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Specific and total *N*-nitrosamines formation potentials of nitrogenous micropollutants during chloramination

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Abstract

N-nitrosamines are a group of potent human carcinogens that can be formed during oxidative treatment of drinking water and wastewater. Many tertiary and quaternary amines present in consumer products (e.g., pharmaceuticals, personal care and household products) are known to be *N*-nitrosodimethylamine (NDMA) precursors during chloramination, but the formation of other *N*-nitrosamines has been rarely studied. This study investigates the specific and total *N*-nitrosamine (TONO) formation potential (FP) of various precursors from nitrogen-containing micropollutants (chlorhexidine, metformin, benzalkonium chloride and cetyltrimethylammonium chloride) and tertiary and quaternary model amines (trimethyl amine, *N,N*-dimethylbutyl amine, *N,N*-dimethylbenzyl amine and tetramethyl ammonium). All the studied nitrogenous micropollutants displayed quantifiable TONO FP, with molar yields in the range 0.04-11.92%. However, the observed TONO pools constituted mostly of uncharacterized species, not included in US-EPA 8270 *N*-nitrosamines standard mix. Only the quaternary ammonium compound benzalkonium chloride showed quantifiable NDMA FP (0.56% molar yield), however, explaining only a minor fraction of the observed TONO FP. The studied model amines showed molar NDMA yields from 0.10% (trimethyl amine) to 5.05% (*N,N*-dimethylbenzyl amine), very similar to the molar TONO yields. The comparison of the FPs of micropollutants and model compounds showed that the presence of electron donating functional groups (such as a benzyl group) in tertiary and quaternary amine precursors leads to a higher formation of NDMA and uncharacterized *N*-nitrosamines, respectively. LC-qTOF screening of a list of proposed *N*-nitrosamine structures has enabled to identify a novel *N*-nitrosamine (*N*-nitroso-*N*-methyl dodecylamine) from the chloramination of benzalkonium chloride. This finding supports the hypothesis that different functional groups in quaternary amines can act as leaving groups during chloramination and form differing *N*-nitrosamine structures at significant yield.

36 Molar TONO yields determined for micropollutants were finally validated under experimental
37 conditions closer to real water matrices, confirming their representativeness also for lower
38 concentration ranges.

39

40 **Keywords:** total *N*-nitrosamines (TONO), NDMA, nitrogenous micropollutants, *N*-nitrosamines
41 formation potential, chloramination

42

43 1. Introduction

44 *N*-nitrosamines are a group of emerging disinfection by-products that can be formed during
45 chloramination, chlorination and ozonation of drinking water and wastewater (Krasner et al., 2013;
46 Mitch and Sedlak, 2004; Richardson and Ternes, 2014). Several *N*-nitrosamines are classified as
47 probable human carcinogens and mutagens, with low nanogram-per-litre drinking water
48 concentrations associated with a 10^{-6} lifetime excess cancer risk (IRIS US EPA, 2018; Mitch et al.,
49 2003). US-EPA listed five *N*-nitrosamines on the Contaminant Candidate List 4 (CCL4) as contaminants
50 potentially occurring in public water systems (US EPA CCL4, 2016).

51 Previous studies (Kemper et al., 2010; Le Roux et al., 2011; Mitch and Sedlak, 2004; Zeng and Mitch,
52 2015) showed that various consumer products, such as pharmaceuticals, personal care and household
53 products, play an important role in the formation of *N*-nitrosamines. These classes of compounds are
54 typically present in municipal wastewater effluents, wherefrom they are discharged into the aquatic
55 environment and may reach water resources used for drinking water production (Schwarzenbach et
56 al., 2006). Zeng and Mitch (2015) found that domestic greywater and blackwater streams are
57 significant sources of chloramine-reactive and ozone-reactive *N*-nitrosamine precursors. Even though
58 specific precursors have not been characterized, *N*-nitrosamine formation was associated with the
59 presence of shampoos, handsoaps, dishsoaps and pharmaceuticals. Specific *N*-nitrosamine precursors
60 may include tertiary or quaternary amines, being common macro-constituents of many consumer
61 products (Dai and Mitch, 2013; Shen and Andrews, 2011a).

62 Various tertiary amines commonly present in herbicides and in pharmaceutical and personal care
63 products were found to be *N*-nitrosodimethylamine (NDMA) precursors during chloramination,

64 showing molar conversion yields varying from <1% up to 60-90% for ranitidine (Le Roux et al., 2011;
65 Lee et al., 2007; Mitch and Schreiber, 2008; Mitch and Sedlak, 2004; Sacher et al., 2008; Schmidt and
66 Sacher, 2006; Selbes et al., 2013; Shen and Andrews, 2011a, 2011b; Spahr et al., 2016). Various
67 mechanisms have been proposed for NDMA formation during chloramination of tertiary amines (Choi
68 and Valentine, 2002a, 2002b, 2003; Shah et al., 2012; Shah and Mitch, 2011). NDMA formation was
69 initially attributed to the liberation of the dimethylamine (DMA) moiety via reactions of chlorine or
70 monochloramine followed by reactions between DMA and dichloramine (Krasner et al., 2013; Mitch
71 and Schreiber, 2008; Mitch and Sedlak, 2004). Further studies suggested that tertiary amines can also
72 form *N*-nitrosamines without proceeding through a secondary amine intermediate, but through
73 nucleophilic attack of the amine on chloramine (Selbes et al., 2013). Recent studies showed that the
74 stability and the electron distribution of the moieties bound to the DMA group play a key role in NDMA
75 formation and strongly affect its production yields (Shah and Mitch, 2011). In particular, the presence
76 of an electron-donating group close to the DMA moiety can increase the electron density on the
77 nitrogen atom, enhancing a nucleophilic attack on the chloramines (Le Roux et al., 2011; Selbes et al.,
78 2013; Shen and Andrews, 2011a).

79 Compared to tertiary amines, only few studies investigated the formation of *N*-nitrosamines from
80 specific quaternary ammonium compounds (QACs). Various quaternary ammonium cationic polymers
81 (e.g., polyDADMAC, polyamine, polyacrylamide) used for water coagulation and flocculation are
82 known to promote NDMA formation during chloramination, due to direct polymer nitrosation and
83 mainly the formation of intermediate polymer degradation products (Cornwell et al., 2015; Park et al.,
84 2009b, 2015; Zeng et al., 2016). Kemper et al. (2010) found that NDMA is formed during
85 chloramination of different QACs used as surfactants and disinfection agents, such as quaternary
86 alkylamines and benzyl trialkylamines, with relatively low molar yields (0.03-0.28%). So far, the
87 mechanism of the formation of *N*-nitrosamines from QACs remains unclear. The full substitution of the
88 nitrogen and the positive charge on the quaternary amine should hinder the reaction with chloramines
89 (Krasner et al., 2013). QACs can contain impurities that might constitute potential *N*-nitrosamines
90 precursors. However, Kemper et al. (2010) have shown that the formation of *N*-nitrosodibutylamine
91 (NDBA) by chloramination of benzyltributylammonium chloride only slightly decreased after

92 purification. This indicates that the QAC itself, rather than impurities, was the dominant *N*-nitrosamine
93 precursor. Even though the reaction pathways have not been defined, the same authors suggested that
94 the NDMA formation from QACs may involve amidogen or chloramino radicals formed from
95 chloramines. Some quaternary amine monomers (e.g., choline, cocoamidopropyl betaine) and
96 polymers (e.g., polyDADMAC, polyquaternium-7, polyacrylamide, polyamine and DMA-epi-DMA) were
97 also found to form NDMA during ozonation (Marti et al., 2015; Padhye et al., 2011), free chlorine and
98 chlorine dioxide disinfection (Park et al., 2009a; Zhang et al., 2014), even though at molar yields lower
99 than those observed for chloramination.

100 The majority of the studies in the literature deals with the formation of NDMA, being the most often
101 detected *N*-nitrosamine in tap and wastewater (Russell et al., 2012). However, previous studies
102 showed that NDMA is often a minor part of the total *N*-nitrosamines (TONO) pool (Dai and Mitch,
103 2013; Kulshrestha et al., 2010; Zeng and Mitch, 2015). Analysing 36 drinking water treatment plants
104 and distribution systems, Dai and Mitch (2013) found that NDMA made up only 5% of the TONO pool
105 on a median basis. Zeng and Mitch (2015) observed that the contribution of NDMA to TONO was
106 1-11% and 3-60% in domestic greywater streams treated by chloramine and ozone, respectively. The
107 same authors suggested that the precursors of the uncharacterized *N*-nitrosamine fraction were
108 associated to common macro-constituents of consumer products, such as QACs present in shampoos
109 and soaps.

110 While many NDMA precursors were identified in the literature, the determination of the TONO
111 formation potential (FP) of specific model compounds is not a common procedure. The investigations
112 of *N*-nitrosamine formation are typically limited to the species detected by the US-EPA method 521
113 (US EPA, 2005), while the identity of the other components of the TONO pool remains unclear.

114 The goal of this study was to investigate the formation of *N*-nitrosamines during chloramination of
115 various nitrogenous compounds that can be discharged to municipal wastewater and may end up in
116 the receiving water bodies. The investigated compounds included two quaternary ammonium
117 compounds used as disinfectants and surfactants (benzalkonium chloride, BZK, a mixture of benzyl
118 trialkylamines, and cetyltrimethylammonium chloride, CTMA, a quaternary alkylamine) and two

119 biguanides (chlorhexidine, CHD, an antiseptic agent used in disinfectants, pharmaceuticals and
120 cosmetics, and metformin, MET, a prescription drug used for treating diabetes).
121 Laboratory experiments on model solutions were performed to determine the formation potential of
122 total and specific *N*-nitrosamines for the studied precursors as a function of the pH. Besides analyses
123 of TONO and *N*-nitrosamines from the US-EPA 8270 standard mix (US EPA, 2005), high resolution
124 mass spectrometry analyses were carried out to identify the components of the unknown fraction of
125 TONO. The role of the molecular structure of the precursors on *N*-nitrosamine formation was further
126 investigated by model tertiary and quaternary amines (trimethyl amine, *N,N*-dimethylbutyl amine,
127 *N,N*-dimethylbenzyl amine and tetramethyl ammonium) containing functional groups similar to the
128 selected micropollutant moieties.

