_{water} \times Lv] / 3600 \quad (7)$$

256  $L_v$  is the latent heat of vaporization equal to 2257 kJ kg<sup>-1</sup>;  
257  $m$  water (kg day<sup>-1</sup>) is the daily mass of water in the solid digestate.

258 The lower and higher heating value (MJ m<sup>-3</sup>) of the producer gas were calculated using the  
259 molar fractions of the detected compounds in the produced gas and the following equations (8 and 9)  
260 (Li et al., 2004; Lv et al., 2004) :

$$261 \text{LHV} = [30 \times v/v\% \text{CO} + 25.7 \times v/v\% \text{H}_2 + 85.4 \times v/v\% \text{CH}_4 + 151.3 \times v/v\% (\text{C}_2\text{H}_4 + \text{C}_2\text{H}_6)] \times 0.42$$

262 (8)

$$263 \text{HHV} = [12.75 \times v/v\% \text{H}_2 + 12.63 \times v/v\% \text{CO} + 39.82 \times v/v\% \text{CH}_4 + 63.43 \times v/v\% (\text{C}_2\text{H}_4 + \text{C}_2\text{H}_6)] / 100$$

264 (9)

265 Furthermore, additional qualitative indicators were introduced to provide a more deliberate  
266 characterisation of the obtained gasification gas. These included syngas yield (H<sub>2</sub>+CO), and H<sub>2</sub>/CO  
267 ratio. The H<sub>2</sub>/CO ratio is an important parameter for a further utilization of producer gas as a feedstock  
268 in the synthesis of chemicals, such as methanol or Fischer Tropsch fuels. H<sub>2</sub>/CO ratio desired values  
269 were considered those which exceed 2:1, depending on selectivity (Wender, 1996). In general, a gas  
270 with high syngas yield and high H<sub>2</sub>/CO ratio is a gas with a high calorific value, suitable as fuel and  
271 for a wide range of applications.

272 Finally, to estimate the energetic balance of the entire process, it was assumed that biogas and  
273 syngas produced from AD and gasification processes respectively were used to fuel a CHP unit. The  
274 energetic efficiency of the conversion system was considered as 35 % for electricity and 50% for heat  
275 (Monlau et al., 2013b). Furthermore, it was assumed that heat wastes (exhaust gases and hot cooling  
276 water) from the CHP system can be used to cover the thermal needs of the digestate process.

277

### 278 3 Results

279 Results on gasification producer gas, char yields and their quality are discussed here; yields and  
280 quality depend on the digestate characteristics. Energy analysis is also provided.

#### 281 3.1 Digestate's characteristics

282

283 The chemical composition of the AD digestate used in this study, is presented in **Table 1**. The  
284 digestate was composed of 17.0 g 100 g<sup>-1</sup> TS of cellulose, 12.2 g 100 g<sup>-1</sup> TS of hemicelluloses and 34 g  
285 100 g<sup>-1</sup> TS of lignin. It contained a high C content of 43 g 100 g<sup>-1</sup> TS and ash content of 9.5 g 100 g<sup>-1</sup>  
286 TS. The results are in agreement with reported results by Santi (Santi et al., 2015), showing that an  
287 important quantity of cell wall polymers remained in the digestate fraction.

288

289

290 **Table 1.** Solid digestate analysis

Parameters (Units)	Digestate
pH	8.6 ( $\pm 0.04$ )
VS (%TS)	89.5 ( $\pm 0.1$ )
Ash (%TS)	9.5 ( $\pm 0.1$ )
C (g 100 g <sup>-1</sup> TS)	43.0 ( $\pm 0.07$ )
H (g 100 g <sup>-1</sup> TS)	6.2 ( $\pm 0.03$ )
N (g 100 g <sup>-1</sup> TS)	1.3 ( $\pm 0.03$ )
S (g 100 g <sup>-1</sup> TS)	0.14 ( $\pm 0.05$ )
O (g 100 g <sup>-1</sup> TS) <sup>a</sup>	39.5
Cellulose (g 100 g <sup>-1</sup> TS)	17.0 ( $\pm 0.9$ )
Hemicellulose (g 100 g <sup>-1</sup> TS)	12.2 ( $\pm 1.7$ )
Lignin (g 100 g <sup>-1</sup> TS)	34.0 ( $\pm 1.2$ )

291

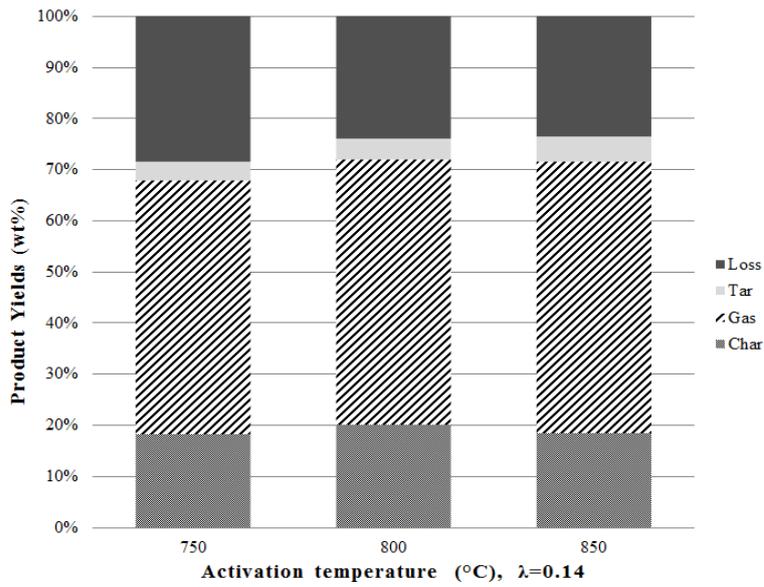
292

293 *3.2 Gasification products: effect of temperature and  $\lambda$  on products yield*

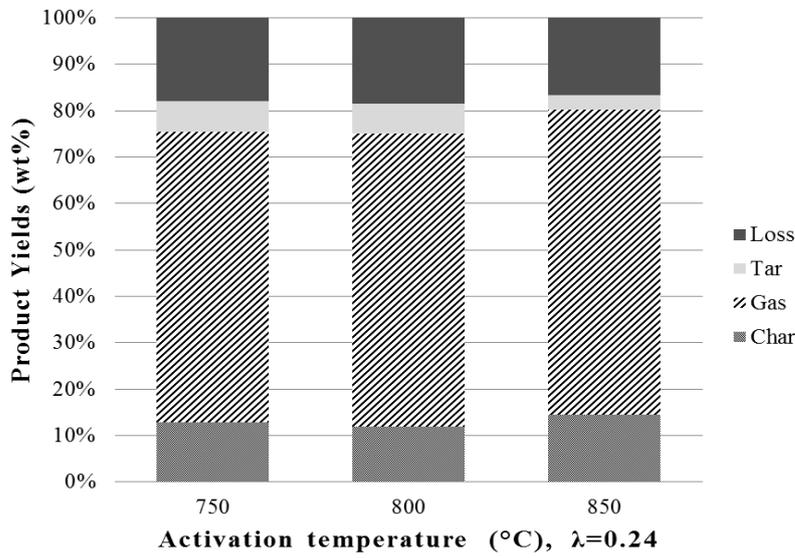
294

295 Digestate was subjected to gasification at 750, 800 and 850°C, using air as gasification agent, at  $\lambda$   
 296 values ( $\lambda = \text{Air}_{\text{experimental}}/\text{Air}_{\text{theoretical}}$ ) ranging from 0.14 to 0.34. The effect of process parameters (T and  
 297  $\lambda$ ) on product yields is depicted in **Fig. 2**.

