

1 Use of CFD For Sampling Hood Design: An Example For Methane Flux

2 Assessment on Landfill Surfaces

3 Federico Lucernoni, Matteo Rizzotto, Federica Tapparo, Laura Capelli, Selena Sironi, Valentina Busini*

4 Politecnico di Milano - Department of Chemistry, Materials and Chemical Engineering "Giulio Natta" -

5 Piazza Leonardo da Vinci 32, 20133 Milano, Italy

6 *Corresponding author: valentina.busini@polimi.it; Tel. +390223993186

7

8 ABSTRACT

9 This study provides a method and a set of guidelines for the assessment of landfill gas surface emissions.
10 The work focuses on the principles for the design of a specific static hood and on the definition of an
11 optimal sampling procedure. This is carried out by means of computational fluid dynamics (CFD)
12 simulations to investigate the fluid dynamics conditions of the hood. The study proves how CFD may be as
13 useful tool for the design and evaluation of sampling systems, thus corroborating the principle that
14 understanding the fluid dynamic conditions is fundamental in order to understand the sampling results and
15 correctly interpret the measured concentration values by relating them to a suitable volatilization model,
16 and therefore to estimate emission rates. The procedure here discussed, which is specific for the case of
17 the investigated landfill, can be generalized to be applied also to different scenarios, where hood sampling
18 is involved.

19 **Keywords:** landfill gas; emissions assessment; computational fluid dynamics; passive area sources; static
20 hood; flux chamber.

21

22 1. INTRODUCTION

23 Environmental sampling for the assessment of emissions from area sources is a quite complicated task
24 (Capelli et al., 2013). The most widely-used techniques for the evaluation of emission rates from area
25 sources the so called “hood methods”, involve an enclosure of some sort. Emission rates are then derived
26 from the data regarding the concentration of the compounds of interest, measured in the samples
27 collected at the outlet of the sampling device combined with the dimensions of the device and the
28 operating conditions (Hudson and Ayoko, 2008; Jiang and Kaye, 1996).

29 Sampling on landfill surfaces for the assessment of landfill gas (LFG) emissions is even more problematic,
30 due to the specific characteristics of this kind of source; in facts, although surely it is not an active area
31 source, it cannot be considered as a passive area source neither, as the landfill surface is typically crossed
32 by a low flux, and there is currently no widely accepted method for sampling on landfills (Bogner et al.,
33 1995; Rachor et al., 2013; Reinhart et al., 1992; Sarkar et al., 2002). Landfills have always been considered
34 important sources of atmospheric pollution, mainly due to the emissions of landfill gas (LFG) from the
35 landfill surface, which entails the emissions of greenhouse gases, VOCs and unpleasant odours (Davoli et
36 al., 2003; Palmiotto et al., 2014; Park and Shin, 2001; Sironi et al., 2005). For this reason, the development
37 of specific methods for the assessment of the emissions would be of great interest both for environmental
38 authorities and for landfill managers and operators (Capelli et al., 2015).

39

40 In the scientific community, it is possible to find a certain agreement concerning the kind of device that
41 should be used for methane concentration measures on landfill surfaces: the instrument should be a static
42 accumulation box of some sort. The hood should allow the methane to diffuse in the internal volume
43 undisturbed (i.e. no external flow) making it possible to measure the concentration over time and then
44 obtain the LFG emission flux. There are several possibilities for the static hood design, as there are different
45 proposals found both in regulations and in scientific papers (UK EA, 2010; Rachor et al., 2013; Lucernoni et
46 al., 2015).

47 In truth, there is an alternative for methane concentration measures on landfill surfaces, which is the
48 procedure exploiting a specifically designed Flux Chamber device, as suggested by the US EPA (Klenbusch,
49 1986). In the present study, the Static Hood was preferred for logistical reasons, linked to the high neutral
50 air demand when operating with a Flux Chamber. Moreover, in a previous research, it was proved that the
51 results obtained with the two methodologies – on the same source – are very much alike (Capelli et al.,
52 2014); thus, it is possible to use either one of the two methods.

53 The present study aims to provide a method and a set of useful guidelines for the design of a static
54 chamber to be used in LFG surface emissions measurements. Moreover, the study aspires to provide hints
55 on what should be the most appropriate sampling methodology to adopt in the field. This will be carried
56 out by means of computational fluid dynamics (CFD) simulations of the phenomenon. Anyway, the
57 procedure here discussed, specific for the case of the investigated landfill, can be applied also for different
58 scenarios, where for a different site the goal is to design a hood analogue to the device here presented.

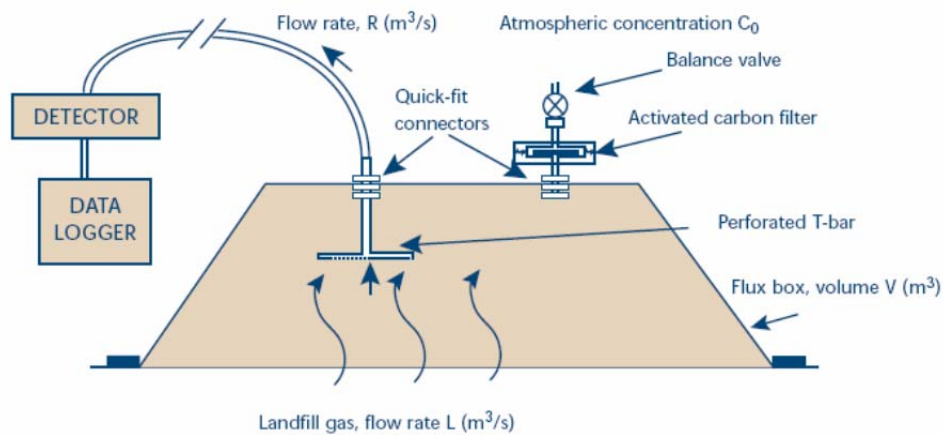
59 As a result, this project will try to corroborate the idea that CFD is a useful tool and can be effectively
60 applied to the design and evaluation of sampling hoods, as discussed in recent studies in this field (Prata Jr.
61 et al., 2016). This idea is based on the fact that the study of the fluid dynamic conditions inside any type of
62 sampling hood is fundamental to understand the sampling results and correctly interpret the measured
63 concentration values by relating them to a suitable volatilization model, in order to be able to estimate
64 emissions.

65 **2. SAMPLING ON LANDFILL SURFACES: STATE OF THE ART**

66 As aforementioned, even though the hood typology for this kind of inquiries is established – i.e. static
67 hoods – there are different proposed designs presented in the scientific literature. In particular, three
68 designs are considered.

69 The first design is the passive flux box proposed by the British regulation (UK EA, 2010). It consists of an
70 enclosure of known volume, with two ports on the top: one inlet port is used for pressure equilibration
71 while the other one, i.e. the outlet port, is used for withdrawing gas samples. Its main design features are: a

72 simple enclosure with an open base that covers an area of approximately 0.33 - 1 m²; two controlled
 73 openings on the top; a sampling line and a suitable gas detection system (e.g., a portable gas detector, such
 74 as a flame ionization detector, FID) for methane concentration measurement inside the box . The device is
 75 typically made of plastic, about 150 mm high. Some sort of paddle may be fitted within the box to promote
 76 gas mixing (UK EA, 2010). The need for this additional tool depends on the box dimensions and the
 77 emission flux rate. The flux boxes should have a base area less than 1 m² and a pyramidal shape. A typical
 78 flux box scheme as described in the UK guideline is shown in Fig. 1.