129

130 **2. Materials and methods**

131 **2.1 Chemicals and reagents**

132 All the experiments were conducted in ultrapure water (Milli-Q, Millipore) buffered with a phosphate
133 buffer (0.5 M) at different pH values (6, 7 or 8). Chlorhexidine dihydrochloride (CHD, >98%),
134 metformin hydrochloride (MET, 99%), benzalkonium chloride (BZK, >95%), cetyltrimethylammonium
135 chloride (CTMA, 25%wt in H₂O), tetramethylammonium chloride (TeMA, >99%), trimethylamine
136 (TMA, 98%), *N,N*-dimethylbenzyl amine (DMBzA, >99%) and *N,N*-dimethylbutyl amine (DMBA, 99%)
137 were purchased from Sigma-Aldrich. The molecular structures and main properties of the selected
138 compounds are provided in Table 1. Concentrated stock solutions of each precursor were prepared in
139 ultrapure water. The US-EPA 8270 *N*-nitrosamine standard mix, containing *N*-nitrosodimethylamine
140 (NDMA), *N*-nitrosodiethylamine (NDEA), *N*-nitrosomethylethylamine (NMEA), *N*-nitrosomorpholine
141 (NMOR), *N*-nitrosodibutylamine (NDBA), *N*-nitrosopiperidine (NPIP), *N*-nitrosopyrrolidine (NPYR),
142 *N*-nitrosodipropylamine (NDPA) and *N*-nitrosodiphenylamine (NDPhA) at the same mass
143 concentrations (2000 µg mL⁻¹ of each component in dichloromethane), was obtained from Sigma-
144 Aldrich. Monochloramine stock solutions were prepared daily as described by Le Roux et al. (2011)
145 using sodium hypochlorite (NaOCl, 15-20%, Sigma-Aldrich) and ammonium chloride (99.9%, VWR

146 Chemicals). The concentrations of monochloramine (NH_2Cl) and dichloramine (NHCl_2) (see section
147 2.3) in the stock solutions were measured after preparation; the NHCl_2 concentration was always not
148 quantifiable (see section 2.3). Ascorbic acid, HPLC grade acetonitrile, dichloromethane and methanol
149 were purchased from Sigma-Aldrich.

150

151 **2.2 Experimental procedures**

152 The self-decay of monochloramine at various initial concentrations (1, 5, 20 and 100 mM) and pH
153 values (6, 7 and 8) was investigated by measuring the NH_2Cl concentration over time (see section 2.3)
154 in buffered ultrapure water solutions. Chloramination experiments were conducted in amber
155 borosilicate bottles at room temperature (23°C) under two experimental setups (ES), denoted ES1 for
156 high concentrations of precursors and ES2 for more realistic lower concentrations of precursors.

157 Under ES1, reactions were conducted in 10 mL solutions, composed of variable volumes of precursor
158 stock solutions, sodium phosphate buffer (125 mM in the sample), pre-formed monochloramine stock
159 solution and ultrapure water. The initial concentration of the precursors was set to 0.1 mM for CHD
160 and 0.5 mM for all the other precursors, while the one of the buffer was set to have a buffer excess
161 with respect to the initial monochloramine concentration. *N*-nitrosamine FP is operationally defined
162 as the maximum *N*-nitrosamine concentration that can be formed from a precursor, using a NH_2Cl
163 concentration and a contact time much higher than applied during drinking water treatment (Krasner
164 et al., 2013). In this study, the experimental conditions to be adopted in the FP tests (i.e., reaction time,
165 initial concentrations of precursors and monochloramine) under ES1 were selected according to a
166 series of preliminary kinetic tests. The kinetics of *N*-nitrosamine formation from CHD, MET, BZK and
167 CTMA at pH 7 were investigated by testing different initial NH_2Cl concentrations (corresponding to
168 10-, 20-, 40-, 100- and 200-fold molar excess) and reaction times (1, 2, 5, 7 and 10 days) and
169 measuring TONO concentrations and the residual precursor concentrations (only for CHD and BZK)
170 (SI, Text S3). According to the results of the kinetic tests (discussed in section 3.1), FP tests at pH 6, 7
171 and 8 were conducted with all the precursors (micropollutants and model compounds) at a reaction
172 time of 7 days and a 200-fold molar excess of NH_2Cl relative to the precursor compound; only in the
173 case of BZK, FP tests were performed with a 10-fold NH_2Cl molar excess, assuring the highest

174 *N*-nitrosamine production (according to kinetic test results discussed in section 3.1). The
175 concentrations of TONO and *N*-nitrosamines of the US-EPA 8270 standard mix (see section 2.3) were
176 measured at the end of the FP tests.

177 In ES2, 7 days FP tests were conducted in 0.5 L solutions, containing deionized water, 1 mM sodium
178 phosphate buffer (pH 8), 100 μM and 0.5 μM initial concentrations of monochloramine and precursor,
179 respectively (corresponding to a 200-fold NH_2Cl molar excess). The TONO concentration was
180 measured after sample pre-concentration by solid phase extraction (see section 2.3).

181 The molar yield of *N*-nitrosamines formation was calculated as the ratio between the produced molar
182 concentration of the *N*-nitrosamine and the molar concentration of the precursor.

183 For all the chloramination experiments, reaction solutions were quenched at the end of the
184 pre-determined reaction time by dosing an ascorbic acid stock solution (280 mM) to get a
185 stoichiometric excess relative to the initial NH_2Cl concentration. Quenched solutions were stored at
186 4°C in the dark. All the experiments were performed in duplicate.

187

188 **2.3 Analytical methods**

189 NH_2Cl and NHCl_2 concentrations were determined spectrophotometrically (Shimadzu, UV-1800) using
190 their respective molar extinction coefficients at 245 and 295 nm, respectively and solving the
191 equations expressing absorbance values at the two wavelengths simultaneously (Schreiber and Mitch,
192 2005). The NHCl_2 concentration is quantifiable only when the ratio between absorbance values at
193 295 nm and 245 nm is $\geq 3\%$, corresponding to a $\text{NHCl}_2:\text{NH}_2\text{Cl}$ molar ratio $\leq 0.01\%$.

194 Total *N*-nitrosamine concentrations were measured by a method based on UV photolysis and
195 chemiluminescence measurement of nitric oxide (see Breider and von Gunten (2017) for more
196 details). The chemiluminescence signal was related to the TONO concentration by a standard curve,
197 obtained with different concentrations of the US-EPA 8270 *N*-nitrosamine standard mix (from 0.8 to
198 16.2 μM of each *N*-nitrosamine). The test solutions in absence of monochloramine were compared
199 with ultrapure water, to verify the absence of a background signal. The calibration was repeated daily
200 before each measurement session. The limit of detection (LOD) and quantification (LOQ) at the 95%

201 confidence level of the analytical method were calculated based on the standard error (SE) and the
202 slope (m) of the calibration curve ($LOD=3.3\times SE/m$, $LOQ=10\times SE/m$). Average (\pm standard deviation)
203 LOD and LOQ resulted equal to $0.3\pm 0.15\ \mu\text{M}$ and $1.0\pm 0.45\ \mu\text{M}$, respectively. For each sample, TONO
204 was measured in triplicate.

205 US-EPA 8270 standard mix *N*-nitrosamines were analysed after sample pre-concentration (the
206 procedure is described in Text S1, SI) by GC-MS/MS (Trace 1310 Series GC, Thermo) according to the
207 method described by Chen et al. (2012). LOD and LOQ for the investigated *N*-nitrosamines are
208 reported in Table S1, SI. The measurements were done in duplicate.

209 Samples obtained from FP tests in ES1 were analysed by liquid chromatography quadrupole time-of-
210 flight mass spectrometry (LC-qTOF). All the details about the analytical method and instruments are
211 reported in Text S2, SI. Data obtained by LC-qTOF analysis were elaborated to verify the presence of a
212 list of proposed compounds, according to the following procedure: the exact mass (m/z) of each
213 compound was extracted from the chromatogram. If a peak with intensity at least 3 orders of
214 magnitude higher than the background noise was observed, the mass spectrum in correspondence of
215 the peak retention time was further analysed. A list of possible molecular formulas was generated by
216 the ChemCalc software (Patiny and Borel, 2013) from the masses observed in the MS spectrum. If the
217 mass of a possible molecular formula corresponded to the theoretical one with an error <5 ppm, the
218 identification was considered positive.

219 Samples obtained in FP tests under ES2 were pre-concentrated by solid phase extraction (SPE)
220 following a modified US-EPA Method 521 protocol: 0.5 L samples were passed under vacuum through
221 pre-conditioned Superclean Coconut charcoal $2\text{g}/6\text{cm}^3$ cartridges (Supelco), subsequently dried with
222 ultrapure nitrogen gas flow for 15 min. Cartridges were eluted with 6 mL of HPLC grade acetonitrile.
223 Dichloromethane was not used as eluent because this solvent interferes with the signal measured with
224 the nitric oxide analyser. The solvent was evaporated at ambient temperature with a nitrogen flow to a
225 volume of about 2 mL. The recovery of TONO by SPE (η) was specifically assessed for the TONO pools
226 produced by CHD, BZK, CTMA and MET, according to the following procedure: firstly, a 7 days
227 chloramination test was conducted in ES1 (as described in section 2.2) and TONO concentration (C_1)

228 was measured; samples ($V_1=10$ mL) were then diluted by ultrapure water to 0.5 L (V_2) of a known
229 TONO concentration (C_2), calculated considering the dilution factor ($C_2=C_1\cdot V_1/V_2$). The sample was
230 then concentrated by SPE to about 2 mL (V_3) and the TONO concentration (C_3) was measured. The
231 TONO recovery was finally calculated ($\eta=C_3\cdot V_3/C_2/V_2$). The recovery efficiency of SPE was also
232 evaluated for a solution containing 0.16 μ M US-EPA 8270 *N*-nitrosamine standard mix, 1 mM
233 phosphate buffer (pH 8) and 2 mM ascorbic acid.

234

235 **3. Results and discussion**

236 **3.1 Total *N*-nitrosamine formation potential (FP)**

237 The results of kinetic tests are shown in Figure 1, reporting the TONO concentrations as a function of
238 the reaction time and the monochloramine dose for BZK, CTMA, CHD and MET at pH 7. For each NH_2Cl
239 dose, all the selected micropollutants showed a similar trend for the TONO concentrations, rapidly
240 increasing in the first 2-5 days and reaching a stable concentration after about 7 days. For CTMA and
241 MET, a higher NH_2Cl dose yielded a higher TONO formation (TONO concentrations from MET detected
242 only for 100- and 200-fold NH_2Cl molar excess); a similar trend was observed for CHD, for which
243 TONO production was comparable at the highest NH_2Cl doses (100- and 200-fold molar excess). An
244 opposite behaviour was observed for BZK, for which TONO concentrations obtained at 10- and 20-fold
245 molar excess were about 2-times higher than those obtained at 100- and 200-fold molar excess. This
246 opposing trend for BZK compared to the other precursors is unclear and further studies are needed to
247 elucidate this phenomenon. The monitoring of residual CHD and BZK concentrations during kinetic
248 tests (SI, Text S3 and Figure S2) showed that the precursors were rapidly consumed, with a >85%
249 abatement after 1 day of reaction. At the same time, the NH_2Cl concentration was expected to decrease
250 over time by disproportionation, as indicated by the monitoring of NH_2Cl in control solutions (SI, Text
251 S3 and Figure S1).