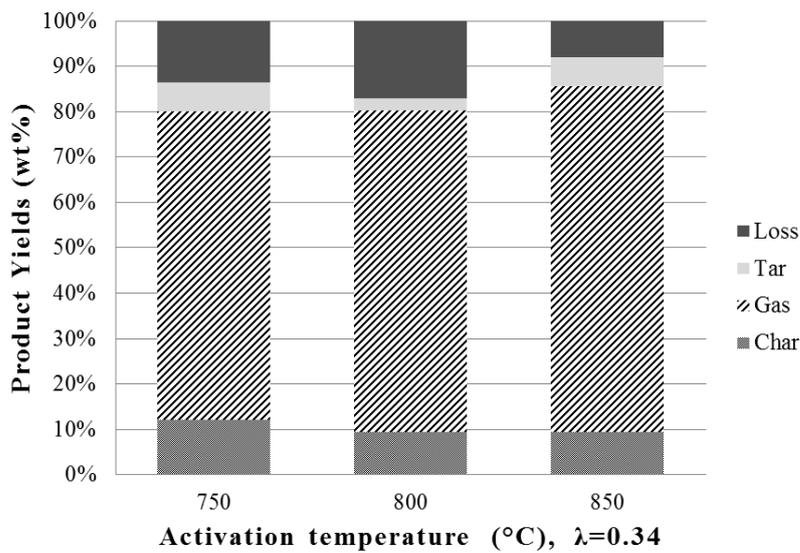
298



299



300

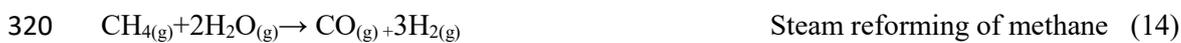
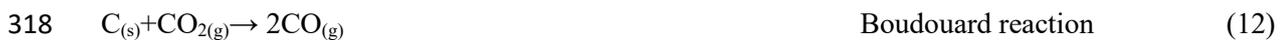
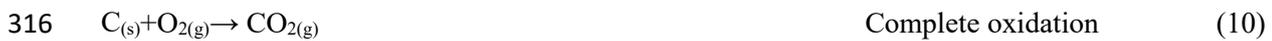


301

302 **Figure 2.** Effect of activation Temperature and  $\lambda$  on a product yields (wt%) and mass balances.

303 The experimental results highlighted the major effect of temperature on product yields. At 750  
 304 and 800°C the char and gas yields remained almost constant at every  $\lambda$ -value studied. A further  
 305 temperature increase favoured the production of gaseous product, because of the extensive  
 306 devolatilization of the digestate, thus reducing char yield. Moreover, due to the intense temperature  
 307 applied (850°C), devolatilization and oxidation of condensed tar also occurred, thus reducing tar yield,  
 308 resulting in the presence of condensed hydrocarbons in the interior of the reactor. As a result, under  
 309 higher temperature, higher carbon conversion was achieved (Devi et al., 2003). At the lowest  $\lambda$  values  
 310 used, high temperature pyrolysis (cracking) occurred rather than gasification (oxidation). An increase  
 311 of  $\lambda$  resulted in further devolatilisation (volatiles released from the solid phase) and oxidation of the  
 312 sample, thus decreasing char yield and tar yield (wt%) in favour of gaseous products.

313 A simplified mechanism describing the evolution of permanent gases during gasification is  
 314 presented below (Cao et al., 2006; Chen et al., 2013; Fryda et al., 2008; Nipattummakul et al., 2010;  
 315 Panopoulos et al., 2006):



322

### 323 3.3 Gasification products: effect of temperature and $\lambda$ on producer gas composition

324

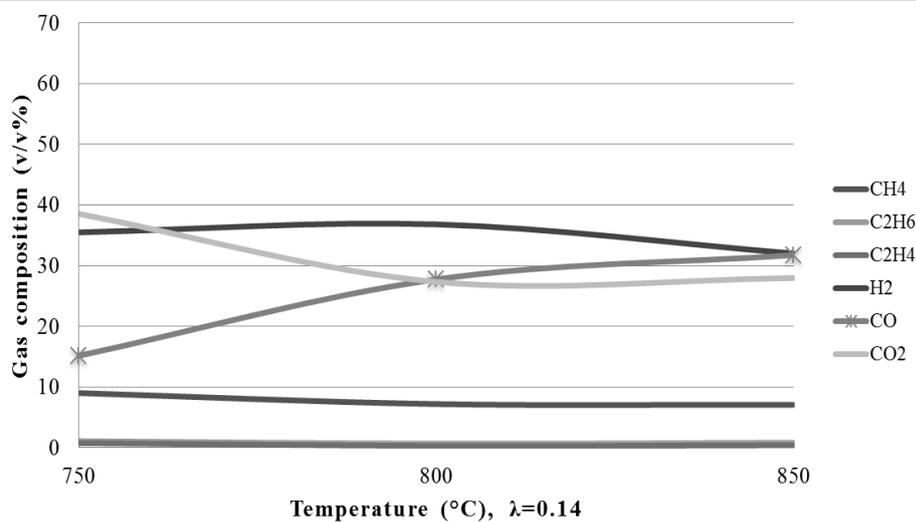
325 For the gasification experiments,  $\lambda$  varied between 0.2 and 0.3. With a lower  $\lambda$  value, the char  
 326 was not fully converted into gases, giving rise to higher tar production, resulting in several problems  
 327 including incomplete gasification and excessive char formation. Gasification at a  $\lambda >0.25$ , resulted  
 328 mostly in gaseous products. At higher temperatures, further oxidation was resulted. However, a high  $\lambda$   
 329 value resulted to an excessive formation of undesired  $CO_2$ , reducing the heating value of the producer  
 330 gas (Basu, 2010). The effect of process parameters on the composition of the gaseous product is  
 331 depicted in Fig. 3.

332 During gasification at the lower temperature, oxidation reactions (exothermic) took place,  
 333 resulting to  $CO_2$  and CO. As temperature increased, Boudouard reaction towards the production of CO  
 334 in expense of  $CO_2$ , was favoured. Under elevated temperatures,  $CH_4$  reforming occurred, increasing  
 335 slightly the generation of CO and  $H_2$ . Likewise, under the same conditions, reforming of  $C_2H_4$  and  
 336  $C_2H_6$  was also noticed for all  $\lambda$ -values studied.