79
 80 *Figure 1. Typical UK EA passive flux box scheme (UK EA, 2010).*

81 The device depicted in Fig. 1 shows a reasonable compromise between footprint area, volume to footprint
 82 ratio, ease of construction and practicability on site. The sample must be withdrawn at a point where the
 83 gas is representative of the mixture in the flux box. The sampled volume must not be so great as to cause
 84 significant dilution of the remaining gas. From the measured concentration, it is possible to obtain the
 85 emission according to Eq.1:

86
$$Q = \frac{V_{sb}}{S_{sb}} \cdot \left(\frac{dc}{dt} \right) \quad (1)$$

87 Where Q is the methane emission flux in [mol m⁻² s⁻¹], V_{sb} is the volume of the sampling device in [m³], S_{sb} is
 88 the base area of the sampling device in [m²], c is the methane concentration variation in [mol m⁻³] and t is
 89 the considered time period in [s].

90 The second design is the one proposed in 2013 by a German research group (Rachor et al., 2013). In this
91 case, the device consists of a cylindrical aluminium chamber with a height of 0.5 m and a base area of 0.12
92 m². The base of the chamber is sharpened in order to enable a better insertion into the soil to a depth of
93 few centimetres, which is necessary in order to avoid undesired exchanges of air/gas between the interior
94 and exterior of the chamber. Gas tightness is further guaranteed by the presence of a water-saturated
95 foam on the hood outside. Even though the hood practically works as a static chamber, its functioning has
96 been improved by preventing overpressures inside the chamber due to the emission of landfill gas from the
97 soil: the chamber has in fact an open tube (inner diameter 4 mm) connecting the inner volume with the
98 atmosphere. The tube has been designed long enough (3m) as to compensate for possible pressure
99 differences without allowing significant gas exchange by diffusion. Moreover, the chamber is equipped with
100 three sampling ports connected to a switching valve. As described in the paper by Rachor et al. (2013), their
101 mobile analyser (FID) for the methane concentration measurement inside the chamber, was connected to
102 the valve, logging CH₄ concentrations every minute. The adopted sampling procedure provides that 15
103 seconds before logging takes place, the open tube was closed and the valve to the chamber opened. Given
104 the suction flow rate of the FID (approximately 1 l/min) and the quick instrumental response time (5 s), this
105 procedure prevents the withdrawal of too much gas, which, according to the authors, produced stable and
106 representative readings. As stated in the above mentioned paper, chamber operation according to this
107 procedure usually lasted at least 6 minutes. Gas emissions from the landfill can be calculated from the CH₄
108 concentration measured inside the chamber according to Eq. 2:

$$109 \quad E = \frac{p \times m \times V_{Ch} \times 1440}{10^6 \times R \times T} \quad (2)$$

110 Where E is the methane rate of emission in [mol_{CH₄}/day], p is the ambient pressure in [hPa], m is the slope
111 of the linear regression (C_{gas}/t) in [μmol/mol/min], V_{Ch} is the actual volume of the chamber in [l], R is the gas
112 constant in [J/mol/K] and T is the temperature in the chamber in [K].

113 The third design is the one adopted in the present work, which will be described in the following section.

114 3. MATERIALS AND METHODS

115 3.1 The sampling hood developed at LabOlf

116 The design adopted in this study, consists of a steel static hood with a square base area. The hood realized
117 at POLIMI LabOlf is based on the first design proposed, relying on theoretical considerations, inspired as
118 well by the scientific literature (Rachor et al., 2013; UK EA, 2010). The sampling chamber has a 50 cm x 50
119 cm base: this allows to cover a representative portion of the landfill surface, while granting that the internal
120 volume is adequate for the Flame Ionization Detector (FID) operation. In facts, the FID adopted for CH₄
121 concentration measurement is a Crowcon GasTec FID that sucks about 1,06 l/min for a measure time of 2
122 min, amounting to a total of 2 l circa. This implies that the hood needs to have a volume at least 10 times
123 greater than that value. The device needs also to be easily transportable, i.e. reasonably small dimension,
124 since the sampling points are scattered over a quite large surface. Therefore, a height of 10 cm was chosen
125 (i.e. a chamber volume of 25 l); this value is significantly lower than the 50 cm proposed by Rachor et al.
126 (2013) but it was chosen considering that the phenomenon is mainly diffusive and thus slow. The POLIMI
127 hood is made of steel, a resistant material that allows to insert the hood in the soil without damaging it and
128 minimizing the undesired LFG leaks; the hood is equipped with lateral flanges assuring a determined
129 penetration into the turf in order to grant a fixed effective hood volume of 25 l in all measurements. The
130 chamber is also equipped with 2 pipes: the former 3 m long, located on a lateral wall that assures isobaric
131 conditions during sampling; the latter 10 cm long, located on the top of the chamber that is the sampling
132 port for the FID. The short tube determines also the exact sampling point, where the CH₄ is measured. Such
133 point needs to grant a measure representative of the average concentration inside the hood, in order to
134 have useful results. The position of the sampling point should vary according to the hood height and the
135 emitted LFG flow, since as these two variables change, the average concentration and the concentration
136 distribution in the chamber will change as well. It is easy to understand that this is a very delicate choice
137 and that it is important to investigate how the sampling point position varies for different situations. An
138 example of such chamber is depicted in Fig. 2.

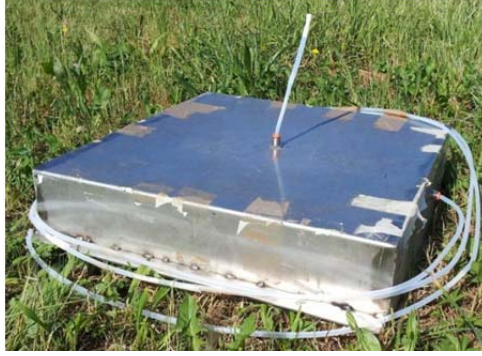


Figure 2. The POLIMI LabOlf static hood.

139

140

141 The sampling lasts about 10 min of hood deposition time plus 2 min of FID operation. Hypothesizing that
 142 the methane concentration measured at the sampling point is representative of the average concentration
 143 in the hood and that the concentration growth over time in the considered horizon time is linear, it is
 144 possible to compute the LFG emission according to Eq. 3. The two hypothesis made will be verified by
 145 means of CFD simulations.

146

$$\hat{Q}_{LFG} = \frac{c_{CH_4,mis} \cdot V}{c_{CH_4,LFG} \cdot \Delta t \cdot S_h} \quad (3)$$

147 Where \hat{Q}_{LFG} is the LFG emission flux in [$l \text{ m}^{-2} \text{ s}^{-1}$], V is the volume of the sampling device in [l], S_h is the base
 148 area of the sampling device in [m^2], $c_{CH_4,mis}$ is the measured methane concentration in [ppm], $c_{CH_4,LFG}$ is
 149 the methane concentration in the LFG in [ppm] and t is the measure time in [s].

150 As mentioned above, in order to determine the specific LFG flow rate emitted from the landfill surface, the
 151 hood is positioned over the soil, for a period defined in 10 min (Lucernoni et al., 2015), then the CH₄
 152 concentration is measured with the FID. The time of 10 minutes was fixed based on two main
 153 considerations. The former, it is a sufficient time to obtain readable CH₄ concentration values inside the
 154 hood. In facts, in a work by Palmiotto et al. (2014), which describes a study carried out on a Municipal Solid
 155 Waste (MSW) landfill in Lombardy, the evaluated mean LFG emission rate turned out to be 0.2 l/m²/h,
 156 giving a concentration of few hundreds ppm inside the hood after 10 minutes, which can be read with the
 157 FID. The latter, is that a period of 10 minutes is short enough to assume that the concentration grows
 158 linearly with time, even if it has an asymptotic “plateau” trend.