252 Based on the results of kinetic tests, a 200-fold NH_2Cl molar excess was adopted for FP tests with CHD,
253 MET and CTMA, leading to the maximum TONO formation, while a 10-fold NH_2Cl molar excess was

254 adopted for BZK. For all the precursors, a reaction time of 7 days was selected for FP tests, being
255 sufficient to reach the maximum TONO concentrations.

256 Samples obtained from FP tests at pH 6, 7 and 8 were firstly analysed for the specific *N*-nitrosamines
257 present in the US-EPA 8270 standard mix. Among the analysed samples, NDMA concentrations higher
258 than the LOQ (0.09 μM), corresponding to molar yields $>0.09\%$ for CHD and $>0.02\%$ for the other
259 precursors, were detected only for BZK at pH 8, while in all the other samples NDMA was not
260 detectable. None of the other *N*-nitrosamines of the US-EPA 8270 standard mix were detected after the
261 FP tests for any of the precursors.

262 The average NDMA concentration formed by BZK was $2.81\pm 0.13 \mu\text{M}$, corresponding to a molar
263 formation yield of $0.56\pm 0.03\%$ (Table 2). This value is consistent with previous observations (Kemper
264 et al., 2010) at pH 7 for the two main constituents of the benzalkonium chloride mixture:
265 benzyldimethyldodecylamine (0.28% molar NDMA yield) and benzyldimethyltetradecylamine (0.26%
266 molar NDMA yield).

267 No NDMA formation was observed for CHD, CTMA and MET (NDMA concentrations $<\text{LOD}$). While, to
268 our knowledge, NDMA formation from CHD has never been reported, the results obtained for CTMA
269 and MET were not unexpected, based on previous findings. In fact, relatively low molar NDMA yields
270 for CTMA (0.03%) and MET ($<0.04\%$) have been observed in previous studies (Kemper et al., 2010;
271 Shen and Andrews, 2011a). Hence, considering these low molar yields, NDMA concentrations close to
272 the LOQ would have been expected in the current study and were therefore not detected.

273 The determined TONO FP for CHD, MET, BZK and CTMA at various pH values are reported in Table 2,
274 expressed as molar TONO yields. Unlike the US-EPA 8270 standard mix for *N*-nitrosamines, TONO
275 measurements displayed detectable *N*-nitrosamine concentrations in all the samples. The TONO
276 formation increased with increasing pH for all precursor compounds. The pH-dependence of the
277 *N*-nitrosamine formation can be explained by (i) the increasing NH_2Cl stability at higher pH and (ii) the
278 acid-base speciation of some of the precursors. The NH_2Cl self-decomposition (SI, Figure S1) is slower
279 at higher pH values, resulting in a higher NH_2Cl exposure during the FP test, which in turn may lead to
280 a higher total *N*-nitrosamine production. For CHD and MET, the increase in TONO formation with

281 increasing pH can be explained by the acid-base speciation of these compounds (pK_a -values in Table
282 1). The fractions of the deprotonated forms of CHD and MET (as well as TMA, DMBA and DMBzA)
283 increase with increasing pH, which enhances the nucleophilic attack on NH_2Cl , resulting in significantly
284 higher *N*-nitrosamine yields (Selbes et al., 2013).

285 Among the studied precursor compounds, CHD exhibited the highest molar TONO yields at all pH
286 values, ranging from $2.64 \pm 1.45\%$ at pH 6 to $11.92 \pm 0.83\%$ at pH 8. Molar TONO yields from
287 $1.15 \pm 0.09\%$ (pH 6) to $4.93 \pm 1.42\%$ (pH 8) were obtained for CTMA, while BZK showed molar
288 conversions of between $0.68 \pm 0.24\%$ and $2.12 \pm 0.40\%$, depending on the pH. Among the studied
289 micropollutants, MET showed the lowest TONO yields ($\leq 0.21 \pm 0.03\%$), being at least one order of
290 magnitude lower than for the other compounds, for a given pH value.

291 These results demonstrate the relatively low NDMA formation yields of the studied micropollutants,
292 consistent with the results recently reported by Woods-Chabane et al. (2017) for tertiary and
293 quaternary amines in chloraminated wastewater samples. However, despite the low NDMA formation
294 yield, the studied precursors can produce significant TONO concentrations. The composition of the
295 formed TONO is unknown and is further discussed in section 3.3. Even though the overall mutagenicity
296 potential of the TONO pool cannot be easily evaluated, the molar TONO yields suggest that the
297 corresponding precursors should be taken into consideration during chloramination of wastewaters
298 and drinking waters. Being associated to consumer products, the studied precursors can easily enter
299 the aquatic system by wastewater discharge and get into the drinking water by this way. For example,
300 BZK homologues have been detected in wastewater effluents, river waters and groundwaters at
301 concentrations ranging from 10 ng L^{-1} to $65 \text{ } \mu\text{g L}^{-1}$ (Clara et al., 2007; Ding and Liao, 2001; Estévez et
302 al., 2012; Ferrer and Furlong, 2001; Martínez-Carballo et al., 2007). Relevant concentrations
303 ($\leq 0.66 \text{ } \mu\text{g L}^{-1}$) of alkyltrimethylammonium chloride compounds, including CTMA, were found in
304 Taiwanese rivers (Ding and Tsai, 2003), while concentrations in the range $10\text{-}150 \text{ ng L}^{-1}$ were detected
305 in Austrian surface waters (Martínez-Carballo et al., 2007). CHD concentrations of up to 5 ng L^{-1} have
306 been found in a drinking water reservoirs in Turkey (Yavuz et al., 2015), while concentrations
307 between 46.1 and 78.1 ng L^{-1} were detected in a Spanish DWTP influent treating river water (Boleda et

308 al., 2011). Scheurer et al. (2012) detected MET concentrations between 18 and 105 $\mu\text{g L}^{-1}$ in five
309 German WWTP influents and estimated a discharge in recipient rivers between 8 and 700 g d^{-1} , due to
310 their incomplete elimination during wastewater treatment. The same authors found MET
311 concentrations between 0.06 and 2.1 $\mu\text{g L}^{-1}$ in eight German rivers. Considering the highest
312 concentrations reported in the literature for surface waters or DWTP intakes and applying the
313 maximum TONO molar yields determined in this study (at pH 8), TONO FP of the selected precursors
314 could vary from 0.02 nM (1.2 ng L^{-1} expressed as NDMA) to 3.97 nM (293.5 ng L^{-1} expressed as NDMA),
315 for CHD and BZK, respectively. Although intake concentrations are strongly case specific, these
316 estimates indicate that the contribution of the studied precursors could be relevant. The determined
317 TONO FPs are consistent with those previously observed by Dai and Mitch (2013) in drinking water
318 systems, with measured TONO concentrations in the same order of magnitude (1 nM median
319 concentration). An additional important factor to be considered for evaluating the actual exposure to
320 *N*-nitrosamine precursors is their fate in aquatic systems, leading to a natural attenuation of their
321 concentrations at DWTP intakes. For instance, QACs are known to be easily adsorbed onto negatively
322 charged solids, such as soils and sludge, and to have a relatively low biodegradability, especially when
323 containing benzyl or alkyl functional groups (Garcia et al., 1999; Ikehata and El-Din, 2004; Li and
324 Brownawell, 2010; Ying, 2006). In contrast to a low *N*-nitrosamine molar yield, MET is characterized
325 by high mobility in natural aquatic systems and can be easily distributed in the aquatic environment,
326 due to the high persistence and low hydrophobicity ($\log K_{ow}=-2.64$) (Lindim et al., 2017; Mansour et
327 al., 2016; Trautwein et al., 2014). Even if the precursors are partially metabolized, they can contribute
328 to *N*-nitrosamine formation (Shen and Andrews, 2011a). For instance, typical biodegradation products
329 of QACs are tertiary and secondary amines, which can lead to *N*-nitrosamine formation yields much
330 higher than their parent compounds (Ying, 2006; Li and Brownawell, 2010).

331

332 **3.2 Total *N*-nitrosamine formation potential of model tertiary and quaternary amines**

333 The *N*-nitrosamine formation potential during chloramination was also determined for various
334 tertiary and one quaternary amine model compounds (Table 1): trimethyl amine (TMA),

335 *N,N*-dimethylbutyl amine (DMBA), *N,N*-dimethylbenzyl amine (DMBzA) and tetramethyl ammonium
336 (TeMA). Molar NDMA and TONO yields obtained in the pH range 6-8 are reported in Table 2. Similar to
337 the selected micropollutants, for these model compounds, only NDMA was found at detectable
338 concentrations from the US-EPA 8270 standard *N*-nitrosamines mix. Quantifiable NDMA and TONO
339 concentrations were found for all the studied model tertiary amines (TMA, DMBA and DMBzA). TMA
340 displayed TONO concentrations at pH 7 slightly higher than the LOQ, resulting in a molar conversion
341 yield of 0.03%. A NDMA concentration higher than the LOQ was determined only at pH 8, with an
342 average molar yield of 0.10%, while TONO concentrations corresponded to a molar yield of 0.29%. The
343 TONO FP for TMA at pH 8 was higher than the NDMA FP, which is probably due to the accuracy of the
344 two analytical methods and related mainly to the different sample pre-treatment procedures. In
345 particular, GC-MS/MS analysis could be affected by a significant loss of NDMA due to the sample
346 extraction procedure, which was not necessary for TONO analysis. A formation of non-NDMA
347 *N*-nitrosamines is not possible from TMA. Furthermore, the TONO method applied in this study
348 (Breider and von Gunten, 2017) tends to slightly overestimate NDMA concentrations (115%±5%
349 recovery of NDMA from standard curve of US-EPA 8270 *N*-nitrosamines standard mix), which may be
350 an additional explanation for the observed discrepancy.

351 NDMA was not detected in DMBA samples at pH 6 and 7, while TONO concentrations slightly above the
352 LOQ (corresponding to 0.02% molar yield) were measured at pH 7. NDMA and TONO molar yields of
353 0.05% and 0.17%, respectively, were obtained at pH 8. Similar to TMA, the NDMA FP observed for
354 DMBA at pH 8 was lower than the TONO FP and the difference in NDMA and TONO concentrations can
355 again be mainly attributed to the differences in the analytical methods.