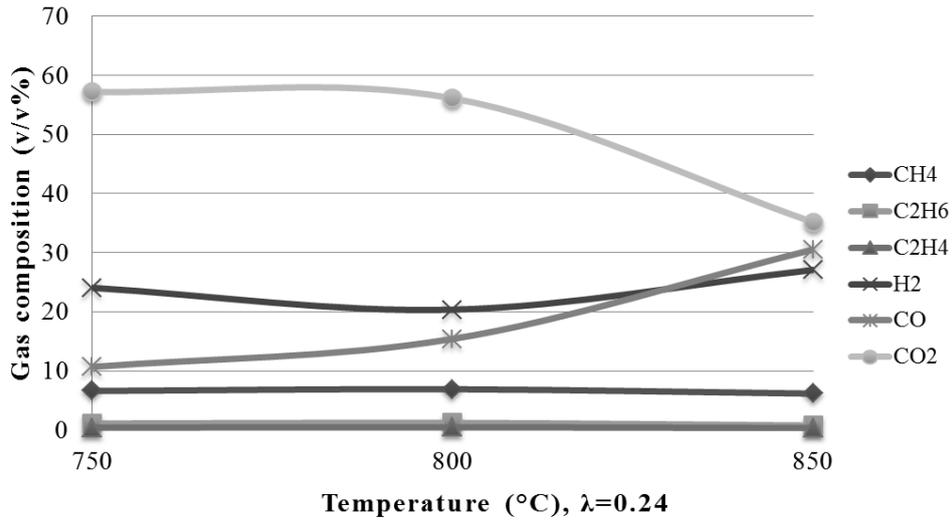
337 The quantitative characteristics of the produced gas, as identified by metrics such as syngas  
338 yield, LHV, HHV and H<sub>2</sub>/CO ratio, are presented in **Table 2**. Producer gas yield increased as  
339 temperature increased (reactions 11-15), at all  $\lambda$  values. On the other hand, syngas yield decreased  
340 when  $\lambda$  values increased, under the same temperature (reaction 10).

341 The aforementioned process parameters greatly affected the heating value of the obtained gas.  
342 Over the studied temperature range (750 up to 850 °C), the LHV and HHV varied from 1.13 up to 4.25  
343 MJ m<sup>-3</sup> and from 1.24 to 4.58 MJ m<sup>-3</sup>, respectively. It seems that a higher temperature favoured H<sub>2</sub> and  
344 CO production (reactions 11-15). Both LHV and HHV showed better values with a  $\lambda=0.14$ . This can  
345 be attributed to the fact that under these conditions ( $\lambda=0.14$ ), the predominant step was high  
346 temperature pyrolysis, thus producing a gas of high calorific content.

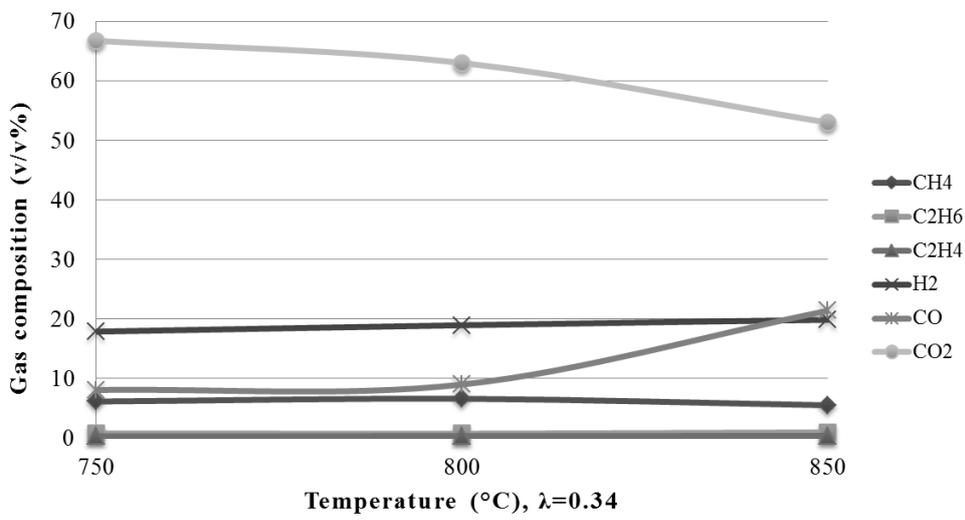
347



348



349



350

351 **Figure 3.** Effect of gasification temperature and  $\lambda$  values on gas composition

352 **Table 2.** Producer gas characteristics.

Temperature °C	750	800	850
<b><math>\lambda = 0.14</math></b>			
syngas yield (%v/v)	17.58	23.73	25.18
H <sub>2</sub> /CO	2.34	1.32	1.01
LHV (MJ m <sup>-3</sup> )	3.52	3.92	4.25
HHV (MJ m <sup>-3</sup> )	3.87	4.27	4.58
<b><math>\lambda = 0.24</math></b>			
syngas yield (% v/v)	8.64	9.15	17.27
H <sub>2</sub> /CO	2.25	1.36	0.90
LHV (MJ m <sup>-3</sup> )	1.80	1.97	2.88
HHV (MJ m <sup>-3</sup> )	1.98	2.13	3.10
<b><math>\lambda = 0.34</math></b>			
syngas yield (% v/v)	5.05	5.40	10.25
H <sub>2</sub> /CO	2.21	2.20	0.88
LHV (MJ m <sup>-3</sup> )	1.13	1.22	1.88
HHV (MJ m <sup>-3</sup> )	1.24	1.34	2.01

353

354 At higher  $\lambda$ , the production of CO<sub>2</sub> through favoured combustion reactions, increased. The  
 355 presence of CO<sub>2</sub> in the gasification gas is not desirable, since it implies both a dilution effect of the gas  
 356 heating value and a reduction in the formation of CO, (reactions of production and consumption of CO  
 357 and CO<sub>2</sub> are the water-gas shift or the Boudouard reactions).

358 The H<sub>2</sub>/CO ratio is a critical value for the choice of producer valorisation pathway. This ratio  
 359 exhibited a maximum value (2.2-2.3) at lower temperature, for all  $\lambda$  values studied. With increasing  
 360 temperature, CO produced at the expense of CO<sub>2</sub>, thus reducing the H<sub>2</sub>/CO ratio in values <1, not  
 361 recommended for synthesis of chemicals, or Fischer Tropsch fuels. Based on the experimental results,  
 362  $\lambda=0.24$  and temperature~850°C were considered as the optimal conditions for a gaseous fuel  
 363 production to be used for CHP.

364

### 365 3.4 Char characteristics

366 In order to use gasification solid remained by-product (char) in soil amendment applications,  
 367 an extensive analysis of chars was performed (Zabaniotou, 2014). During the last decades, biochar has  
 368 attracted attention as soil amendment for soil properties improvement with valuable nutrients (P, K,  
 369 Ca, Mg), contributing to carbon sequestration, (Kuligowski and Luostarinen, 2011; Monlau et al.,

2016; Srinivasan et al., 2015). Literature data on the agronomic properties of char derived from the gasification of solid digestate, are scarce (Hansen et al., 2015).

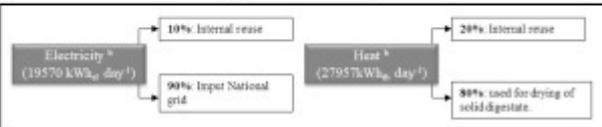
In this study, the digestate showed an alkaline pH (8.6). Digestate alkalinity is commonly observed and it is mainly due to the degradation of VFAs and the production of ammonia during the AD process. Control of pH can improve the addition of basic compounds or carbonates into the digester (Tambone et al., 2010). As Table 3 depicts, the gasification-char exhibited a high pH value of 11.4, which is in agreement with reported values (Monlau et al., 2016; Opatokun et al., 2017), VS of 61.8% TS and ash content of 38.2 ( $\pm$  2.6) % TS. The content of C on the gasification-char was found higher by 50% in the dry mass (DM) than the digestate, which is in agreement with the European Biochar Certificate (EBC, 2012).