159 Actually, the nature of the phenomenon is almost purely diffusive and the sampling hood can be modeled
160 as a Plug Flow Reactor (PFR). After a certain time, the methane concentration inside the hood will reach a
161 stationary value (i.e. the methane concentration in LFG, 50%). In order to have a rough estimation of the
162 time required to reach the plateau, it is possible to compute the ratio between the hood volume and the
163 LFG flow rate, which for the case here considered – with a LFG emission in the range of the tenths of l/m²/h
164 – turned out to be 24000 min circa. For this reason, it is reasonable that in the first 10 minutes the CH₄
165 concentration growth inside the hood can be approximated with a linear trend.

166 The study shows how CFD can be a useful tool for the design of a suitable sampling procedure for particular
167 and complex area sources requiring the use of specific sampling hoods (Prata Jr. et al., 2016).

168 **3.2 Computational Fluid Dynamics (CFD)**

169 Computational Fluid Dynamics (CFD) is the field of fluid mechanics that uses numerical analysis and
170 software to solve problems concerning fluid flows, heat transfer and chemical reactions. The chosen
171 software for the present work was ANSYS Fluent v14.5, one of the most used and most validated CFD
172 program available in the market (Fluent User's Guide, 2012). This software helped in defining the most
173 appropriate sampling procedure, allowing to verify whether or not the first design proposed for the hood
174 and the sampling methodology devised were acceptable, in terms of hood dimensions, sampling point and
175 measurement time.

176 In CFD analysis the control volume is divided into numerous cells, and the nonlinear partial differential
177 equations (PDEs), which govern the system, are discretized and solved numerically. So, in order to perform
178 the simulations, some steps have to be done:

- 179 1) Define and create the model geometry
- 180 2) Create the mesh
- 181 3) Set up the solver and choose the proper physical models

182 For steps 1 and 2, the grids were built with the aid of the software GAMBIT, which allows to define the
183 mesh with the desired characteristics. This is very important, in order to guarantee accuracy and

184 convergence of the FLUENT solution. The meshes were made of tetrahedral cells and presented a
185 “boundary layer” in order to increase accuracy in proximity of the inlet.

186 Then, the boundary conditions need to be defined. The base surface of the hood was set as *mass flow inlet*,
187 the outlet surface of the 3 m long tube (during the 10 min deposition time) as *pressure-outlet*, and the
188 outlet surface of the 10 cm long tube (during the 2 min FID operation) as *velocity-inlet*.

189 For step 3, in the present case the equations considered were those for the mass, momentum and energy
190 conservation, and for the species transport, which allows to account for the diffusion of the LFG – a binary
191 gas mixture of CH₄ and CO₂ – in air.

192 After setting up the whole case, it is possible to run the simulation. When the simulation has finished, a
193 built-in post-processor allows to visualize and investigate the results.

194 It is worth clarifying that the “deposition time” in the present work refers to the period when the chamber
195 is put on top of the sampling area and no induced flows are present. After this “deposition time” is the “FID
196 operation time”, the period when the FID is turned on sucking a defined flow rate from the chamber.

197 3.3 CFD Simulations

198 In order to verify the hypotheses discussed in Par. 3.1, allowing to devise a proper hood design
199 methodology and an adequate sampling procedure, several simulations were set up and performed. In
200 different simulations, key parameters were changed such as hood height and LFG emitted flow, while the
201 base surface of the hood is kept constant. Tab.1 summarizes the different input data set for each
202 simulation.

203 *Table 1 Summary of the performed simulations.*

Simulation	Q_{LFG} [l/(m ² *h)]	Hood Height [cm]	Number of cells	Number of layers of the boundary layer
1	0.25	10	~ 2.7 million	7
2	0.05	10	~ 2.7 million	7
3	1.00	10	~ 2.7 million	7
4	0.25	20	~ 3 million	21
5*	0.25	10	~ 4 million	16

204

205 In Tab. 1, simulation 5 is marked with an asterisk to indicate that it is different from simulation 1, because
206 of the different geometry due to the introduction of a mini-fan, as discussed in the following. In each
207 simulation, both the hood deposition period (10 min) and the FID operation period (2 min) were simulated,
208 in order to verify the effects that suction produces on the system during the analysis.

209 In simulation 1, the hood was simulated according to the first design proposed - i.e. 0.10 m high - and the
210 LFG flow was set equal to $0.25 \text{ l/m}^2/\text{h}$. This flow value was chosen based on the values obtained in previous
211 similar works (Palmiotto et al., 2014; Lucernoni et al., 2015) investigating landfills in northern and central
212 Italy similar to the site here considered. This simulation aims to validate the two hypotheses on which Eq. 3
213 relies for the estimation of the LFG emission rate from the CH_4 concentration measured inside the sampling
214 hood: i) the linear growth of CH_4 concentration during the first 10 min of deposition time, and ii) the
215 representativeness of the sampling point, i.e., the assumption that the value read with the FID is
216 representative of the average concentration inside the hood. It is important to highlight that for all the
217 simulations a first improvement of the hood design was introduced by moving the 3m-long open tube from
218 the lateral to the top of the chamber in order to prevent perturbations on the developing concentration
219 front that might deviate due to the lateral opening.

220 Once this first hypotheses were proven, it was deemed interesting to further investigate the position of the
221 optimal sampling point, i.e. the point in which the methane concentration is equal to the average methane
222 concentration inside the hood, for different sampling conditions.

223 For this reason, different variables were examined by simulating the methane concentration variation
224 inside the hood in modified conditions, in order to verify and prove whether or not the outcomes of the
225 first simulation can be generalized to different field situations (Lucernoni et al., 2015).

226 First, the LFG flow rate was modified from the first attempt value of $0.25 \text{ l/m}^2/\text{h}$, to a significantly lower
227 value of $0.05 \text{ l/m}^2/\text{h}$ (simulation 2) and to a significantly higher value is of $1 \text{ l/m}^2/\text{h}$ (simulation 3), by
228 keeping the hood features constant. The flow rate values were chosen as to remain within the validity

229 boundaries of the considered fluid dynamic regime, which is the laminar diffusive regime, and thus within
230 the boundaries set by the ANSYS Fluent software in terms of cell dimensions, Courant number, etc. Those
231 values are also representative of the limits for LFG emissions from landfill surfaces, typically observed in the
232 field, for northern Italian MSW landfills.

233 Then, the hood volume was modified by changing its height from 10 cm to 20 cm (simulation 4).

234 Simulation 5 was set up in order to investigate possible modifications and improvements to the proposed
235 hood design based on the outcomes of the previous simulations. One main aspect that was considered is
236 the homogenisation of the CH₄ concentration inside the hood before the FID operation, which could be
237 improved by the introduction of a fan, thus giving a sampling procedure less sensitive to the sampling
238 point. The fan was located in the top left corner of the chamber, with a 2.5 cm diameter, spinning at a
239 speed of about 400 rev/min.

240 For all simulations, it was set an air temperature of 17°C and a soil temperature of 16°C, representative of
241 the situation in the field.