356 Among the model tertiary amines, DMBzA displayed the highest *N*-nitrosamine formation potentials.
357 Molar NDMA yields were, on average, 1.22% at pH 7 and 5.05% at pH 8, while molar TONO yields were
358 0.72% or 3.27% at pH 7 or 8, respectively. The NDMA FPs resulted to be higher than TONO FPs in this
359 case. However, considering the uncertainties associated to the two analytical methods, measured
360 NDMA and TONO concentrations can be considered to be comparable, indicating that NDMA was the
361 most prevalent *N*-nitrosamine present in the analysed samples. The relatively high formation of NDMA
362 from DMBzA can be attributed to the low stability of the benzyl group, which is easily released from

363 the N-atom (Kemper et al., 2010; Selbes et al., 2013; Spahr et al., 2016). In fact, the benzyl group is a
364 weaker base compared to alkyl groups and plays a stronger electron donating effect on the DMA
365 moiety, resulting in higher NDMA formation rates (Shen and Andrews, 2011a).

366 Comparing the molar yields of NDMA and TONO for all the selected tertiary amines (MET, TMA, DMBA
367 and DMBzA), the role of the molecular structures of the precursors, in particular the type of the
368 functional group bound to the DMA moiety, becomes evident. Even though it was not possible to
369 evaluate the TONO composition, experimental results confirmed that the benzyl functional group is
370 much more easily released than the methyl, butyl or biguanide moieties, leading to significantly higher
371 NDMA yields.

372 For TeMA, the only tested quaternary amine model compound, NDMA concentrations higher than the
373 LOD were never detected, while quantifiable TONO FPs were found at all the pHs. TONO
374 concentrations at pH 6 and 7 were slightly higher than LOQ, while quantifiable concentrations were
375 measured at pH 8, corresponding to a 0.90% molar TONO yield. Since NDMA is expected to be the
376 most prevalent *N*-nitrosamine formed from TeMA, the significant difference between NDMA and TONO
377 yields observed at pH 8 could be explained by an underestimation of NDMA concentration due to the
378 sample extraction procedure. Even though the TONO composition observed for TeMA was not
379 characterized, the results obtained for other precursors show a link between the molecular structures
380 of QACs and *N*-nitrosamine formation. The TONO FP for TeMA was much lower than for the QACs BZK
381 and CTMA, with a significant difference in molar yields (difference >60%, considering the average
382 molar yields reported in Table 2 for pH 8). This indicates that the functional groups bound to the N-
383 atom of QACs play an important role for the *N*-nitrosamine conversion efficiency.

384 A previous study showed that formation of specific *N*-nitrosamines (i.e. NDMA and NDBA) from
385 chloramination of QACs is mostly associated to QACs themselves rather than lower order amine
386 impurities (Kemper et al., 2010). However, the negligible contribution of tertiary amine impurities has
387 not been demonstrated for the TONO FP. The selected tertiary amine model compounds in the current
388 study contain the same, or very similar, functional groups as potential impurities in the QACs.
389 Although the aliphatic chain is shorter in DMBA compared to BZK and CTMA, previous studies have
390 shown that the length of the alkyl group next to the nitrogen atom does not significantly affect the

391 *N*-nitrosamine conversion yield (Selbes et al., 2013). Results of FP tests showed that molar
392 *N*-nitrosamine yields of DMBA and TMA were never high enough to explain *N*-nitrosamine formation
393 from QACs by the presence of impurities (maximum estimated contribution to NDMA or TONO FPs of
394 BZK and CTMA <1%). DMBzA, for which the highest NDMA FP was observed, could contribute up to
395 the 45% of NDMA formation potential of BZK, if considering a maximum content of 5% as impurity
396 (BZK purity >95%). In contrast, the maximum contribution of DMBzA impurities to the TONO FP of
397 BZK would be $\leq 8\%$, indicating that most of the uncharacterized *N*-nitrosamine production is due to
398 BZK itself. Hence, the results from this study indicate that the observed TONO FP from QACs cannot be
399 attributed to tertiary amine impurities and that QACs themselves are the dominant *N*-nitrosamine
400 precursors.

401

402 **3.3 Elucidation of the potential structures of unknown *N*-nitrosamines formed from** 403 **quaternary ammonium compounds**

404 TONO constituents formed from the selected consumer product ingredients were partially
405 characterized only for BZK, for which it was possible to determine the NDMA FP. The ratio between
406 the average observed molar concentrations of NDMA and TONO was 0.26. Even by taking the
407 differences in the two analytical methods into account, the difference between NDMA and TONO
408 concentrations is significant. This indicates that a large fraction of the total *N*-nitrosamine pool
409 consists of unknown compounds. So far, no known *N*-nitrosamine has been identified in the TONO
410 pools formed from CHD, MET and CTMA.

411 Because no compounds from the US-EPA 8270 standard *N*-nitrosamines mix were found, some other
412 *N*-nitrosamines must be responsible for the measured TONO. Possible *N*-nitrosamine structures
413 formed by CHD cannot be easily proposed due to the complex molecular structure of the precursor. In
414 contrast to most of the tertiary or quaternary amines studied in the literature, CHD does not contain a
415 dialkylamino group which can be related to the formation of the corresponding *N*-nitrosamine, such as
416 NDMA, NDEA or NDBA (Kemper et al., 2010; Le Roux et al., 2011; Selbes et al., 2013; Shen and
417 Andrews, 2011a; Spahr et al., 2016). MET contains similar biguanide moieties as CHD (Table 1),
418 however, it showed significantly lower *N*-nitrosamine FP. Comparing the molecular structures of the

419 two precursors, the higher *N*-nitrosamine formation yield of CHD can be attributed to the aromatic
420 moiety, which is bound to the end-membered N-atom of the biguanide group. In comparison, MET
421 contains the same biguanide structure, however, with end membered DMA moiety. The low NDMA FP
422 of MET has been explained by the electron-withdrawing effect of the biguanide group bound to DMA
423 leading to a lower nucleophilicity of the N-atom (Le Roux et al., 2011; Selbes et al., 2013; Shen and
424 Andrews, 2011a). In contrast, the chlorinated aromatic ring in the CHD structure could enhance the
425 nucleophilic attack on the chloramine. Nevertheless, further studies are needed to verify these
426 assumptions.

427 For QACs, the presence of only one N-atom limits the number of possible reactions with NH_2Cl leading
428 to potential *N*-nitrosamine products. It has been demonstrated that *N*-nitrosamine formation from
429 tertiary and quaternary amines involves the release of the functional groups bound to the nitrogen
430 atom in the precursor molecule (Selbes et al., 2013). Various studies have shown that the structure of
431 the leaving groups in precursors containing a DMA moiety strongly influences the formation of NDMA,
432 resulting in a large range of formation yields (Kemper et al., 2010; Le Roux et al., 2011; Selbes et al.,
433 2013; Shen and Andrews, 2011b). However, most of these studies focused on NDMA formation and
434 TONO measurements were rarely performed.

435 In the current study, the structures of other potential *N*-nitrosamines were hypothesized based on all
436 possible leaving groups present in the precursor molecule. The potential *N*-nitrosamines resulting
437 from chloramination of BZK and CTMA are shown in Table 3. For instance, the benzyl group and the
438 aliphatic chain ($\text{C}_n\text{H}_{2n+1}$) bound to the N-atom in BZK are supposed to act as leaving groups during
439 NDMA formation. Similarly, the release of one methyl group and the 16-C aliphatic chain ($\text{C}_{16}\text{H}_{33}$) may
440 lead to the formation of NDMA from CTMA. The other *N*-nitrosamine structures were proposed by
441 considering all the other potential leaving groups, including the methyl group, the benzyl group and
442 the long-chain alkyl group for BZK, the methyl groups and the 16-C aliphatic chain for CTMA.

443 To test the formation of the proposed *N*-nitrosamines, the samples obtained by the TONO FP tests at
444 pH 8 were analysed by high resolution LC-qTOF. The high accuracy and resolution of the obtained
445 mass spectra allow proposing a compound based on its exact mass, in absence of analytical reference
446 standards. The procedure described in section 2.3 was applied to screen the list of potential *N*-

447 nitrosamines from the chromatograms, by extracting their exact masses (m/z) (calculated by the
448 ChemCalc software and reported in Table 3, including the protonated molecules). The procedure was
449 also applied to detect potential denitrosation products of the investigated compounds.

450 Among the hypothesized *N*-nitrosamines, a positive match was obtained for BZK. The full
451 chromatogram of the BZK sample (treated with 10-fold molar excess of monochloramine at pH 8) is
452 reported in Figure S3, SI. The extraction of the exact mass of $C_{13}H_{28}N_2O$ ($[M+H]^+=229.2279$) allowed to
453 obtain a distinct peak (intensity of $8.56 \cdot 10^5$ arbitrary units) at a retention time of 7.34 min (Figure
454 2 a). The relative mass spectrum (Figure 2 b) showed a peak at $m/z=229.2271$, corresponding to a
455 relative abundance of 93% and matching with high accuracy (error<0.1%) with two possible
456 molecular formulas: (i) the hypothesized *N*-nitrosamine ($C_{13}H_{28}N_2O$, error=3.876 ppm, Table 3) and
457 (ii) $C_{11}H_{27}N_5$ (error=1.981 ppm). The formation of the latter compound is highly improbable
458 considering the investigated reaction system. The identified mass corresponded with a high level of
459 confidence to *N*-nitroso-*N*-methyldodecylamine (Figure 2 a), a potential product of the reaction of
460 chloramine with benzyldimethyldodecylamine, which is the most abundant constituent of the
461 benzalkonium chloride mixture (about 70% abundance, according to the supplier (Sigma-Aldrich)).

462 The detection of *N*-nitroso-*N*-methyldodecylamine was not unexpected and its formation at significant
463 yield is consistent with previous results reported in the literature (Kemper et al., 2010; Selbes et al.,
464 2013; Spahr et al., 2016). Because the benzyl group is an electron-donating group, which increases the
465 electron density on the N-atom, a nucleophilic attack on the chloramine is favoured. In addition,
466 benzyl-type functional groups are known to be better leaving groups than alkyl groups, both in tertiary
467 and quaternary amine precursors (Kemper et al., 2010; Selbes et al., 2013). Hence, during
468 chloramination of BZK homologues, *N*-nitroso-*N*-metylalkylamines are expected to have higher yields
469 than benzyl-containing *N*-nitrosamines. Consistently, it was found previously that the molar yield of
470 *N*-nitrosomethylbenzylamine from benzyldimethyltetradecylamine (the C-14 constituent of BZK) was
471 only 0.016%, about 16-times lower than the molar NDMA yield (Kemper et al., 2010).