**Table 3.** Proximate, ultimate analysis and concentrations of nutrients and heavy metals of the digestate and gasification char (compared with standards).

Compounds	Digestate (Monlau et al., 2016)	Char	EBC guidelines	IBI standard
pH_1:25 (H <sub>2</sub> O)	8.6 ( $\pm$ 0.04)	11.4 ( $\pm$ 0.01)		
VS (% TS)	89.5 ( $\pm$ 0.1)	61.8 ( $\pm$ 2.5)		
Ash (% TS)	9.5 ( $\pm$ 0.1)	38.2 ( $\pm$ 2.6)		
R <sub>50</sub> (%)	34	47.6		
C (g. 100g <sup>-1</sup> TS)	43.0	66.5	> 50	
H (g. 100g <sup>-1</sup> TS)	6.2	8.7		
N (g. 100g <sup>-1</sup> TS)	1.3	1.4		
S (g. 100g <sup>-1</sup> TS)	0.14	0.196		
Mg (mg kg <sup>-1</sup> TS)	3689	19446		
Ca (mg kg <sup>-1</sup> TS)	10814	69139		
Cd (mg kg <sup>-1</sup> TS)	-	0.15	< 1.5	1.4-39
Cr (mg kg <sup>-1</sup> TS)	1	786	< 90	64-1200
Cu (mg kg <sup>-1</sup> TS)	11	94	< 100	63-1500
Fe (mg kg <sup>-1</sup> TS)	630	8971		
K (mg kg <sup>-1</sup> TS)	11966	48132		
Al (mg kg <sup>-1</sup> TS)	356	3418		
Mn (mg kg <sup>-1</sup> TS)	97	550		
Na (mg kg <sup>-1</sup> TS)	1886	16285		
Ni (mg kg <sup>-1</sup> TS)	1	1555	< 50	47-600
P (mg kg <sup>-1</sup> TS)	5289	17400		

Pb (mg kg <sup>-1</sup> TS)	-	48	< 120	70-500
Zn (mg kg <sup>-1</sup> TS)	35	256	< 400	200-7000

383 **Table 4.** Characteristics of the agricultural biogas plant. a) 10 kWh Nm<sup>-3</sup> methane; b) The  
 384 energetic efficiency of the conversion system CHP was considered as 35 % for electricity and 50% for  
 385 heat.

Feedstocks (t/d)	Total: 120 t FM day <sup>-1</sup> 43% pig manure, 20% cow manure, 25% maize and triticale silages and 12% cereal bran	
Operational conditions	HRT: 53 days Digester Volume (m <sup>3</sup> ): 5840 T°: 45°C pH: 7.5-7.8	
Biogas characteristics	Biogas (Nm <sup>3</sup> day <sup>-1</sup> ): 9477 Methane (%): 59 Methane (Nm <sup>3</sup> day <sup>-1</sup> ): 5591 Energy (kWh day <sup>-1</sup> ): 55914*	Solid Digestate production Quantity (t day <sup>-1</sup> ): 27.8 Humidity (%): 76 Water (t day <sup>-1</sup> ): 21.1 Energy drying (kWh day <sup>-1</sup> ): 16561
Energy production		

386  
387

388 The distribution of the main macronutrients (N, P, K, S, Ca, Mg) present in both the digestate  
 389 and its resulting char were presented in **Table 3**. The content of N and S elements remained similar in  
 390 digestate and char; no char enrichment was noticed due to the fact that most N and S compounds  
 391 volatilize above 200°C and 375°C (Kookana et al., 2011). A significant increase of the P (17,400 mg  
 392 kg<sup>-1</sup>TS), K (48,132 mg kg<sup>-1</sup>TS), Ca (69,139 mg kg<sup>-1</sup>TS) and Mg (19,446 mg kg<sup>-1</sup>TS) macronutrients  
 393 content was observed on the char. Similar results have been previously reported (Kuligowski and  
 394 Luostarinen, 2011; Monlau et al., 2016; Opatokun et al., 2017). Kuligowski and Luostarinen  
 395 (Kuligowski and Luostarinen, 2011), reported that the main macronutrients of gasification char (from  
 396 solid digestate of AD plant treating manure) were calcium (311 g kg<sup>-1</sup>), phosphorus (54.4 g kg<sup>-1</sup>) and  
 397 potassium (34.7 g kg<sup>-1</sup>) making the char a good candidate for fertilizer. In parallel, an enrichment of  
 398 micronutrients (Fe, Cu, Mn, Ni, Zn) was noticed on char. This is in agreement with previous studies,  
 399 that reported an enrichment of the main macronutrients (P, K, Mg, Ca) and micronutrients on biochar,  
 400 after pyrolysis of solid anaerobic digestate of food wastes and agricultural wastes respectively  
 401 (Monlau et al., 2016; Opatokun et al., 2017).

402 Toxic compounds can be present in the char, preventing its broader use for agronomic purpose,  
 403 depending on the composition of the precursor material. To assess the possibility to use gasification  
 404 char as soil amender, it is of prime interest to estimate the potential toxicants, especially heavy metals  
 405 content, **Table 3**. Except for Cr and Ni, analyses of heavy metals of most samples were under the  
 406 threshold values recommended by International Biochar Initiative and European Biochar Certificate  
 407 (EBC, 2012; IBI, 2014). The higher concentration of Ni and Cr compared to threshold values is not  
 408 worrying because it was due to some experimental limitations: the high content of Ni and Cr were can

409 be attributed to the degradation of K-type thermocouple during the high temperature process or  
410 leaching from the gasification reactor made up of stainless steel. Finally, a high concentration of Al  
411 (3,418 mg kg<sup>-1</sup>TS) was also estimated, which is in agreement with values reported by Kataki et al.  
412 (Kataki et al., 2017) on char from rice husk gasification.

413 Further research is needed on the characteristics of the char obtained from gasification of solid  
414 anaerobic digestate to assess its implication in soil amendment. It will be interesting to investigate the  
415 bioavailability of the main macronutrients presents in gasification char along with growth plant tests,  
416 to assess their suitability benefit use for agronomic application Opatokun et al. (Opatokun et al.,  
417 2017).

418 For the estimation of the carbon sequestration potential by char in soils, the recalcitrance index  
419 ( $R_{50}$ ) was calculated from  $T_{50}$  values, using the obtained thermogravimetric curves corrected for  
420 moisture and ash content.  $R_{50}$  values of 0.34 and 0.48 were determined for digestate and char,  
421 respectively (Table 3). Such results confirmed the recalcitrant structure of gasification char which is  
422 more apparent in char than in digestate sample, confirming char's potential in carbon sequestration.  
423 The  $R_{50}$  was interpreted considering the following recalcitrance/carbon sequestration classes: Class A  
424 ( $R_{50} \geq 0.70$ ), Class B ( $0.50 \leq R_{50} < 0.70$ ), or Class C ( $R_{50} < 0.50$ ). Considering these  $R_{50}$  values, the  
425 obtained char from gasification should be ranked in the third class. A lower stability was observed  
426 comparing the obtained results with those of biochar from pyrolysis of manure and residues of  
427 agricultural crops (Harvey et al., 2012). Nonetheless, the  $R_{50}$  values were in the same range ( $R_{50}$   
428 =0.40/0.41) with the previously values reported by Monlau et al. (Monlau et al., 2016), on biochar  
429 obtained from pyrolysis (at 600°C) of solid anaerobic digestates.