242 **4. RESULTS AND DISCUSSION**

243 **4.1 SIMULATIONS RESULTS – FIRST HOOD DESIGN**

244 In order to verify linearity of the CH₄ concentration, the CH₄ concentration was monitored with Fluent at
245 the top of the chamber (i.e. at the inlet of the FID sampling tube), for the 10 min deposition period.

246 The obtained values were linearly interpolated, providing a satisfactory R² correlation index of 0.9994, thus
247 validating the initial hypothesis. Fig. 3 shows the obtained plot for the trend and the corresponding table
248 with the numerical values of the concentration, allowing to visually appreciate this evidence. Moreover, in
249 Fig. 3 also the mean CH₄ concentration in the hood at each minute is reported. This can be evaluated by
250 means of a material balance on the hood that – for the considered situation – can be written as shown in
251 Eq. (4).

252

$$c_{CH_4}^{mean}(t_i) = \frac{q_{LFG} * x_{CH_4}^{LFG}}{3600} * \frac{S_{hood}}{V_{hood}} * (t_i) * (10^6) = \frac{2.5}{7.2} * (t_i) \quad (4)$$

253

Where ($c_{CH_4}^{mean}$) is the mean methane concentration in the hood at time (t_i) in [ppm], (q_{LFG}) is the specific

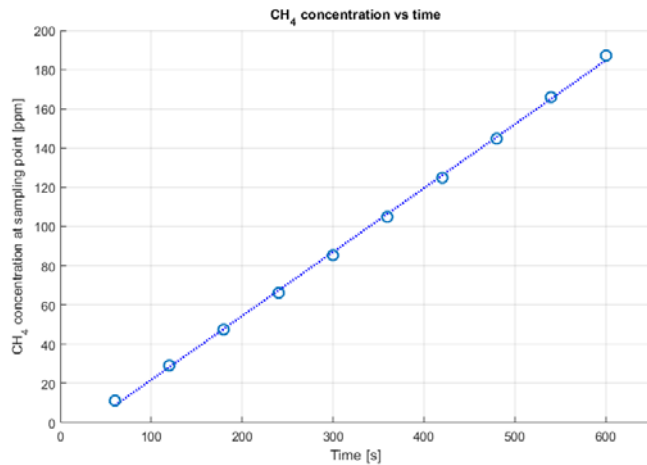
254

LFG flow rate entering the chamber in [l/m²/h], ($x_{CH_4}^{LFG}$) is the methane molar fraction in the LFG, (S_{hood}) is

255

the base area of the hood in [m²], (V_{hood}) is the volume of the hood in [l], (t_i) is the time elapsed in [s].

OUTLET		MEAN
t [s]	c,CH ₄ Fluent [ppm]	c,CH ₄ mb [ppm]
60	11	21
120	29	42
180	48	43
240	66	83
300	86	104
360	105	125
420	125	146
480	145	167
540	166	188
600	187	208



256

Figure 3 CH₄ concentration trend at the FID sampling tube inlet.

257

In Fig. 3, the tabulated methane concentration values are those obtained from the software at the sampling

258

point and those computed as shown in Eq. (4). Comparing them, it is possible to notice how the %

259

discrepancy decreases as time increases, i.e. after 10 min the outlet concentration is representative of the

260

mean value in the hood.

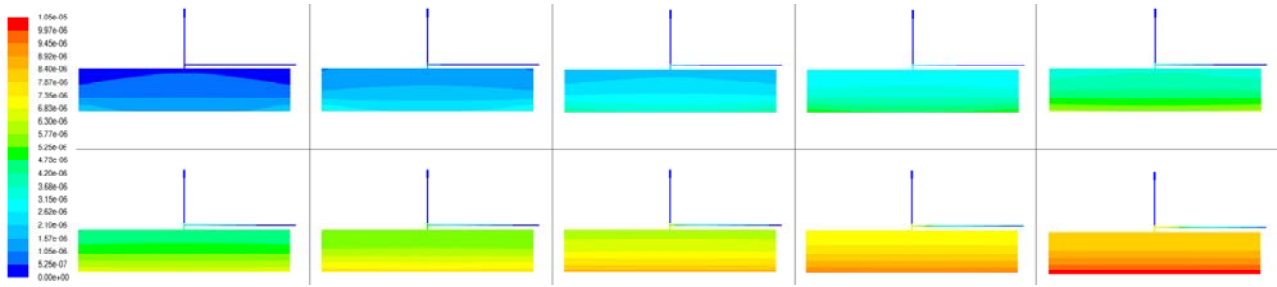
261

The second hypothesis verified with this CFD simulation is the assumption that the collected sample with

262

the FID has a concentration representative of the average concentration in the entire hood volume.

263



264

Figure 4 CH₄ concentration profile in [kmol/m³] inside the chamber during the 10 min deposition time.

265

266

267

268

269

As it is possible to appreciate in Fig. 4, during the 10 min deposition period the CH₄ stratifies in the internal volume, with concentration decreasing with height. Such behaviour is due to the nature of the mass transfer phenomenon, which is purely diffusive. This aspect can be better understood considering that the diffusive velocity in the hood is about 10⁻⁴ m/s, while the convective velocity is about 10⁻⁷ m/s, which can be computed as shown in Eq. (5)-(6).

270

$$v_{diff} \cong \frac{\varphi_{mix}}{h} \quad (5)$$

271

$$v_{conv} \cong \frac{\hat{Q}_{LFG}}{S_h} \quad (6)$$

272

273

274

Where (v_{diff}) is the diffusive velocity in [m/s], (φ_{mix}) is the mixture diffusivity equal to 10⁻⁵ [m²/s], (h) is the hood height in [m], (v_{conv}) is the convective velocity in [m/s], (\hat{Q}_{LFG}) is the LFG emission flux [m³/s], (S_h) is the hood base area in [m²].

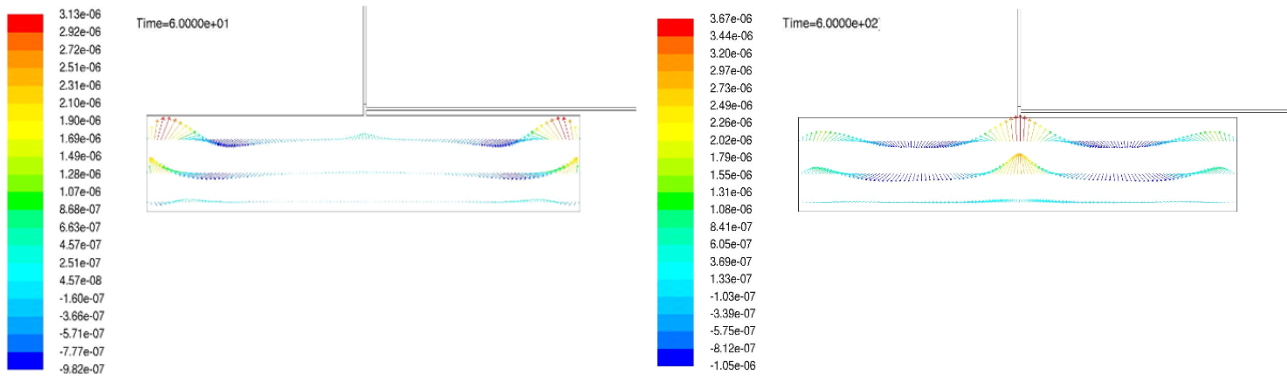
275

276

277

278

Looking at the shape of the concentration profile (Fig. 4), it is possible to notice that in the early moments it assumes a lanceolate pattern, then becoming flat in the following. In order to explain this behaviour, an investigation was performed concerning the velocity vectors in the chamber. In Fig. 5, the velocity vectors in the vertical direction are depicted, after 1 min and after 10 min of hood deposition time.