472 High resolution LC-qTOF analysis of a CTMA sample did not lead to an identification of the
473 hypothesized compound ($C_{17}H_{36}N_2O$) with enough confidence (results are reported in Figures S3 and
474 S4, SI). The formation of a compound during FP tests cannot be excluded from the absence of a positive

475 match. It may be caused by artefacts in the detection method, e.g., the stability of the *N*-nitroso group
476 during ionization. Moreover, the absence of reference standards did not allow determining an optimal
477 sample extraction procedure and optimized analytical conditions for each proposed *N*-nitrosamine.
478 Thus, the performed analyses must be considered as a first attempt to characterize the unknown
479 fraction of TONO. Nevertheless, in the current study, in addition to NDMA, a novel *N*-nitrosamine from
480 the TONO FP pool of BZK was potentially identified. This finding strongly supports the hypothesis that
481 different leaving groups in the precursor molecule can lead to the formation of different
482 *N*-nitrosamines.

483 The novel *N*-nitrosamine, *N*-nitroso-*N*-methyldodecylamine was found to be carcinogenic in
484 experimental animal tests and its presence in personal care and household products has already been
485 cause of concern in previous studies (Hecht et al., 1982; Kamp and Eisenbrand, 1991; Lijinsky et al.,
486 1983; Morrison and Hecht, 1982; SCCS, 2012). For instance, a T25 of 0.46 mg kg⁻¹ d⁻¹ (being the
487 chronic dose rate giving tumours in the 25% of the tested animals (Dybing et al., 1997)) was
488 determined in laboratory rats for *N*-nitroso-*N*-methyldodecylamine, which is comparable to the values
489 obtained for other *N*-nitrosamines of the US-EPA 8270 standard mix (varying from 0.058 mg kg⁻¹ d⁻¹ of
490 NDMA to 0.57 mg kg⁻¹ d⁻¹ of NPYR) (SCCS, 2012). Hence, it may significantly contribute to the overall
491 cancer potency of the total *N*-nitrosamine pool. Even though the high-resolution LC-qTOF results were
492 not quantitative, they indicate that the uncharacterized fraction of TONO can be composed of species
493 as hazardous as the compounds from the US-EPA 8270 standard *N*-nitrosamine mix. Even if a full
494 characterization of the TONO pool is not always possible, coupling total and specific (i.e., US-EPA 8270
495 standard mix) *N*-nitrosamine analyses is an effective way to fully assess the potential hazardousness of
496 *N*-nitrosamine precursors. Product identification in the TONO FP pool by LC-qTOF may lead to further
497 insights into the formation of individual compounds, which might be of toxicological concern.

498

499 **3.4 Effect of precursors concentration on TONO formation potential**

500 Experimental conditions adopted in ES1 were suitable for determination of the *N*-nitrosamine FP of
501 the studied precursors and investigating the composition of the formed TONO compound pool.
502 However, the precursor and the monochloramine concentrations were much higher compared to

503 realistic systems, leading to the formation of extremely high *N*-nitrosamine concentrations. To validate
504 the molar TONO yield values obtained in ES1, FP tests were also conducted in ES2, with experimental
505 conditions closer to real water matrices (pH 8, 0.5 μM precursor and 100 μM monochloramine,
506 corresponding to 5.1 mg L^{-1} as NH_2Cl). The relatively long reaction time in FP tests (7 days) was mainly
507 selected to maximize *N*-nitrosamine formation and can be considered as an upper limit of residence
508 times in drinking water distribution systems. In contrast, it is longer than typical contact times for
509 wastewater disinfection. Due to the lower *N*-nitrosamine concentrations produced under ES2, TONO
510 concentrations were measured after sample pre-concentration by solid phase extraction (SPE).

511 As discussed above, the composition of the generated TONO compound pool is different for each
512 precursor. Since the physico-chemical properties of the formed *N*-nitrosamines can strongly affect the
513 efficiency of SPE (Dai and Mitch, 2013), the recovery of TONO concentrations by SPE was determined
514 for the specific *N*-nitrosamine pools formed by each precursor, as reported in section 2.2. The TONO
515 recovery efficiencies (values reported in Table S2, SI) resulted to be quite low for all the precursors,
516 with values of about 10%. Applying the obtained recovery efficiency values to the TONO
517 concentrations measured in the FP tests (ES2), the molar TONO yields resulted, on average as
518 $6.54\pm 0.39\%$, $4.05\pm 0.38\%$, $4.05\pm 0.02\%$ and $0.77\pm 0.32\%$ for CHD, BZK, CTMA and MET, respectively.
519 Despite the extremely different experimental conditions and the uncertainty associated to SPE, the
520 molar TONO yields obtained under ES2 were comparable with those obtained in ES1 at pH 8 (Table 2).
521 CHD and MET displayed the highest and the lowest molar TONO yields, respectively, while
522 intermediate values were obtained for QACs. Even though the molar TONO yields are slightly different,
523 they are in the same order of magnitude for ES1 and ES2. This indicates that results obtained in ES1
524 can be considered representative for more realistic conditions.

525 When dealing with TONO measurements in real water matrices, the recovery efficiency of SPE plays a
526 key role and it can strongly vary with the TONO pool composition (Dai et al., 2015; Zeng and Mitch,
527 2015). In this study, the SPE extraction efficiency was also evaluated for a 0.16 μM US-EPA 8270
528 standard *N*-nitrosamine mix solution, both in terms of TONO and specific *N*-nitrosamine constituents.
529 Some specific *N*-nitrosamines (such as NMOR, NDEA and NDBA) were not detected after SPE, denoting

530 a very low recovery. Nevertheless, the average recovery efficiency of the TONO concentrations was
531 42%. The variability in the recovery efficiencies obtained for samples from FP tests demonstrates that
532 the studied precursors lead to the formation of quite a large range of *N*-nitrosamines, having different
533 affinities to the SPE material and probably yielding higher analytical uncertainty. The relatively low
534 recovery of TONO concentration observed for the reacted samples could be due to the presence of
535 unreacted precursor molecules, having surface active properties and being present at high initial
536 concentrations, which could form micellar aggregates with the *N*-nitrosamines present in solution and
537 increase their hydrophilicity. Further studies are needed to elucidate the composition and the
538 properties of the observed TONO pools and to optimize the SPE procedure.

539

540 **4. Conclusions**

541 Specific and total *N*-nitrosamine (TONO) formation potentials (FP) for different micropollutants
542 (chlorhexidine, CHD, metformin, MET, benzalkonium chloride, BZK, and cetyltrimethylammonium
543 chloride, CTMA) and model compounds (trimethylamine, TMA, *N,N*-dimethylbutylamine, DMBA,
544 *N,N*-dimethylbenzylamine, DMBzA, and tetramethylammonium, TeMA) were determined at pH 6, 7
545 and 8. All the micropollutants displayed quantifiable molar TONO yields, with maximum values at pH 8
546 varying between 0.21% (MET) and 11.92% (CHD). A quantifiable NDMA FP was determined only for
547 BZK, with a molar yield of 0.56% (at pH 8) and corresponding to about 26% of the TONO FP. Generally,
548 the formed TONO compounds consisted mostly of uncharacterized species, not included in the US-EPA
549 8270 *N*-nitrosamines standard mix.

550 The study of tertiary and quaternary model amines revealed a role of precursor molecular structures,
551 in particular the nature of functional groups, in NDMA and TONO conversion yields: DMBzA displayed
552 the highest molar NDMA yields, indicating that the presence of a benzyl functional group in tertiary
553 amines containing a DMA moiety results in significantly higher NDMA formation. BZK and CTMA
554 showed molar TONO yields much higher than TeMA, suggesting that the formation of uncharacterized
555 *N*-nitrosamines from QACs is favoured by the presence of benzyl and long-chain alkyl functional
556 groups, instead of only methyl groups.

557 A screening procedure with high resolution LC-qTOF analysis allowed to identify with high probability
558 an uncharacterized *N*-nitrosamine (*N*-nitroso-*N*-methyldodecylamine) from the chloramination of
559 BZK. The detection of a second *N*-nitrosamine species in addition to NDMA points out that different
560 functional groups in QAC molecules can act as leaving groups during chloramination and form
561 different *N*-nitrosamine structures at significant yield. These results represent a first step in the
562 understanding of *N*-nitrosamine formation during chloramination of QACs.

563

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568

569 References

- 570 Agarwal, A., Nelson, T.B., Kierski, P.R., Schurr, M.J., Murphy, C.J., Czuprynski, C.J., McAnulty, J.F., Abbott,
571 N.L., 2012. Polymeric multilayers that localize the release of chlorhexidine from biologic
572 wound dressings. *Biomaterials* 33, 6783–6792.
- 573 Boleda, M.R., Galceran, M.T., Ventura, F., 2011. Behavior of pharmaceuticals and drugs of abuse in a
574 drinking water treatment plant (DWTP) using combined conventional and ultrafiltration and
575 reverse osmosis (UF/RO) treatments. *Environ. Pollut.* 159, 1584–1591.
576 <https://doi.org/10.1016/j.envpol.2011.02.051>
- 577 Breider, F., von Gunten, U., 2017. Quantification of Total N-Nitrosamine Concentrations in Aqueous
578 Samples via UV-Photolysis and Chemiluminescence Detection of Nitric Oxide. *Anal. Chem.* 89,
579 1574–1582. <https://doi.org/10.1021/acs.analchem.6b03595>
- 580 Chen, A., Huebschmann, H.J., Fangyan, L., Foong, C.Y., Harn, C.S., 2012. High Sensitivity Analysis of
581 Nitrosamines Using GC-MS/MS, Alpha Analytical Pte. Ltd., Thermo Fisher Scientific. Health Sci.
582 Auth. HAS Singap.
- 583 Choi, J., Valentine, R.L., 2003. N-nitrosodimethylamine formation by free-chlorine-enhanced
584 nitrosation of dimethylamine. *Environ. Sci. Technol.* 37, 4871–4876.
- 585 Choi, J., Valentine, R.L., 2002a. A kinetic model of N-nitrosodimethylamine (NDMA) formation during
586 water chlorination/chloramination. *Water Sci. Technol.* 46, 65–71.
- 587 Choi, J., Valentine, R.L., 2002b. Formation of N-nitrosodimethylamine (NDMA) from reaction of
588 monochloramine: a new disinfection by-product. *Water Res.* 36, 817–824.
- 589 Clara, M., Scharf, S., Scheffknecht, C., Gans, O., 2007. Occurrence of selected surfactants in untreated
590 and treated sewage. *Water Res.* 41, 4339–4348.
- 591 Cornwell, D., Krasner, S., Mitch, W., Pignatello, J., 2015. Investigating Coagulant Aid Alternatives to
592 PolyDADMAC Polymers. *Water Res. Found. Denver CO* 313.
- 593 Dai, N., Mitch, W.A., 2013. Relative importance of N-nitrosodimethylamine compared to total N-
594 nitrosamines in drinking waters. *Environ. Sci. Technol.* 47, 3648–3656.
- 595 Dai, N., Zeng, T., Mitch, W.A., 2015. Predicting N-nitrosamines: N-Nitrosodiethanolamine as a
596 significant component of total N-nitrosamines in recycled wastewater. *Environ. Sci. Technol.*
597 *Lett.* 2, 54–58.