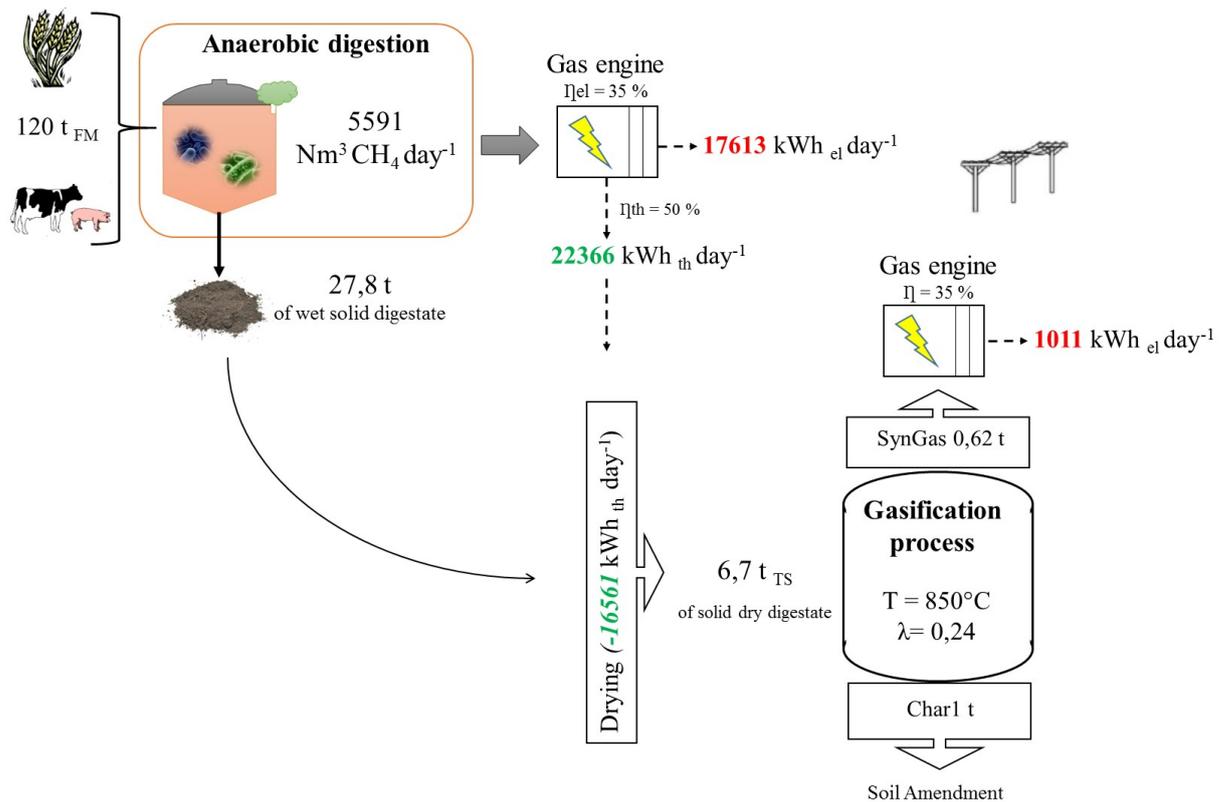
430

### 431 3.5 Dual system's overall energy balance

432

433 In Fig. 4, the overall energy balances of coupling AD with gasification process, in a basis of on  
434 one-day running process is presented. The methodology for the detailed calculations, were described  
435 by Monlau et al. (Monlau et al., 2015b). The biogas produced from the AD plant converted into  
436 electricity and heat through a CHP unit, for which 10% of electricity and 20% of heat were internally  
437 reused for the AD plant operation. A net daily production of 17,613 kWh<sub>el</sub> and 22,366 kWh<sub>th</sub> were  
438 documented.

439



440

441

442 **Figure 4.** Energy balance of coupling AD with gasification process ( base: one-day running process)

443 It was shown that the daily excess heat produced from the CHP unit (22,366 kWh<sub>th</sub>) was  
 444 sufficient for a daily drying of the solid digestate. It was demonstrated that a higher energy recovery  
 445 could be achieved by coupling AD/gasification process, due to avoiding part of heat losses into the  
 446 atmosphere. Such results were in agreement with previous results reported by Monlau et al. (Monlau  
 447 et al., 2015b), demonstrating that waste-heat from the biogas unit was enough to fulfil the drying  
 448 needs for the solid digestate, through a belt dryer system.

449 At the gasification process conditions 850°C and λ=0.24, gas yield exhibited its maximum  
 450 heating value (LHV=2.88 MJ m<sup>-3</sup>), classified as a medium heating value fuel. Based on the above, the  
 451 overall energy balance of the dual AD/gasification process was determined by considering: a daiy  
 452 processing of solid digestate of 6.7 t<sub>TS</sub>; a daily syngaz production of 0.62 tons equivalent through  
 453 conversion into a CHP system to 971 kWh<sub>el</sub> day<sup>-1</sup>..

454 In a similar study on sewage sludge, Lacroix et al. (Lacroix et al., 2014) improved energy  
455 recovery by 90%. The produced 971 kWh<sub>el</sub> per day was a relatively small amount of energy,  
456 appropriate for a small scale decentralised gasification system to be used. Such system, as reported by  
457 Manara and Zabaniotou. (Manara and Zabaniotou, 2014), was proved to be suitable. The dual system  
458 could also effectively reduce any production of secondary wastes. It was calculated that approximately  
459 1 t of final residues (char) are obtained from around 6.9 t of solid digestate treated, thus efficiently  
460 contributing to waste minimization.

#### 461 *4. Discussion and recommendations for policy making and businesses development*

462 The study is relevant not only from an academic but also from a policy making and business  
463 perspective, particularly in the EU context. The European Commission took several important  
464 initiatives in resource efficiency during the years 2011–2015, culminating with the Circular  
465 Economy Package in December 2015. Useful managerial suggestions for policy makers and  
466 businesses about the implementation of innovative options for agricultural waste management can be  
467 drawn, related to gains in energy recovery with the efficient reduction of wastes, elimination of  
468 pathogens in digestate and production of a carbonaceous material for soil preservation and long-  
469 term carbon sequestration.

470 The coupled AD/digestate gasification (DG) system proposed in this study shows real promises for  
471 future implementation, offering a continual circulation of resources in the long-term rather than  
472 offering a temporary solution to the problem of agricultural waste. It provides a source of renewable  
473 energy in the form of CHP besides biogas and a soil amendment made from the digestate, which is  
474 rich in organic matter and nutrients.