279

280

Figure 5 Vertical (y-direction) velocity vectors after 1 min (left) and after 10 min (right) of hood deposition time.

281

Looking at Fig. 5 it is possible to see how after 1 min the velocity zeroes at the walls but has maximum

282

intensity in the top corners near the walls, while flattening towards the centre of the hood. This indicates

283

that there are recirculation phenomena in the upper part of the chamber – even though the modules of the

284

velocity are almost negligible – and that there is a “chimney effect” caused by the tube on the top wall. On

285

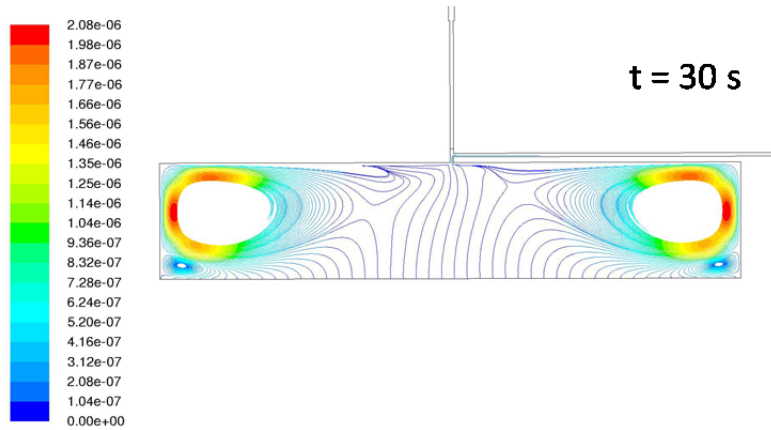
the contrary, after 10 min the situation is reversed, with the higher intensity located at the central pipe

286

inlet (i.e. sampling point), while the velocity vectors near the walls are very small, indicating that there is no

287

more recirculation in the corners of the chamber.



288

289

Figure 6 The velocity path-lines at 30 s of hood deposition time.

290

Fig. 6 shows the velocity path-lines after 30 s of hood deposition time. Such behavior can be explained since

291

in the initial times the CH_4 concentration in the hood is low, thus the convection is the controlling

292

phenomenon, and therefore the effect of the walls and the top tube produces the swirls in the top corners

293

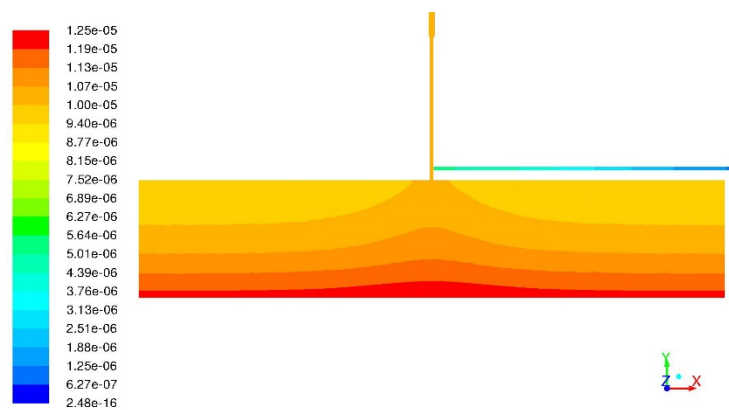
with negative (downward) velocities. This causes an accumulation of CH_4 in the lower corners and gives the

294

peculiar lanceolate shape to the concentration profile. Then, when CH_4 concentration increases, the

295 diffusion becomes dominant “evening up” these effects and producing the flat concentration profile. As
296 discussed in the next paragraphs, this evolution is observed in all situations.

297 Both Fig. 5 and Fig. 6 show that it is possible to have unexpected situations inside the sampling hood during
298 the initial moments whereby sampling would be critical. This is due to the procedure required to position
299 the hood on the soil, which may produce pressure and temperature differences that cause unexpected
300 behaviors inside the chamber. For the case at hand, it was found that a time of 10 minutes is required to
301 dampen the significance of these phenomena and obtain a representative result.



302
303 *Figure 7 CH₄ concentration profile in [kmol/m³] after the 2 min FID measure period.*

304 As stated, after the simulation of the first 10 min, the FID operation was simulated. Looking at Fig. 7, it is
305 possible to see that after the FID operation, the system sustains a perturbation: the system results no
306 longer completely segregated. The CH₄ is distributed more uniformly in the volume. Comparing the
307 concentration values at the sampling tube inlet (242 ppm or 1.02 kmol/m³) and the average one (250 ppm
308 or 1.05 kmol/m³), at the end of the 2 min FID operation, it is confirmed that the chosen sampling point is
309 even more representative.

310 The concentration values are all reported in Tab. 2, for all simulations, at end of the hood deposition time
311 and at the end of the FID operation. It is worth remarking that the values reported in Tab. 2 refer to the
312 concentrations extrapolated from Fluent.

313 Looking now at Fig. 4, at the end of the 10 min deposition period, which is the height where the punctual
314 concentration value is exactly equal to the average, it is possible to see that this point is roughly at 4 cm

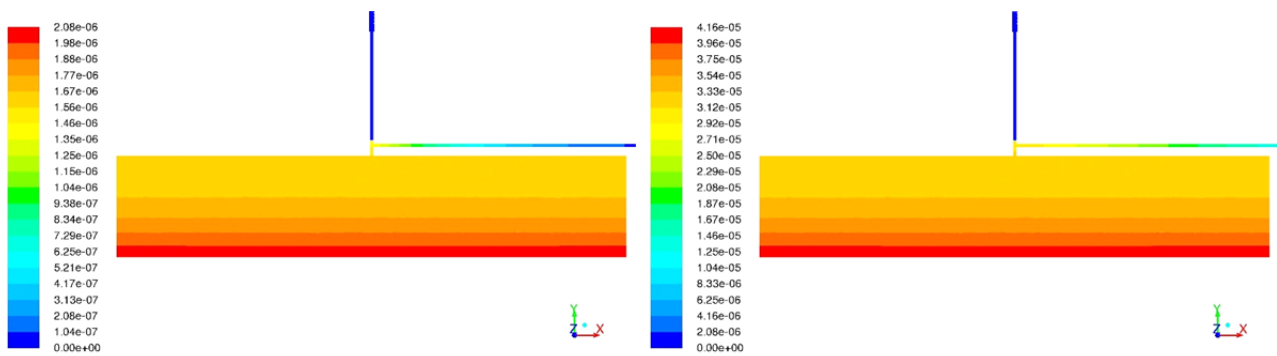
315 from the ground, i.e. at about 40% of the chamber height. The position where the concentration inside the
316 hood is equal to the mean was checked also in the following simulations performed, in order to prove
317 whether or not this situation is purely casual or it is a constant also in different situations.

318 4.2 SIMULATIONS RESULTS – DIFFERENT CONFIGURATIONS

319 4.2.1 DIFFERENT SOURCE CONFIGURATIONS

320 A comparison was carried out among three different situations, with different values of LFG emission flux,
321 for a fixed hood height and volume: standard situation (simulation 1), low emission (simulation 2), high
322 emission (simulation 3) (see par. 3.3).