- 598 Devi, S.R., 2013. Studies on Stability and Configurational Changes of Nickel and Copper Ethambutol
599 Dihydrochloride and Metformin Hydrochloride in both Cationic and Anionic Surfactants. *Int J*
600 *Enginee Sci Inven* 2.
- 601 Ding, W.-H., Liao, Y.-H., 2001. Determination of alkylbenzyltrimethylammonium chlorides in river
602 water and sewage effluent by solid-phase extraction and gas chromatography/mass
603 spectrometry. *Anal. Chem.* 73, 36–40.
- 604 Ding, W.-H., Tsai, P.-C., 2003. Determination of alkyltrimethylammonium chlorides in river water by
605 gas chromatography/ion trap mass spectrometry with electron impact and chemical
606 ionization. *Anal. Chem.* 75, 1792–1797.
- 607 Dybing, E., Sanner, T., Roelfzema, H., Kroese, D., Tennant, R.W., 1997. T25: a simplified carcinogenic
608 potency index: description of the system and study of correlations between carcinogenic
609 potency and species/site specificity and mutagenicity. *Basic Clin. Pharmacol. Toxicol.* 80, 272–
610 279.
- 611 Estévez, E., del Carmen Cabrera, M., Molina-Díaz, A., Robles-Molina, J., del Pino Palacios-Díaz, M., 2012.
612 Screening of emerging contaminants and priority substances (2008/105/EC) in reclaimed
613 water for irrigation and groundwater in a volcanic aquifer (Gran Canaria, Canary Islands,
614 Spain). *Sci. Total Environ.* 433, 538–546.
- 615 Ferrer, I., Furlong, E.T., 2001. Identification of alkyl dimethylbenzylammonium surfactants in water
616 samples by solid-phase extraction followed by ion trap LC/MS and LC/MS/MS. *Environ. Sci.*
617 *Technol.* 35, 2583–2588.
- 618 Garcia, M.T., Campos, E., Sanchez-Leal, J., Ribosa, I., 1999. Effect of the alkyl chain length on the
619 anaerobic biodegradability and toxicity of quaternary ammonium based surfactants.
620 *Chemosphere* 38, 3473–3483.
- 621 Hecht, S.S., Morrison, J.B., Wenninger, J.A., 1982. N-Nitroso-N-methyldodecylamine and N-nitroso-N-
622 methyltetradecylamine in hair-care products. *Food Chem. Toxicol.* 20, 165–169.
- 623 Ikehata, K., El-Din, M.G., 2004. Degradation of recalcitrant surfactants in wastewater by ozonation and
624 advanced oxidation processes: a review. *Ozone Sci. Eng.* 26, 327–343.
- 625 IRIS US EPA, 2018. Chemical Search | IRIS | US EPA [WWW Document]. URL
626 <https://cfpub.epa.gov/ncea/iris/search/index.cfm?keyword=n-nitrosodimethylamine>
627 (accessed 1.20.18).
- 628 Kamp, E., Eisenbrand, G., 1991. Long-chain N-nitroso-N-methylalkylamines in commercial cosmetics,
629 light-duty dishwashing liquids and household cleaning preparations. *Food Chem. Toxicol.* 29,
630 203–209.
- 631 Kemper, J.M., Walse, S.S., Mitch, W.A., others, 2010. Quaternary amines as nitrosamine precursors: a
632 role for consumer products? *Environ. Sci. Technol.* 44, 1224–1231.
- 633 Krasner, S.W., Mitch, W.A., McCurry, D.L., Hanigan, D., Westerhoff, P., 2013. Formation, precursors,
634 control, and occurrence of nitrosamines in drinking water: a review. *Water Res.* 47, 4433–
635 4450.
- 636 Kulshrestha, P., McKinstry, K.C., Fernandez, B.O., Feelisch, M., Mitch, W.A., 2010. Application of an
637 optimized total N-nitrosamine (TONO) assay to pools: placing N-nitrosodimethylamine
638 (NDMA) determinations into perspective. *Environ. Sci. Technol.* 44, 3369–3375.
- 639 Le Roux, J., Gallard, H., Croué, J.-P., 2011. Chloramination of nitrogenous contaminants
640 (pharmaceuticals and pesticides): NDMA and halogenated DBPs formation. *Water Res.* 45,
641 3164–3174.
- 642 Lee, C., Schmidt, C., Yoon, J., Von Gunten, U., 2007. Oxidation of N-nitrosodimethylamine (NDMA)
643 precursors with ozone and chlorine dioxide: kinetics and effect on NDMA formation potential.
644 *Environ. Sci. Technol.* 41, 2056–2063.
- 645 Li, X., Brownawell, B.J., 2010. Quaternary ammonium compounds in urban estuarine sediment
646 environments—a class of contaminants in need of increased attention? *Environ. Sci. Technol.* 44,
647 7561–7568.
- 648 Lijinsky, W., Reuber, M.D., Riggs, C.W., 1983. Carcinogenesis by combinations of N-nitroso compounds
649 in rats. *Food Chem. Toxicol.* 21, 601–605.
- 650 Lindim, C., van Gils, J., Cousins, I.T., Kühne, R., Georgieva, D., Kutsarova, S., Mekenyan, O., 2017. Model-
651 predicted occurrence of multiple pharmaceuticals in Swedish surface waters and their flushing
652 to the Baltic Sea. *Environ. Pollut.* 223, 595–604.

- 653 Mansour, F., Al-Hindi, M., Saad, W., Salam, D., 2016. Environmental risk analysis and prioritization of
654 pharmaceuticals in a developing world context. *Sci. Total Environ.* 557, 31–43.
- 655 Marti, E.J., Pisarenko, A.N., Peller, J.R., Dickenson, E.R., 2015. N-nitrosodimethylamine (NDMA)
656 formation from the ozonation of model compounds. *Water Res.* 72, 262–270.
- 657 Martínez-Carballo, E., Sitka, A., González-Barreiro, C., Kreuzinger, N., Fürhacker, M., Scharf, S., Gans, O.,
658 2007. Determination of selected quaternary ammonium compounds by liquid chromatography
659 with mass spectrometry. Part I. Application to surface, waste and indirect discharge water
660 samples in Austria. *Environ. Pollut.* 145, 489–496.
- 661 Mitch, W.A., Schreiber, I.M., 2008. Degradation of tertiary alkylamines during
662 chlorination/chloramination: implications for formation of aldehydes, nitriles,
663 halonitroalkanes, and nitrosamines. *Environ. Sci. Technol.* 42, 4811–4817.
- 664 Mitch, W.A., Sedlak, D.L., 2004. Characterization and fate of N-nitrosodimethylamine precursors in
665 municipal wastewater treatment plants. *Environ. Sci. Technol.* 38, 1445–1454.
- 666 Mitch, W.A., Sharp, J.O., Trussell, R.R., Valentine, R.L., Alvarez-Cohen, L., Sedlak, D.L., 2003. N-
667 nitrosodimethylamine (NDMA) as a drinking water contaminant: a review. *Environ. Eng. Sci.*
668 20, 389–404.
- 669 Morrison, J.B., Hecht, S.S., 1982. N-nitroso-N-methyldodecylamine and N-nitroso-N-
670 methyltetradecylamine in household dishwashing liquids. *Food Chem. Toxicol.* 20, 583–586.
- 671 Padhye, L., Luzinova, Y., Cho, M., Mizaikoff, B., Kim, J.-H., Huang, C.-H., 2011. PolyDADMAC and
672 dimethylamine as precursors of N-nitrosodimethylamine during ozonation: reaction kinetics
673 and mechanisms. *Environ. Sci. Technol.* 45, 4353–4359.
- 674 Park, S.H., Padhye, L.P., Wang, P., Cho, M., Kim, J.-H., Huang, C.-H., 2015. N-nitrosodimethylamine
675 (NDMA) formation potential of amine-based water treatment polymers: Effects of in situ
676 chloramination, breakpoint chlorination, and pre-oxidation. *J. Hazard. Mater.* 282, 133–140.
- 677 Park, S.-H., Piyachaturawat, P., Taylor, A.E., Huang, C.-H., 2009a. Potential N-nitrosodimethylamine
678 (NDMA) formation from amine-based water treatment polymers in the reactions with
679 chlorine-based oxidants and nitrosifying agents. *Water Sci. Technol. Water Supply* 9, 279–288.
- 680 Park, S.-H., Wei, S., Mizaikoff, B., Taylor, A.E., Favero, C., Huang, C.-H., 2009b. Degradation of amine-
681 based water treatment polymers during chloramination as N-nitrosodimethylamine (NDMA)
682 precursors. *Environ. Sci. Technol.* 43, 1360–1366.
- 683 Patiny, L., Borel, A., 2013. ChemCalc: A Building Block for Tomorrow's Chemical Infrastructure. *J.*
684 *Chem. Inf. Model.* 53, 1223–1228. <https://doi.org/10.1021/ci300563h>
- 685 Richardson, S.D., Ternes, T.A., 2014. Water analysis: emerging contaminants and current issues. *Anal.*
686 *Chem.* 86, 2813–2848.
- 687 Russell, C.G., Blute, N.K., Via, S., Wu, X., Chowdhury, Z., others, 2012. Nationwide assessment of
688 nitrosamine occurrence and trends. *J.-Am. Water Works Assoc.* 104, E205–E217.
- 689 Sacher, F., von Gunten, U., Lee, C., Schmidt, C., 2008. Strategies for minimizing nitrosamine formation
690 during disinfection. Water Environment Research Foundation.
- 691 SCCS, S.C. on C.S.-E.C., 2012. Opinion on Nitrosamines and Secondary Amines in Cosmetic Products
692 [WWW Document]. URL
693 [https://ec.europa.eu/health/sites/health/files/scientific_committees/consumer_safety/docs/
694 sccs_o_090.pdf](https://ec.europa.eu/health/sites/health/files/scientific_committees/consumer_safety/docs/sccs_o_090.pdf) (accessed 8.26.17).
- 695 Scheurer, M., Michel, A., Brauch, H.-J., Ruck, W., Sacher, F., 2012. Occurrence and fate of the antidiabetic
696 drug metformin and its metabolite guanylurea in the environment and during drinking water
697 treatment. *Water Res.* 46, 4790–4802.
- 698 Schmidt, C.K., Sacher, F., 2006. Strategies for minimizing formation of NDMA and other nitrosamines
699 during disinfection of drinking water, in: 2006 Water Quality Technology Conference and
700 Exposition Proceedings.
- 701 Schreiber, I.M., Mitch, W.A., 2005. Influence of the order of reagent addition on NDMA formation
702 during chloramination. *Environ. Sci. Technol.* 39, 3811–3818.
- 703 Schwarzenbach, R.P., Escher, B.I., Fenner, K., Hofstetter, T.B., Johnson, C.A., Gunten, U. von, Wehrli, B.,
704 2006. The Challenge of Micropollutants in Aquatic Systems. *Science* 313, 1072–1077.
705 <https://doi.org/10.1126/science.1127291>
- 706 Selbes, M., Kim, D., Ates, N., Karanfil, T., 2013. The roles of tertiary amine structure, background
707 organic matter and chloramine species on NDMA formation. *Water Res.* 47, 945–953.