475 The system can be stand alone or incorporated in a Circular Economy territorial model. The stand-  
476 alone dual system offers a management option for the agricultural wastes, mitigating waste-to-landfill,  
477 enhancement options for the digestate, and the energy required for the AD plant. It offers a resource  
478 closed loop by converting digestate generated by biogas plant into commodity consumed by the plant  
479 (energy) and a product to be used in agriculture (char), as biofertilizer, complying with the Circular  
480 Economy concept, where waste resulting from a process, AD in the present case, can be used as  
481 primary inputs for other process, gasification in the present case, towards achieving environmental,  
482 economic, and social advantages (Fig. 5A). In a scheme that involves separate industries in a  
483 collective approach to competitive advantage involving exchange of materials, energy and services,  
484 many AD plants could be coupled with one central gasification plant for a Circular Economy  
485 territorial model. In the case of regionally centralised large-scale gasification unit, digestate from  
486 various AD plant can be collected, transported, and be used as feedstocks in the large-scale centralised  
487 gasification-based CHP plant (Fig. 5B).

488 The amount of electric energy and biochar produced in the system depend on the amount of digestate  
489 and agricultural waste production rates. Therefore, the efficiency of the application of the approach

490 may be different depending on specific area and case considered: stand-alone dual system (Fig. 5A) or  
491 multiple biogas plants in cascade conjunction with a central gasification unit (Fig. 5B). Three main  
492 environmental benefits can be obtained:

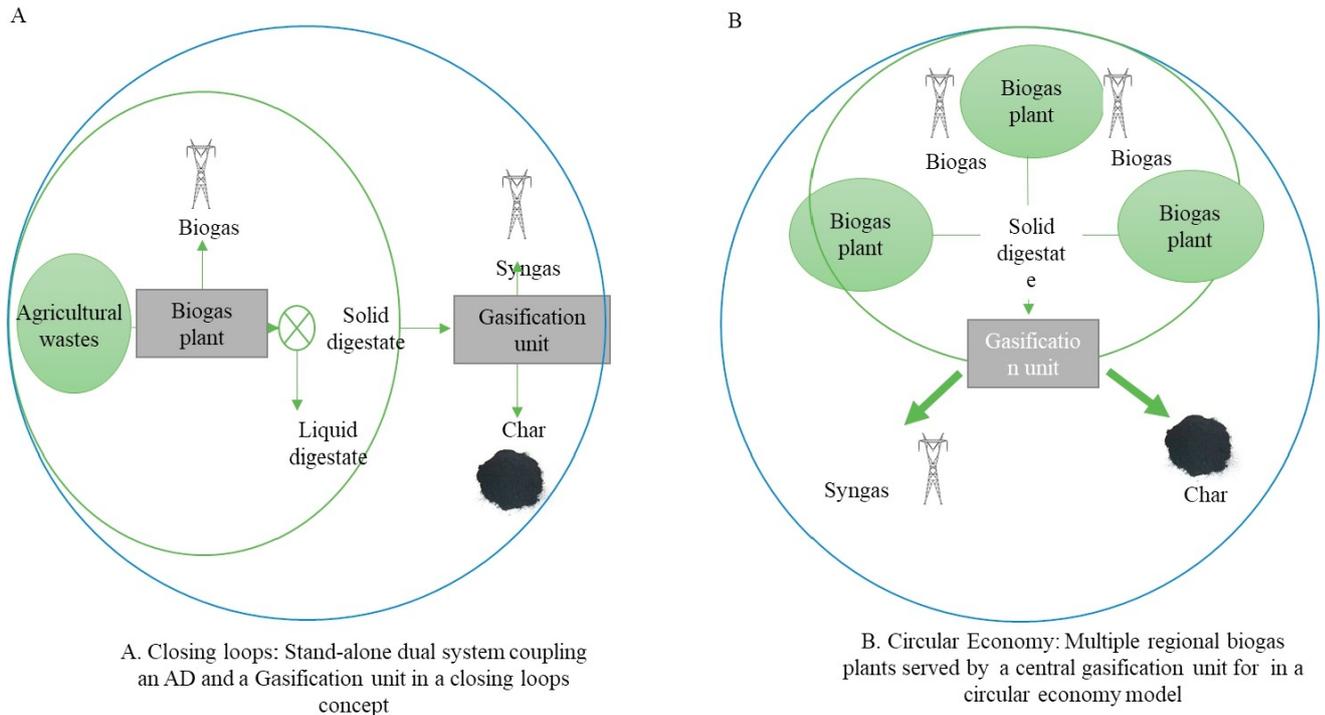
- 493 i) Sustainable management of digestate.
- 494 ii) Less amount of energy produced using traditional processes.
- 495 iii) Less GHG emissions in atmosphere, due to lower energy production using conventional  
496 processes.

497 In addition, economic advantages can be expected in terms of cost reduction about waste disposal and  
498 energy procurement from outside, even if pre-feasibility study is required.

499 The application of the above system in the ‘Circular Economy’ across the agri-food sector, will reduce  
500 wastes while also making best use of the ‘wastes’ produced by using economically viable processes  
501 and procedures to increase their value. It presents a major opportunity for the development of a  
502 Circular Economy (CE) using innovative conjunction of applied technologies and profitable business  
503 practices to address the utilization of agricultural wastes, byproducts and co-products.

504 The proposed system could be a smart sustainable rural energy infrastructure, in rural farming  
505 communities, suited to the Circular Economy principle. Dual AD and gasification system with farms  
506 supplying crop and slurry feedstock and local industry supplying food waste, can generate electricity  
507 and fertilizer which can then be used by the local community (Blades et al., 2017), rather focusing on  
508 the social benefits that a transformation from a linear to a Circular Economy would entail.

509



510  
511

512 **Figure 5.** Conceptual model of: **A)** a standalone dual system of AD with gasification for electricity  
513 and biochar recovery; **B)** a symbiotic scheme of various biogas plants with a central gasification unit  
514 for electricity and biochar recovery.

515

## 516 5. Conclusions

517

518 This study investigated the coupling of a typical anaerobic digestion (AD) with a gasification  
519 system in a cascade flow of materials. It showed the dual system's potentiality to increase the  
520 renewable energy efficiency and to produce a carbonaceous material for agronomy purposes, towards  
521 the transition to an inclusive and Circular Economy. Although the proposed process system might be  
522 used in the conventional practice, the closed loop components constitute the differences from the  
523 linear model to a circular model of waste management. In this Circular Economy concept, higher  
524 energy recovery could be achieved. The coupled AD/digestate gasification system proposed in this  
525 study shows real promises for future implementation, offering a continual circulation of resources in  
526 the long-term rather than offering a temporary solution to the problem of agricultural waste.

527 The dual system of AD with air gasification of the digestate produced by a commercial biogas plant  
528 was studied.

529 The optimisation of the digestate gasification was achieved at 850 °C and  $\lambda = 0.24$ , resulting in a  
530 medium heating value gas fuel with LHV of 2.88 MJ Nm<sup>-3</sup> and H<sub>2</sub>/CO = 2.3, classified as medium  
531 heating value fuel, suitable for CHP production.

532 From an energetic and economic point of view, results have shown that the heat excess produced from  
533 the anaerobic digestion plant is sufficient for a complete drying of the solid digestate. Furthermore  
534 after gazification process, it can be generated a surplus of electricity of about 971 kWh<sub>el</sub> day<sup>-1</sup>,  
535 enhancing thus the economic viability of the AD plant.