323 The CH₄ concentration profile inside the hood, for simulations 2 and 3, at the end of the 10 min deposition
324 period is depicted in Fig. 8, left and right respectively:



325
326 *Figure 8 CH₄ concentration profile in [kmol/m³] after the 10 min deposition period, low (left) and high LFG emission (right).*

327 The pictures in Fig. 8 look identical, but since the LFG flow rates differ of a factor 20, also the concentration
328 scales differ of a factor 20, i.e. the concentrations in the bottom are much higher than the ones in the top.

329 Moreover, Fig. 8 shows that the CH₄ concentration is stratified inside the chamber as it was in the previous
330 configuration. With lower flow rates – thus lower convective velocities – the time required to reach the
331 stationary regime is greater.

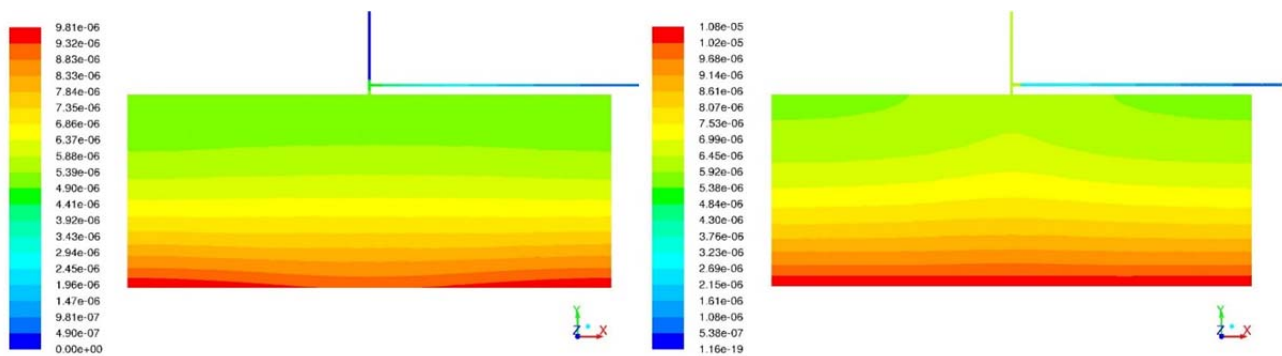
332 Now, looking at the mean CH₄ concentration in the hood at the end of the 10 min deposition time and the
333 concentration at the sampling point, as reported in Tab. 2, it is possible to see that in both cases of low and
334 high emission, the values are comparable, thus confirming the appropriateness of the sampling point.

BEFORE FID	Simulation 1	Simulation 2	Simulation 3	Simulation 4	Simulation 5
c,ave [ppm]	209	42	831	156	208
c,sample [ppm]	187	37	750	120	203
AFTER FID	Simulation 1	Simulation 2	Simulation 3	Simulation 4	Simulation 5
c,ave [ppm]	250	50	1001	178	250
c,sample [ppm]	242	49	965	147	248

335 *Table 2 Sampling point and average CH₄ concentration, after the 10 min hood deposition period and after the 2 min FID operation.*
336 Similarly to the previous situation, after the 2 min FID operation period, a certain degree of mixing of the
337 system is observed for both configurations, giving that the mean concentration and the concentration at
338 the sampling point are similar, as shown in Tab. 2. The height of the iso-concentration surface
339 corresponding to the average concentration in the chamber is found at about 40-45% of the hood height.

340 4.2.2 DIFFERENT SAMPLING SYSTEM CONFIGURATION

341 In this simulation, the hood volume was modified by changing its height from 10 cm to 20 cm.
342 The CH₄ concentration profile in the hood, at the end of the 10 min deposition period, is depicted in Fig. 9
343 (left).



344
345 *Figure 9 CH₄ concentration profile in [kmol/m³] inside the chamber, at the end of the 10 min deposition time (left) and after the 2*
346 *min FID operation (right).*

347 In this case as well, stratification of the methane concentration is observed, even though the profile results
348 less flat. Referring to the discussion concerning Fig. 4, this is probably because the volume is double and it
349 takes longer for the profile to “even up”.

350 Similarly to the previous situations, after the 2 min FID operation period, a certain degree of mixing of the
351 system is observed, even if such effect is much less marked in this case, as shown in Fig. 9 (right).

352 Looking at Tab. 2, it is possible to see that also in this case the mean CH_4 and the value at the sampling
353 point are comparable, both before (i.e. after the 10 min deposition period) and after the sampling (i.e. after
354 the 2 min FID operation period).

355 Also in this case, it is possible to see that the height at which the concentration is equal to the main value is
356 at about 40-45% of the chamber height.

357

358 **4.3 DISCUSSION**

359 The simulations presented in the previous sections yielded several interesting results. First, they allowed to
360 validate the hypothesis made concerning the linear growth of the CH_4 concentration in the hood over the
361 first minutes of sampling and the choice of a proper sampling point where the concentration is
362 representative of the mean value inside the chamber, for the case of the first hood design proposed. In
363 addition, it was highlighted that in any case, after the 2 min FID operation, the system results mixed.

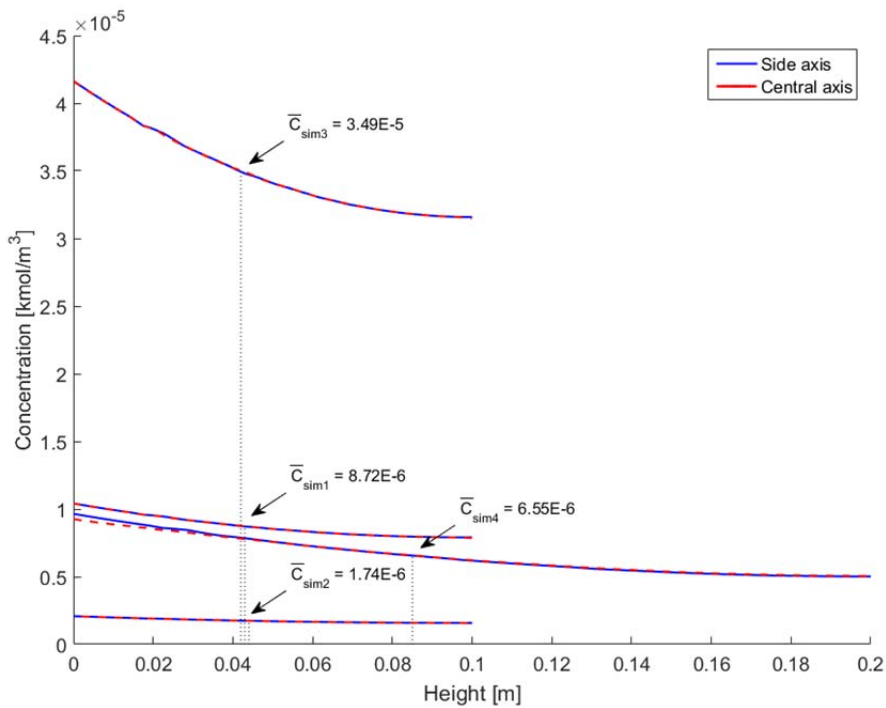
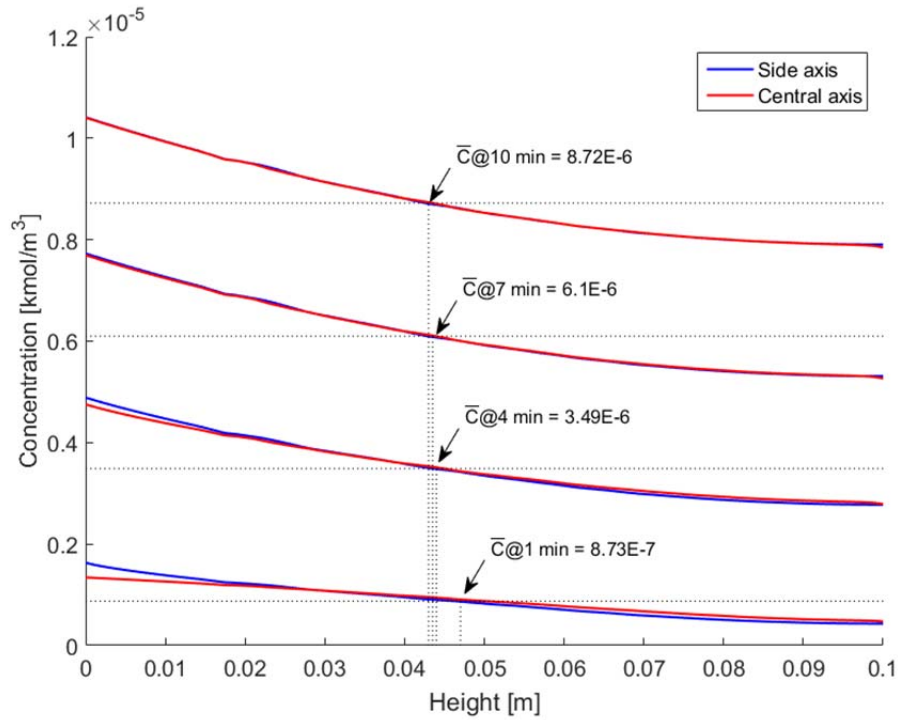
364 It was possible to see that the height where the concentration iso-surface is the mean value in the hood –
365 averaged on the entire volume – is always at a level above ground roughly equal to the 40-45% of the
366 chamber height. As it is shown in Fig. 10, it is possible to see that in all the considered situations, the
367 methane concentration values on the central axis of the hood the average value is always found at a height
368 slightly lower than 50% of hood height (i.e. 40-45%).