- 708 Shah, A.D., Krasner, S.W., Lee, C.F.T., von Gunten, U., Mitch, W.A., 2012. Trade-offs in disinfection
709 byproduct formation associated with precursor preoxidation for control of N-
710 nitrosodimethylamine formation. *Environ. Sci. Technol.* 46, 4809–4818.
- 711 Shah, A.D., Mitch, W.A., 2011. Halonitroalkanes, halonitriles, haloamides, and N-nitrosamines: a critical
712 review of nitrogenous disinfection byproduct formation pathways. *Environ. Sci. Technol.* 46,
713 119–131.
- 714 Shen, R., Andrews, S.A., 2011a. Demonstration of 20 pharmaceuticals and personal care products
715 (PPCPs) as nitrosamine precursors during chloramine disinfection. *Water Res.* 45, 944–952.
- 716 Shen, R., Andrews, S.A., 2011b. NDMA formation kinetics from three pharmaceuticals in four water
717 matrices. *Water Res.* 45, 5687–5694.
- 718 Spahr, S., Cirpka, O.A., von Gunten, U., Hofstetter, T.B., 2016. Formation of N-nitrosodimethylamine
719 during chloramination of secondary and tertiary amines: Role of molecular oxygen and radical
720 intermediates. *Environ. Sci. Technol.* 51, 280–290.
- 721 Trautwein, C., Berset, J.-D., Wolschke, H., Kümmerer, K., 2014. Occurrence of the antidiabetic drug
722 Metformin and its ultimate transformation product Guanylurea in several compartments of the
723 aquatic cycle. *Environ. Int.* 70, 203–212.
- 724 US EPA, O., 2014. Chemical Contaminants - CCL 4 [WWW Document]. US EPA CCL4. URL
725 <https://www.epa.gov/ccl/chemical-contaminants-ccl-4> (accessed 1.24.18).
- 726 US EPA, O.R.D., 2005. METHOD 521: DETERMINATION OF NITROSAMINES IN DRINKING WATER BY
727 SOLID PHASE EXTRACTION AND CAPILLARY COLUMN GAS CHROMATOGRAPHY WITH LARGE
728 VOLUME INJECTION AND CHEMICAL IONIZATION TANDEM MASS SPECTROMETRY (MS/MS)
729 [WWW Document]. URL
730 https://cfpub.epa.gov/si/si_public_record_report.cfm?dirEntryId=103912 (accessed 7.18.17).
- 731 Woods-Chabane, G.C., Glover, C.M., Marti, E.J., Dickenson, E.R., 2017. A novel assay to measure tertiary
732 and quaternary amines in wastewater: An indicator for NDMA wastewater precursors.
733 *Chemosphere* 179, 298–305.
- 734 Yavuz, M., Oggioni, M., Yetis, U., Dilek, F.B., 2015. Biocides in drinking water system of Ankara, Turkey.
735 *Desalination Water Treat.* 53, 3253–3262.
- 736 Ying, G.-G., 2006. Fate, behavior and effects of surfactants and their degradation products in the
737 environment. *Environ. Int.* 32, 417–431.
- 738 Zeng, T., Li, R.J., Mitch, W.A., 2016. Structural Modifications to Quaternary Ammonium Polymer
739 Coagulants to Inhibit N-Nitrosamine Formation. *Environ. Sci. Technol.* 50, 4778–4787.
- 740 Zeng, T., Mitch, W.A., 2015. Contribution of N-nitrosamines and their precursors to domestic sewage
741 by greywaters and blackwaters. *Environ. Sci. Technol.* 49, 13158–13167.
- 742 Zhang, A., Li, Y., Song, Y., Lv, J., Yang, J., 2014. Characterization of pharmaceuticals and personal care
743 products as N-nitrosodimethylamine precursors during disinfection processes using free
744 chlorine and chlorine dioxide. *J. Hazard. Mater.* 276, 499–509.

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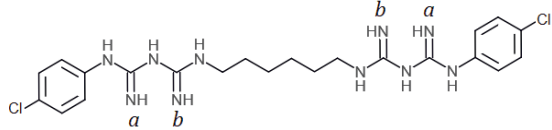
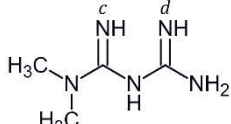
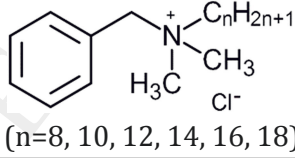
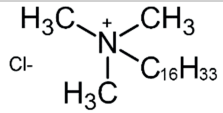
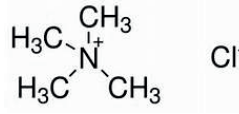
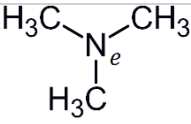
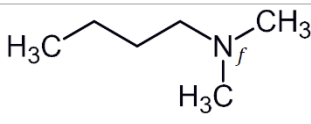
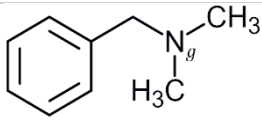
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Table 1 Selected micropollutants and model compounds: molecular structure, molecular weight (MW) and experimentally determined pK_a values (references given in the footnotes). For CHD and MET, the pK_a values are referred to the equilibrium $C=NH_2^+ \rightleftharpoons C=NH + H^+$.

Compound	Structure	MW (g mol ⁻¹)	pK_a
Micropollutants			
Chlorhexidine dihydrochloride (CHD)		578.4	2.20 ^a 10.30 ^b
Metformin hydrochloride (MET)		165.6	10.27 ^c 2.97 ^d
Benzalkonium chloride (BZK)		283.4-423.4 (average 348.4)	-
Cetyl trimethyl ammonium chloride (CTMA)		319.4	-
Model compounds			
Tetramethyl ammonium chloride (TeMA)		109.6	-
Trimethyl amine (TMA)		59.1	9.80 ^e
<i>N,N</i> -dimethylbutyl amine (DMBA)		101.2	10.19 ^f
<i>N,N</i> -dimethylbenzyl amine (DMBzA)		135.2	8.91 ^g

772 ^{a, b} (Agarwal et al., 2012)

773 ^{c, d} (Devi, 2013)
 774 ^{e, f, g} ChemIDplus-Toxnet database (<https://chem.nlm.nih.gov/chemidplus/>).
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778 **Table 2** Molar yields of NDMA and TONO observed in formation potential (FP) experiments under
 779 experimental setup ES1 (reaction time: 7 days; initial precursor concentration: 0.1 mM for CHD,
 780 0.5 mM for the other compounds; initial NH₂Cl molar excess: 10-fold for BZK, 200-fold for the other
 781 compounds) and ES2 (reaction time: 7 days; initial precursor concentration: 0.5 μM; initial NH₂Cl
 782 molar excess: 200-fold). Reported molar yields are the average values of multiple measurement
 783 repetitions (two for NDMA and three for TONO) on two experimental replicates.

Precursor	pH	Molar yield (%)		
		NDMA ^a	TONO ^b	
		average (range)	average ± st.dev.	
		ES1	ES1	ES2
Micropollutants				
Chlorhexidine hydrochloride (CHD)	6	<0.09	2.64 ± 1.45	-
	7	<0.09	4.93 ± 1.33	-
	8	<0.09	11.92 ± 0.83	6.54 ± 0.39
Metformin dihydrochloride (MET)	6	<0.09	0.04 ± 0.003	-
	7	<0.09	0.09 ± 0.01	-
	8	<0.09	0.21 ± 0.03	0.77 ± 0.32
Benzalkonium chloride (BZK)	6	<0.09	0.68 ± 0.24	-
	7	<0.09	1.31 ± 0.21	-
	8	0.56 (0.55-0.56)	2.12 ± 0.40	4.05 ± 0.38
Cetyl trimethyl ammonium chloride (CTMA)	6	<0.09	1.15 ± 0.09	-
	7	<0.09	4.77 ± 1.00	-
	8	<0.09	4.93 ± 1.43	4.05 ± 0.01
Model compounds				
Tetramethyl ammonium chloride (TeMA)	6	<0.09	0.04 ± 0.002	-
	7	<0.09	0.07 ± 0.004	-
	8	<0.09	0.90 ± 0.08	-
Trimethyl amine (TMA)	6	<0.09	<1.0	-
	7	<0.09	0.03 ± 0.005	-
	8	0.10 ^c	0.29 ± 0.007	-
<i>N,N</i> -dimethylbutyl amine (DMBA)	6	<0.09	<1.0	-
	7	<0.09	0.02 ± 0.004	-
	8	0.05 ^c	0.17 ± 0.03	-
<i>N,N</i> -dimethylbenzyl amine (DMBzA)	6	<0.09	<1.0	-

	7	1.22 (1.08-1.36)	0.72 ± 0.03	-
	8	5.05 (3.92-6.18)	3.27 ± 0.34	-

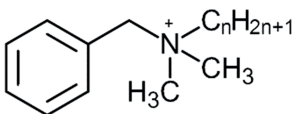
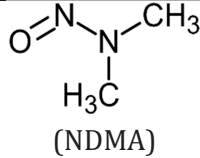
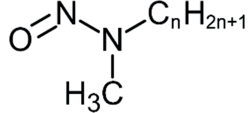
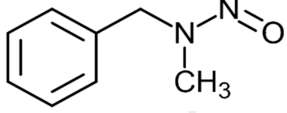
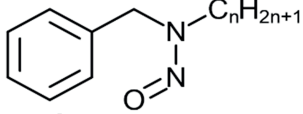
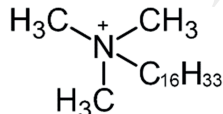
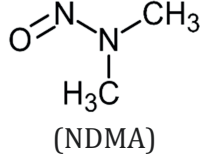
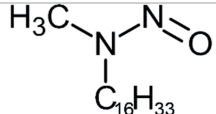
784 ^a LOQ for NDMA: 0.09±0.0001 μM.