536 In parallel, the obtained carbonaceous material exhibited pH of 11.4, TS of 89.5%, over 50% more C  
537 content compared to the precursor material, and a significant increase of P (17,400 mg kg<sup>-1</sup>TS), K  
538 (48,132 mg kg<sup>-1</sup>TS), Ca (69,139 mg kg<sup>-1</sup>TS) and Mg (19,446 mg kg<sup>-1</sup>TS) nutrients content.

539 From environmental point of view, three main environmental benefits can be obtained: i) sustainable  
540 management of digestate; ii) less amount of energy produced using traditional processes, and iii)  
541 less GHG emissions in atmosphere, due to lower energy production using conventional processes iv)  
542 by generating a carbonaceous material rich in nutrients and recalcitrant carbon contributing to carbon  
543 sequestration in soil. In addition, economic advantages are produced in terms of cost reduction about  
544 waste disposal, fertilizers requirement and energy procurement from the grid.

545 Further research is needed on char's implication on soil amendment applications, mainly on  
546 the bioavailability of the main macronutrients present in gasification char, along with growth plants  
547 tests, to assess their use in agronomic application.

548

549 **Acknowledgements.** Acknowledgements are attributed by the APESA members to the Nouvelle  
550 Aquitaine Region for its financial support through the FEDER program.

551

552

## 553 **References**

554 Allesina, G., Pedrazzi, S., Guidetti, L., Tartarini, P., 2015. Modeling of coupling gasification and  
555 anaerobic digestion processes for maize bioenergy conversion. *Biomass Bioenergy* 81, 444-451.

556 APHA, 2005. *Standard Methods for the Examination of Water and Wastewater*, 21th ed. American  
557 Public Health Association, Washington DC, USA.

558 Barakat, A., Chuetor, S., Monlau, F., Solhy, A., Rouau, X., 2014. Eco-friendly dry chemo-mechanical  
559 pretreatments of lignocellulosic biomass: Impact on energy and yield of the enzymatic  
560 hydrolysis. *Applied Energy* 113, 97-105.

561 Basu, P., 2010. *Biomass Gasification and Pyrolysis: Practical Design and Theory*. Elsevier Science.

562 Cao, Y., Wang, Y., Riley, J.T., Pan, W.-P., 2006. A novel biomass air gasification process for  
563 producing tar-free higher heating value fuel gas. *Fuel Process. Technol.* 87, 343-353.

564 Channiwala, S.A., Parikh, P.P., 2002. A unified correlation for estimating HHV of solid, liquid and  
565 gaseous fuels. *Fuel* 81, 1051-1063.

566 Chen, W.-H., Chen, C.-J., Hung, C.-I., Shen, C.-H., Hsu, H.-W., 2013. A comparison of gasification  
567 phenomena among raw biomass, torrefied biomass and coal in an entrained-flow reactor.  
568 Applied Energy 112, 421-430.

569 Chertow, M.R., Industrial symbiosis: literature and taxonomy, 2000. Annual review of energy and the  
570 environment 25, 313-337

571 Devi, L., Ptasiński, K.J., Janssen, F.J.J.G., 2003. A review of the primary measures for tar elimination  
572 in biomass gasification processes. Biomass Bioenergy 24, 125-140.

573 EBC, 2012. 'European Biochar Certificate - Guidelines for a Sustainable Production of Biochar.'  
574 European Biochar Foundation (EBC), Arbaz, Switzerland.  
575 <http://www.europeanbiochar.org/en/download>. Version 6.3E of 14th August 2017, accessed:  
576 14/09/2017.

577 Fabbri, D., Torri, C., 2016. Linking pyrolysis and anaerobic digestion (Py-AD) for the conversion of  
578 lignocellulosic biomass. Curr. Opin. Biotechnol. 38, 167-173.

579 Fryda, L., Panopoulos, K.D., Kakaras, E., 2008. Integrated CHP with autothermal biomass gasification  
580 and SOFC-MGT. Energy Convers. Manage. 49, 281-290.

581 Hansen, V., Müller-Stöver, D., Ahrenfeldt, J., Holm, J.K., Henriksen, U.B., Hauggaard-Nielsen, H.,  
582 2015. Gasification biochar as a valuable by-product for carbon sequestration and soil  
583 amendment. Biomass Bioenergy 72, 300-308.

584 Harvey, O.R., Kuo, L.-J., Zimmerman, A.R., Louchouart, P., Amonette, J.E., Herbert, B.E., 2012. An  
585 Index-Based Approach to Assessing Recalcitrance and Soil Carbon Sequestration Potential of  
586 Engineered Black Carbons (Biochars). Environ. Sci. Technol. 46, 1415-1421.

587 IBI, 2014. International Biochar Initiative (IBI). [http://www.biochar-](http://www.biochar-international.org/sites/default/files/IBI_Biochar_Standards_V2%20final_2014.pdf)  
588 [international.org/sites/default/files/IBI\\_Biochar\\_Standards\\_V2%20final\\_2014.pdf](http://www.biochar-international.org/sites/default/files/IBI_Biochar_Standards_V2%20final_2014.pdf), accessed:  
589 14/11/2014.

590 Kataki, S., Hazarika, S., Baruah, D.C., 2017. Assessment of by-products of bioenergy systems  
591 (anaerobic digestion and gasification) as potential crop nutrient. Waste Manage. (Oxford) 59,  
592 102-117.

593 Kookana, R.S., Sarmah, A.K., Van Zwieten, L., Krull, E., Singh, B., 2011. Chapter three - Biochar  
594 Application to Soil: Agronomic and Environmental Benefits and Unintended Consequences, in:  
595 Donald, L.S. (Ed.), Advances in Agronomy. Academic Press, pp. 103-143.

596 Kratzeisen, M., Starcevic, N., Martinov, M., Maurer, C., Müller, J., 2010. Applicability of biogas  
597 digestate as solid fuel. Fuel 89, 2544-2548.

598 Kuligowski, K., Luostarinen, S., 2011. Thermal Gasification of Manure. Baltic Forum for Innovative  
599 Technologies for Sustainable Manure Management.

600 Lacroix, N., Rousse, D.R., Hausler, R., 2014. Anaerobic digestion and gasification coupling for  
601 wastewater sludge treatment and recovery. Waste Manage. Res. 32, 608-613.

602 Li, H., Larsson, E., Thorin, E., Dahlquist, E., Yu, X., 2015. Feasibility study on combining anaerobic  
603 digestion and biomass gasification to increase the production of biomethane. *Energy Convers.*  
604 *Manage.* 100, 212-219.

605 Li, X.T., Grace, J.R., Lim, C.J., Watkinson, A.P., Chen, H.P., Kim, J.R., 2004. Biomass gasification in  
606 a circulating fluidized bed. *Biomass Bioenergy* 26, 171-193.

607 Li, Y., Zhang, R., He, Y., Zhang, C., Liu, X., Chen, C., Liu, G., 2014. Anaerobic co-digestion of  
608 chicken manure and corn stover in batch and continuously stirred tank reactor (CSTR).  
609 *Bioresour. Technol.* 156, 342-347.