369 Fig. 10 (top) shows the concentration values for Simulation 1, taken at different times (1, 4, 7, 10 min) and
370 at different heights (from ground to top); the average value is always found at a height slightly lower than
371 50% of hood height, starting at a value close to 50% (at 1 min) and then settling around 43%.

372 Fig. 10 (bottom) shows the concentrations for the Simulations 1-4 at the end of the first 10 min: in each
373 case the average concentration is reached around the 40-45% of the hood height.

374 Moreover, in Fig. 10, the methane concentrations are reported for two vertical axis in the hood, one
375 corresponding to the sampling tube axis (central axis), at 0.25 m from the lateral wall and one at a distance
376 of 0.1 m from the lateral wall (side axis).

377 This leads to the conclusion that when operating a sampling hood in a situation similar to the ones
378 considered in the present study, the ideal sampling point should be at a level somewhat lower than 50% of
379 the hood height.



380

381

Figure 10 CH₄ concentration trends over time and space.

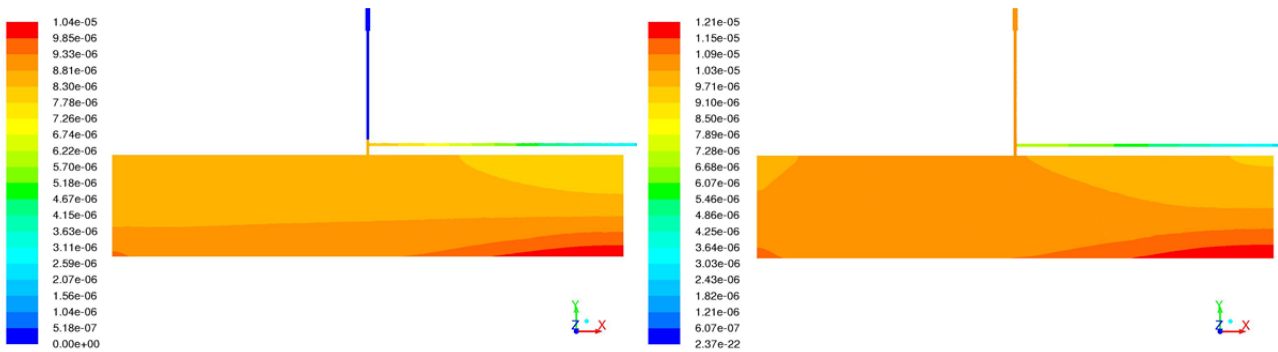
382 **4.4 SIMULATIONS RESULTS – MODIFIED HOOD DESIGN**

383 Simulation 5 entails the presence of mini-fan that should help in achieving a satisfactory mixing degree

384 inside the chamber also before the FID operation. The mini-fan chosen for the simulation is a Jamicon®

385 KF0210S5MR axial fan, 25x25x10 mm.

386 The CH₄ concentration profile inside the hood, at the end of the 10 min deposition time, is represented in
387 Fig. 11 (left).

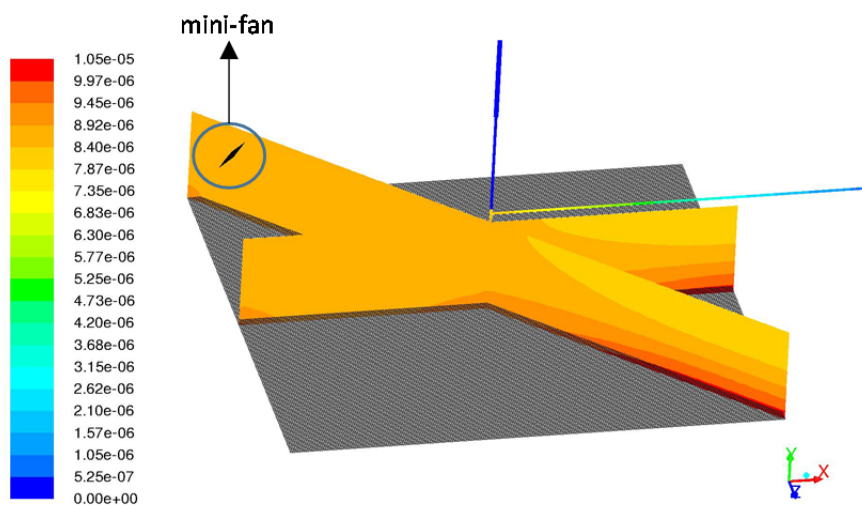


388
389 *Figure 11 CH₄ concentration profile in [kmol/m³] after the 10 min deposition period (left) and after the 2 min FID operation (right),*
390 *with mini-fan.*

391 Looking at Fig. 11 (left), it shows that the introduction of a mini-fan does result in the expected effect – i.e.
392 the achievement of a good mixing degree in the chamber also before the FID operation – in facts the
393 outcome is a fairly uniform concentration in the hood.

394 In this case, after the 2 min FID operation period, Fig. 11 (right), the situation is not significantly different
395 from the one depicted in Fig. 11 (left), i.e. before FID operation. This indicates that the mini-fan serves well
396 its purpose.

397 In order to better appreciate the action of the mini fan, Fig.12 shows a different plane inside the chamber.



398
399 *Figure 12 CH₄ concentration profile in [kmol/m³] after the 10 min deposition period, different angle.*

400 In analogy with paragraph 4.2.2, it is possible to observe the CH₄ concentrations (average and at sampling
401 point) both before and after FID operation, the values are shown in Tab.2.