785 ^b LOQ for TONO: 1.0±0.45 μM.

786 ^c NDMA concentration >LOQ only in one sample.

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788 **Table 3** Structure, formulas and exact masses of the hypothetical *N*-nitrosamines formed from BZK
 789 and CTMA. Only *N*-nitroso-*N*-methyl dodecylamine could be confirmed as a product from
 790 chloramination of BZK.

Precursor	Hypothesized <i>N</i> -nitrosamines			
	Structure	Formula	Exact mass (<i>m/z</i>)	
			[M] ⁺	[M+H] ⁺
BZK  n=8, 10, 12, 14, 16, 18	 (NDMA)	C ₂ H ₆ N ₂ O	74.0480	75.0558
	 n=8, 10, 12, 14, 16, 18	C _{n+1} H _{2n+4} N ₂ O	172.1576 200.1888 228.2202 256.2515 284.2828 312.3141	173.1654 201.1967 229.2279 257.2593 285.2906 313.3219
		C ₈ H ₁₀ N ₂ O	150.0793	151.0871
	 n=8, 10, 12, 14, 16, 18	C _{n+7} H _{2n+8} N ₂ O	248.1888 276.2202 304.2515 332.2828 360.3141 388.3454	249.1967 277.2280 305.2593 333.2906 361.3219 389.3532
	CTMA 	 (NDMA)	C ₂ H ₆ N ₂ O	74.0480
		C ₁₇ H ₃₆ N ₂ O	284.2828	285.2906

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800 Figure captions

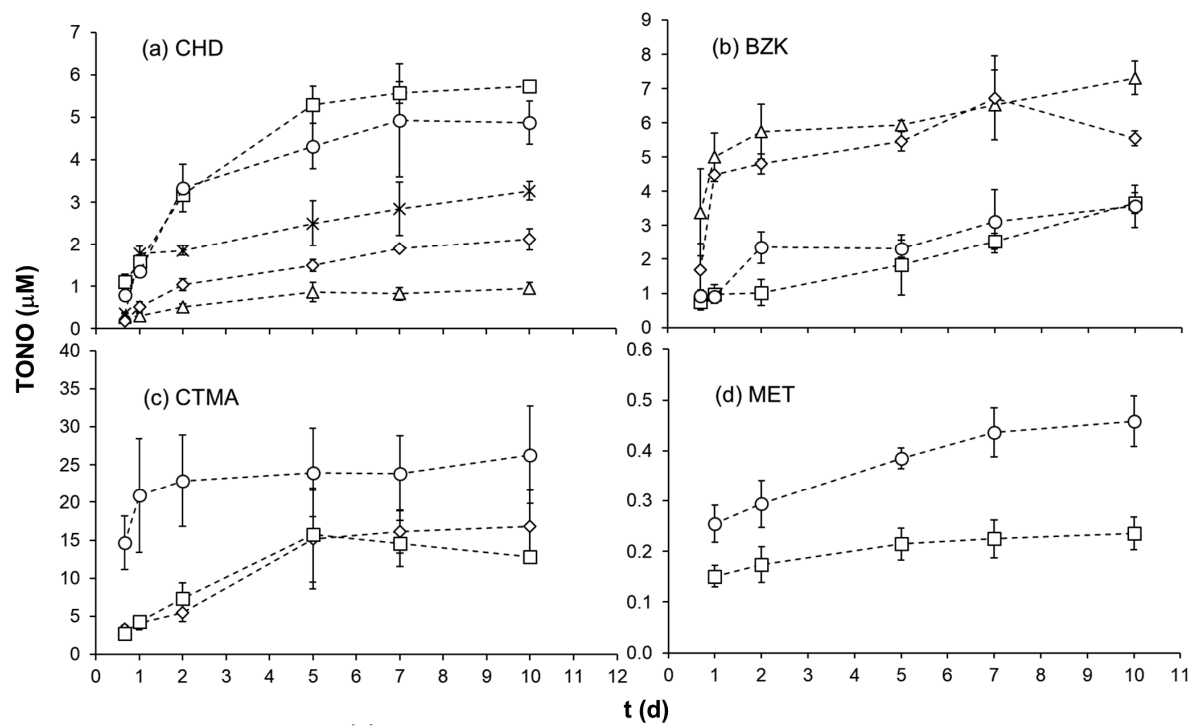
801 **Figure 1** Kinetics of TONO formation at pH 7 for (a) CHD, (b) BZK, (c) CTMA and (d) MET as a function
802 of the initial molar NH₂Cl excess: 10- (Δ), 20- (\diamond), 40- (\times), 100- (\square) and 200-fold (\circ) NH₂Cl molar
803 excess. The initial precursor concentration was set at 0.1 mM for CHD and 0.5 mM for the other
804 compounds. Data represent the average \pm standard deviation of three measurement repetitions on two
805 experimental replicates.

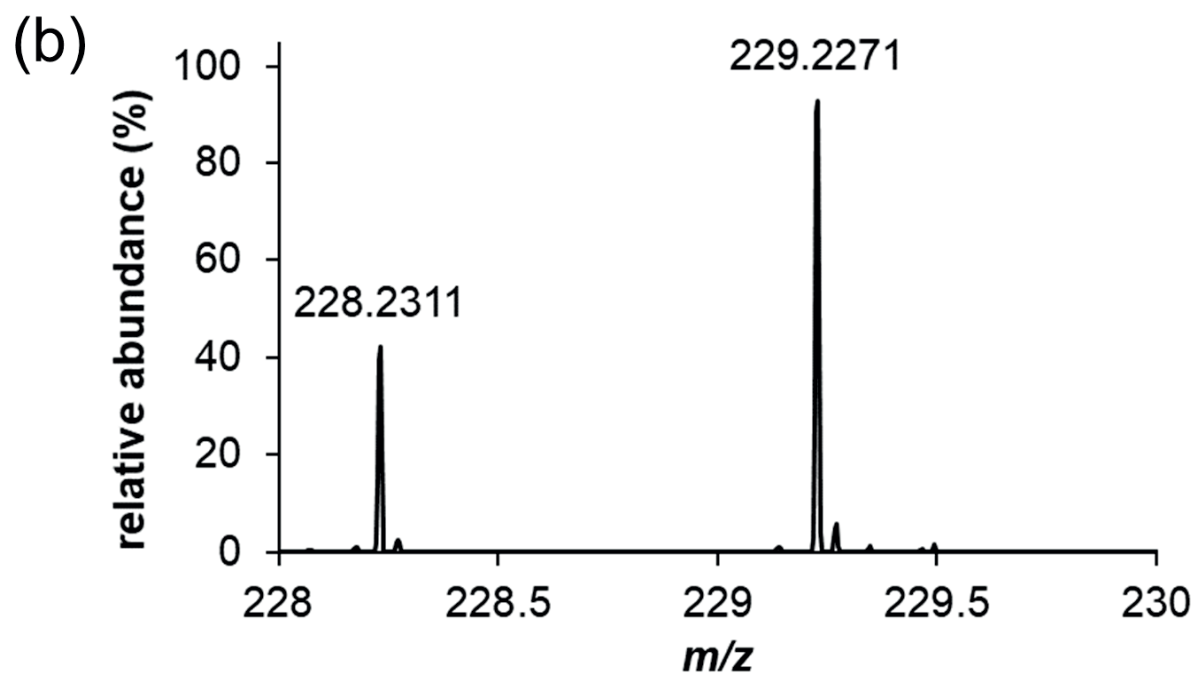
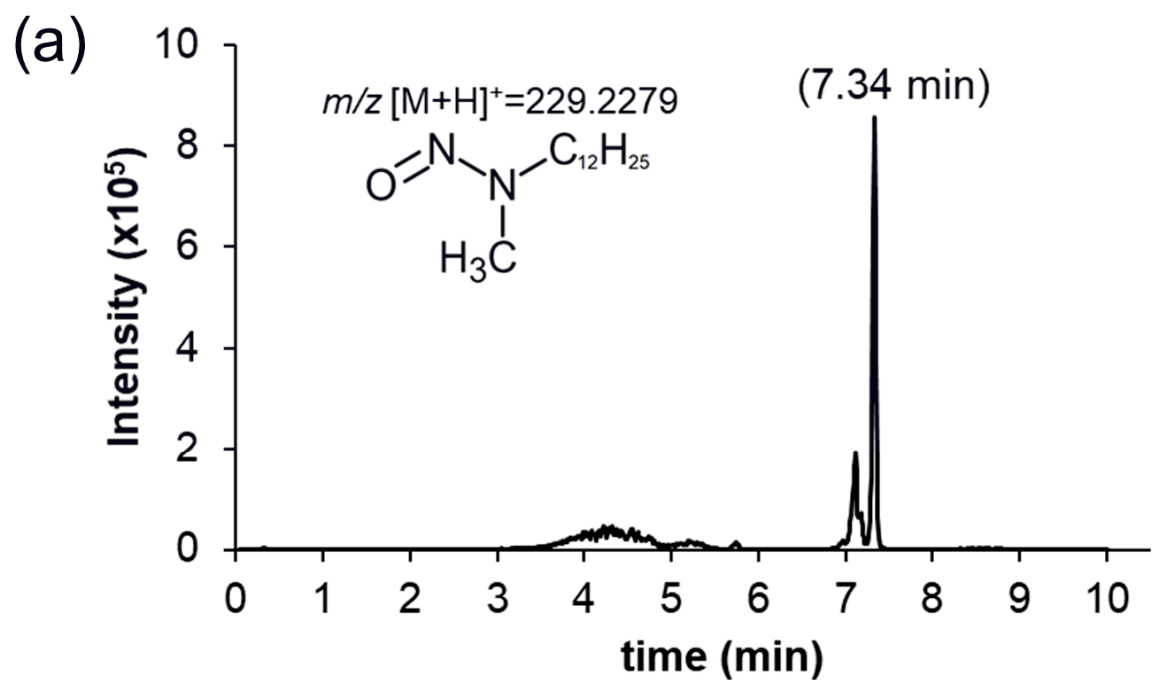
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807 **Figure 2.** High resolution LC-qTOF results for a BZK sample (FP test at pH 8): (a) chromatogram
808 obtained by extraction of the exact mass of C₁₃H₂₈N₂O (m/z [M+H]⁺=229.2279); (b) mass spectrum at
809 retention time=7.34 min.

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Total *N*-nitrosamine (TONO) formation yield of various micropollutants was determined
TONO pools were mostly uncharacterized, only NDMA could be quantified
A LC-qTOF screening procedure was used to characterize the TONO pool
N-nitrosamine formation yield depends on the moieties in precursor molecule

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