610 Lv, P.M., Xiong, Z.H., Chang, J., Wu, C.Z., Chen, Y., Zhu, J.X., 2004. An experimental study on  
611 biomass air-steam gasification in a fluidized bed. *Bioresour. Technol.* 95, 95-101.

612 Manara, P., Zabaniotou, A., 2013. Co-pyrolysis of biodiesel-derived glycerol with Greek lignite: A  
613 laboratory study. *J. Anal. Appl. Pyrolysis* 100, 166-172.

614 Manara, P., Zabaniotou, A., 2014. Indicator-based economic, environmental, and social sustainability  
615 assessment of a small gasification bioenergy system fuelled with food processing residues from  
616 the Mediterranean agro-industrial sector. *Sustainable Energy Technologies and Assessments* 8,  
617 159-171.

618 Monlau, F., Barakat, A., Trably, E., Dumas, C., Steyer, J.-P., Carrère, H., 2013a. Lignocellulosic  
619 Materials Into Biohydrogen and Biomethane: Impact of Structural Features and Pretreatment.  
620 *Crit. Rev. Environ. Sci. Technol.* 43, 260-322.

621 Monlau, F., Latrille, E., Da Costa, A.C., Steyer, J.-P., Carrère, H., 2013b. Enhancement of methane  
622 production from sunflower oil cakes by dilute acid pretreatment. *Applied Energy* 102, 1105-  
623 1113.

624 Monlau, F., Sambusiti, C., Ficara, E., Aboulkas, A., Barakat, A., Carrere, H., 2015a. New  
625 opportunities for agricultural digestate valorization: current situation and perspectives. *Energy*  
626 *& Environmental Science* 8, 2600-2621.

627 Monlau, F., Sambusiti, C., Antoniou, N., Barakat, A., Zabaniotou, A., 2015b. A new concept for  
628 enhancing energy recovery from agricultural residues by coupling anaerobic digestion and  
629 pyrolysis process. *Applied Energy* 148, 32-38.

630 Monlau, F., Sambusiti, C., Antoniou, N., Zabaniotou, A., Solhy, A., Barakat, A., 2015c. Pyrochars  
631 from bioenergy residue as novel bio-adsorbents for lignocellulosic hydrolysate detoxification.  
632 *Bioresour. Technol.* 187, 379-386.

633 Monlau, F., Francavilla, M., Sambusiti, C., Antoniou, N., Solhy, A., Libutti, A., Zabaniotou, A.,  
634 Barakat, A., Monteleone, M., 2016. Toward a functional integration of anaerobic digestion and  
635 pyrolysis for a sustainable resource management. Comparison between solid-digestate and its  
636 derived pyrochar as soil amendment. *Applied Energy* 169, 652-662.

637 Nipattummakul, N., Ahmed, I., Kerdsuwan, S., Gupta, A.K., 2010. High temperature steam  
638 gasification of wastewater sludge. *Applied Energy* 87, 3729-3734.

639 Opatokun, S.A., Yousef, L.F., Strezov, V., 2017. Agronomic assessment of pyrolysed food waste  
640 digestate for sandy soil management. *J. Environ. Manage.* 187, 24-30.

641 Panopoulos, K.D., Fryda, L.E., Karl, J., Poulou, S., Kakaras, E., 2006. High temperature solid oxide  
642 fuel cell integrated with novel allothermal biomass gasification: Part I: Modelling and feasibility  
643 study. *J. Power Sources* 159, 570-585.

644 Pedrazzi, S., Allesina, G., Belló, T., Rinaldini, C.A., Tartarini, P., 2015. Digestate as bio-fuel in  
645 domestic furnaces. *Fuel Process. Technol.* 130, 172-178.

646 Sambusiti, C., Monlau, F., Ficara, E., Carrère, H., Malpei, F., 2013. A comparison of different pre-  
647 treatments to increase methane production from two agricultural substrates. *Applied Energy*  
648 104, 62-70.

649 Santi, G., Proietti, S., Moscatello, S., Stefanoni, W., Battistelli, A., 2015. Anaerobic digestion of corn  
650 silage on a commercial scale: Differential utilization of its chemical constituents and  
651 characterization of the solid digestate. *Biomass Bioenergy* 83, 17-22.

652 Saveyn, H., Edder, P., 2014. End-of-waste criteria for biodegradable waste subjected to biological  
653 treatment (compost & digestate): Technical proposal. IPTS, EC, Seville, Spain.

654 Sheets, J.P., Yang, L., Ge, X., Wang, Z., Li, Y., 2015. Beyond land application: Emerging  
655 technologies for the treatment and reuse of anaerobically digested agricultural and food waste.  
656 *Waste Manage. (Oxford)* 44, 94-115.

657 Skoulou, V., Zabaniotou, A., Stavropoulos, G., Sakelaropoulos, G., 2008. Syngas production from  
658 olive tree cuttings and olive kernels in a downdraft fixed-bed gasifier. *Int. J. Hydrogen Energy*  
659 33, 1185-1194.

660 Srinivasan, P., Sarmah, A.K., Smernik, R., Das, O., Farid, M., Gao, W., 2015. A feasibility study of  
661 agricultural and sewage biomass as biochar, bioenergy and biocomposite feedstock: Production,  
662 characterization and potential applications. *Sci. Total Environ.* 512–513, 495-505.

663 Tambone, F., Scaglia, B., D'Imporzano, G., Schievano, A., Orzi, V., Salati, S., Adani, F., 2010.  
664 Assessing amendment and fertilizing properties of digestates from anaerobic digestion through a  
665 comparative study with digested sludge and compost. *Chemosphere* 81, 577-583.

666 Troy, S.M., Nolan, T., Leahy, J.J., Lawlor, P.G., Healy, M.G., Kwapinski, W., 2013. Effect of sawdust  
667 addition and composting of feedstock on renewable energy and biochar production from  
668 pyrolysis of anaerobically digested pig manure. *Biomass Bioenergy* 49, 1-9.

669 Wender, I., 1996. Reactions of synthesis gas. *Fuel Process. Technol.* 48, 189-297.

670 WRAP, 2012. Enhancement and treatment of digestates from anaerobic digestion.

671 Yao, Z., Li, W., Kan, X., Dai, Y., Tong, Y.W., Wang, C.-H., 2017. Anaerobic digestion and  
672 gasification hybrid system for potential energy recovery from yard waste and woody biomass.  
673 *Energy* 124, 133-145.

674 Zabaniotou, A., 2014. Agro-residues implication in decentralized CHP production through a  
675 thermochemical conversion system with SOFC. *Sustainable Energy Technologies and*  
676 *Assessments* 6, 34-50.

677 Zabaniotou, A., Bitou, P., Kanellis, T., Manara, P., Stavropoulos, G., 2014. Investigating Cynara C.  
678 biomass gasification producer gas suitability for CHP, second generation biofuels, and H2  
679 production. *Ind. Crop Prod.* 61, 308-316.

680 Zabaniotou, A., Rovas, D., Libutti, A., Monteleone, M., 2015. Boosting circular economy and closing  
681 the loop in agriculture: Case study of a small-scale pyrolysis–biochar based system integrated in  
682 an olive farm in symbiosis with an olive mill. *Environmental Development* 14, 22-36.

683

684

685

686

687

688

689

690