402 **5. CONCLUSIONS**

403 The main goal of the present work was to provide a method for designing a static sampling hood, exploiting
404 CFD simulations. Such method was presented, applied to the specific case of methane concentration
405 measurement on landfill surfaces. However, even though the procedure depicted is specific, the method is
406 to be considered as general: in facts, by considering the peculiarities of the case considered, it can be
407 applied to the optimized design and suitable development of any sampling hood for the assessment of
408 emissions from area sources.

409 This study highlights the main aspects to focus on in order to successfully design an effective sampling
410 device. First, it is necessary to verify the fundamental hypothesis made that allow writing the mass balance
411 for the hood. Then, it is necessary to check whether the system is segregated or mixed, leading to a
412 discussion concerning the optimal position for the sampling point. Finally, it is recommendable to
413 investigate what are possible modifications to the design proposed in the beginning and what such
414 modifications imply in terms of fluid-dynamic behavior of the system inside the chamber.

415 The project proved that CFD could be effectively applied for designing specific static sampling devices for
416 area sources; moreover, it showed CFD to be a valuable asset for this task.

417 Referring to the analyzed situation of LFG emissions from landfill surfaces, the simulations performed
418 yielded several interesting outcomes.

419 First, they verified that CH₄ concentration grows linearly over time inside the hood, in the first 10-15
420 minutes, allowing to write the mass balance for the chamber according to Eq.3. Moreover, it was shown
421 that the FID operation provides a certain degree of mixing to the system, thus increasing the reliability of
422 the measured data, since the sampled gas is likely to have a CH₄ concentration very close to the average,
423 making the measurement representative of the general condition.

424 It was also proved that the sampling point should be always at a height above ground about 40% of the
425 hood's height, since this is where the concentration value is almost equal to the average concentration
426 value inside the hood. This statement holds true even as both the dimensions of the hood and the intensity
427 of the LFG emission vary.

428 The simulations also suggested a possible improvement for the design: introducing a mini-fan inside the
429 hood to achieve mixing also before turning on the FID. The modification turned out to be actually an
430 improvement: the introduction of a mini-fan affects the system, providing a satisfactory mixing degree also
431 before FID operation, as it is possible to see the situation before and after FID operation is fairly similar (see
432 Fig. 11).

433

434 **REFERENCES**

435 ANSYS Fluent User's Guide, Release 14.5, 2012.

436 Bogner, J., Spokas, K., Burton, E., Sweeney, R., Corona, V., 1995. Landfills as atmospheric methane sources
437 and inks. *Chemosphere* 9, 4119-4130. DOI 10.1016/0045-6535(95)00273-B.

438 Capelli, L., Sironi, S., Del Rosso, R., Céntola, P., Il Grande, M., 2008. A comparative and critical evaluation of
439 odour assessment methods on a landfill site. *Atmos. Environ.* 42, 7050-7058. DOI
440 10.1016/j.atmosenv.2008.06.009.

441 Capelli, L., Sironi, S., Del Rosso, R., 2013. Odor sampling: techniques and strategies for the estimation of
442 odor emission rates from different source types. *Sensors* 13, 938-955. DOI 10.3390/s130100938.

443 Capelli, L., Sironi, S., Del Rosso, R., Magnano, E., 2014. Evaluation of landfill surface emissions. *Chem. Eng.*
444 *Trans.* 40, 187-192. DOI 10.3303/CET1440032.

445 Capelli, L., Lucernoni, F., Sironi, S., 2015. Odour impact assessment by dispersion modelling: a comparison
446 of different approaches for the estimation of odour emissions from landfill surfaces. *Proceedings of*
447 *Sardinia Symposium 2015, Santa Margherita di Pula, 5-9 October 2015. Organized by the International*
448 *Waste Working Group (IWWG), c/o Hamburg University of Technology, Harburger Schlosstr. 36, 20179*
449 *Hamburg, Germany.*

450 Davoli, E., Gangai, M.L., Morselli, L., Tonelli, D., 2003. Characterisation of odorants emissions from landfills
451 by SPME and GC/MS. *Chemosphere* 51, 357-368. DOI 10.1016/S0045-6535(02)00845-7.

452 Hudson, N., Ayoko, G.A., 2008. Odour Sampling 1: Physical chemistry considerations. *Biores. Technol.* 99,
453 3982-3992. DOI 10.1016/j.biortech.2007.04.034.

454 Jiang, K., Kaye, R., 1996. Comparison study on portable wind tunnel system and isolation chamber for
455 determination of VOCs from areal sources. *Water Sci. Technol.* 34, 583-589.

456 Klenbusch, M.R., 1986. Measurement of gaseous emission rates from land surfaces using an emission
457 isolation flux chamber. Radian Corporation, US-EPA.

458 Lucernoni, F., Tapparo, F., Capelli, L., Busini, V., Del Rosso, R., Sironi, S., 2015. Evaluation of landfill gas
459 surface emissions. Proceedings of IWA Odours 2015, Paris, 15-18 November 2015. Organized by the
460 International Water Association (IWA), c/o Alliance House 12 Caxton Street, London SW1H 0QS, United
461 Kingdom.

462 Palmiotto, M., Fattore, E., Paiano, V., Celeste, G., Colombo, A., Davoli, E., 2014. Influence of a municipal
463 solid waste landfill in the surrounding environment: Toxicological risk and odor nuisance effects. Environ.
464 Int. 68, 16-24. DOI 10.1016/j.envint.2014.03.004.

465 Park, J.W., Shin, H.C., 2001. Surface emissions of landfill gas from solid waste landfill. Atmos. Environ. 35,
466 3445-3451. DOI 10.1016/S1352-2310(01)00118-2.

467 Prata Jr., A.A., Santos, J.M., Beghi, S.P., Fernandes, I.F., Vom Marttens, L.L.C., Neto, L.P., Martins, R.S., Reis
468 Jr., N.C., Stuetz, R.M., 2016. Dynamic flux chamber measurements of hydrogen sulfide emission rate from a
469 quiescent surface – A computational evaluation. Chemosphere 146, 426-434. DOI
470 10.1016/j.chemosphere.2015.11.123.

471 Rachor, I.M., Gebert, J., Grongroft, A. & E., Pfeiffer, M., 2013. Variability of methane emissions from an old
472 landfill over different time-scales. Eur. J. Soil Sci. 64, 16–26. DOI 10.1111/ejss.12004.

473 Reinhart, D.R., Cooper, D.C., Walker, B.L., 1992. Flux chamber design and operation for the measurement of
474 municipal solid waste landfill gas emission rates. Air Waste Manage. Assoc. 42, 1067-1070. DOI
475 10.1080/10473289.1992.10467053.

476 Sarkar, U., Hobbs, S.H., 2002. Odour from municipal waste (MSW) landfills: a study on the analysis of
477 perception. Environ. Int. 27, 655–662.

478 Sironi, S., Capelli, L., Centola, P., Del Rosso, R., Il Grande, M., 2005. Odour emission factors for assessment
479 and prediction of Italian MSW landfills odour impact. *Atmos. Environ.* 39, 5387-5394. DOI:
480 10.1016/j.atmosenv.2005.05.023.

481 United Kingdom Environment Agency, UK EA, 2010. *Guidance On Monitoring Landfill Gas Surface Emissions*,
482 UK EA.